Endangered Species Act Section 7(a)(2) Biological Opinion Biorka Island Dock Replacement Project Sitka, Alaska

NMFS Consultation Number: AKR-2017-9693

Action Agency: Federal Aviation Administration (FAA), and Permits and Conservation

Division, Office of Protected Resources (PR1), National Marine Fisheries

Service, NOAA

Affected Species and Determinations:

ESA-Listed Species	Status	Is Action Likely to Adversely Affect Species?	Is the Action Likely to Adversely Affect Critical Habitat?	Is Action Likely To Jeopardize the Species?	Is Action Likely To Destroy or Adversely Modify Critical Habitat?
Humpback Whale, Mexico DPS (Megaptera novaeangliae)	Threatened	Yes	N/A	No	N/A
Sperm whale (Physeter microcephalus)	Endangered	No	N/A	No	N/A
Steller Sea Lion, Western DPS (Eumatopias jubatus)	Endangered	Yes	No	No	No

Consultation Conducted By: National Marine Fisheries Service, Alaska Region

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TERMS AND ABBREVIATIONS

BOEM	Bureau of Ocean Energy Management
BSAI	Bering Sea/Aleutian Islands
BSEE	Bureau of Safety and Environmental
	Enforcement
CI	Confidence Interval
dB	decibel
DPS	distinct population segment
DTH	down-the-hole
EIS	environmental impact statement
ESA	Endangered Species Act
ESCA	Endangered Species Conservation Act
IHA	incidental harassment authorization
IWC	International Whaling Commission
kHz	kilohertz
kts	knots
MMPA	Marine Mammal Protection Act
NMFS	National Marine Fisheries Service
NOAA	National Oceanic and Atmospheric
	Administration
NPDES	National Pollutant Discharge Elimination
	System
Opinion	this biological opinion
р-р	peak-to-peak
PAM	passive acoustic monitoring
PBF	physical or biological features
PCE	primary constituent element
Permits Division	NMFS Office of Protected Resources, Permits
	and Conservation Division
PSO	protected species observer
PTS	permanent threshold shift
rms	root mean square
SSV	sound source verification
TTS	temporary threshold shift
USFWS	U.S. Fish and Wildlife Service
ZOI	zone of influence
μΡα	micropascal
0-p	peak

1 INTRODUCTION

Section 7(a)(2) of the Endangered Species Act of 1973, as amended (ESA; 16 U.S.C. § 1536(a)(2)), requires each Federal agency to ensure that any action it authorizes, funds, or carries out is not likely to jeopardize the continued existence of any endangered or threatened species or result in the destruction or adverse modification of critical habitat of such species. When a Federal agency's action "may affect" a protected species, that agency is required to consult with the National Marine Fisheries Service (NMFS) or the U.S. Fish and Wildlife Service (USFWS), depending upon the endangered species, threatened species, or designated critical habitat that may be affected by the action (50 CFR § 402.14(a)). Federal agencies may fulfill this general requirement informally if they conclude that an action "may affect, but is not likely to adversely affect" endangered species, threatened species, or designated critical habitat, and NMFS or the USFWS concurs with that conclusion (50 CFR § 402.14(b)).

Section 7(b)(3) of the ESA requires that at the conclusion of consultation, NMFS and/or USFWS provide an opinion stating how the Federal agency's action is likely to affect ESA-listed species and their critical habitat. If incidental take is reasonably certain to occur, section 7(b)(4) requires the consulting agency to provide an incidental take statement (ITS) that specifies the impact of any incidental taking, specifies those reasonable and prudent measures necessary to minimize such impact, and sets forth terms and conditions to implement those measures.

For the actions described in this document, the action agencies are the Federal Aviation Administration (FAA), which proposes to construct a replacement dock on Biorka Island in Symonds Bay near Sitka, Alaska; and NMFS Office of Protected Resources, Permits and Conservation Division (PR1), which proposes to issue an incidental harassment authorization (IHA) pursuant to section 101(a)(5)(D) of the Marine Mammal Protection Act of 1972, as amended (MMPA) (16 U.S.C. § 1361 *et seq.*), to the FAA for harassment of marine mammals incidental to construction operations (82 FR 41229). The consulting agency is NMFS's Alaska Regional Office.

This document represents NMFS's biological opinion (opinion) on the proposed action and its effects on endangered and threatened species and designated critical habitat for those species.

The opinion and ITS were prepared by NMFS in accordance with section 7(b) of the ESA, and implementing regulations at 50 CFR § 402.

The opinion and ITS are in compliance with the Data Quality Act (44 U.S.C. § 3504(d)(1) *et seq.*) and underwent pre-dissemination review.

1.1 Background

This opinion is based on information provided to us in the August 2017 IHA application and marine mammal monitoring and mitigation plan (ECO49 2017), initiation package (FAA 2017b), proposed IHA (82 FR 41229), and the September 2017 Environmental Assessment (FAA

2017a). Other sources of information relied upon included updated project proposals, emails and telephone conversations between NMFS Alaska Region, PR1, FAA staff and FAA's contractor (ECO49 Consulting, LLC. [ECO49]). A complete record of this consultation is on file at NMFS's field office in Juneau, Alaska.

The proposed action involves the improvement and maintenance of the dock and barge landing area located on Biorka Island, and the navigational and weather facilities located on the Island (Figure 1). The existing dock is deteriorated and has reached the end of its useful life. Regular and repetitive heavy surging seas, along with constant use have destroyed the face of the existing floating marine dock. The existing barge landing area is reinforced seasonally by adding fill to the landing at the shoreline, which is periodically washed away by storms and wave action. The Project would reconstruct the deteriorated existing dock and construct an improved barge landing area.



Figure 1. Biorka Island Dock Construction Location (ECO49 2017).

This opinion considers the effects of removal and construction operations, and the associated proposed issuance of an IHA. These actions may affect the following species: Mexico distinct population segment (DPS) humpback (*Megaptera novaeangliae*), sperm whale (*Physeter microcephalus*), and western DPS Steller sea lions (*Eumetopias jubatus*). No designated critical habitat is located within the action area. The nearest designated critical habitat, for Steller sea lions, is Kaiuchali Island located ~1.6 km southwest of Biorka Island.

1.2 Consultation History

Our communication with PR1, FAA and ECO49 regarding this consultation is summarized as follows:

- May 23, 2016: received informal section 7 consultation initiation request for the FAA (FAA 2016).
- May 25, 2016: NMFS submitted an additional information request.
- **June 7, 2016:** NMFS submitted additional information request and indicated initiation package was insufficient for initiation of consultation.
- **July 7, 2016:** FAA responded with additional information.
- **September 13, 2016:** NMFS's review of the proposed action indicated potential for take of ESA-listed species, and formal consultation was recommended.
- August 4, 2017: FAA submitted Draft IHA application.
- **August 15, 2017:** FAA submitted a request to initiate formal section 7 consultation (FAA 2017b).
- August 15, 2017: NMFS initiated consultation with FAA.
- August 17, 2017: sent a notice of Section 7 consultation request for information to the Alaska Department of Fish and Game
- August 29, 2017: NMFS's Permits Division submitted a request to initiate formal section 7 consultation in an initiation package that included ECO 49 IHA application (ECO49 2017), and a draft proposed Federal Register notice for Biorka Island Dock Replacement Project (82 FR 41229).
- August 29, 2017: initiated consultation with NMFS's Permits Division.
- September 25, 2017: FAA submitted Draft Environmental Assessment

2 DESCRIPTION OF THE PROPOSED ACTION AND ACTION AREA

2.1 Proposed Action

"Action" means all activities or programs of any kind authorized, funded, or carried out, in whole or in part, by Federal agencies. "Interrelated actions" are those that are part of a larger action and depend on the larger action for their justification. "Interdependent actions" are those that have no independent utility apart from the action under consideration (50 CFR 402.02).

This opinion considers the effects of FAA's removal of the existing dock structures, and constructing a new dock and improved barge landing facility, as well as issuing an IHA to take marine mammals by harassment under the MMPA incidental to these actions in Symonds Bay near Sitka, AK between May 1, 2018 and April 30, 2019.

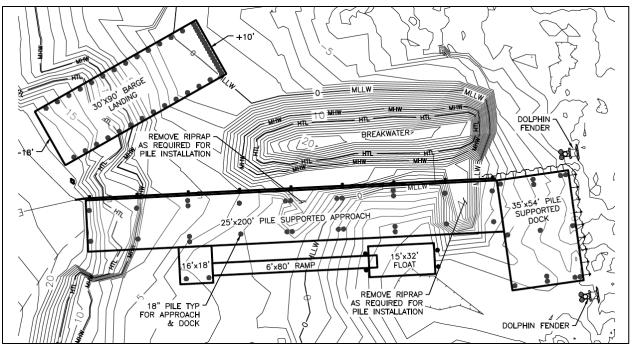


Figure 2. Main structures for the proposed Biorka Island Dock Replacement Project (ECO49 2017).

2.1.1 Demolition of Existing Dock and Pile Removal

The existing dock is a T-shaped, pile-supported structure consisting of a 170-ft long by 16-ft wide approach trestle with a 51-ft wide by 35-ft long end section. The existing infrastructure also includes a 30-ft by 32-ft floating dock that is accessed by a 5-ft wide by 50-ft long steel gangway, a small 10-ft by10-ft pre-fabricated building, and an electric hydraulic pedestal crane.

A total of 46 existing piles would be removed (Table 1). The steel and timber piles would be pulled out of the substrate directly with a crane and sling, by using a vibratory hammer, or with a clamshell bucket. The three concrete piles that are located above the high tide were cast in place. The concrete piles are set in bedrock and will be removed at low tide using standard excavation equipment. Therefore, removal of these piles will not produce underwater noise. The construction contractor would determine the exact method for concrete pile removal.

Table 1. Existing piles to be removed (ECO49 2017).

Pile Type	Quantity	Size (in.)
Concrete	3	24
	14	8
Steel	8	10
	14	12.75
Timber	7	14 tapering to 8
Total	46	

The existing deck and other associated infrastructure would also be disassembled and removed. The existing 4-ton pedestal crane would be salvaged for relocation on the new dock. As necessary, portions of the existing rubble mound/breakwater would be removed to provide enough clearance for construction and then replaced once the dock has been constructed.

2.1.2 Construction of New Dock and Barge Landing

As part of constructing a new dock facility, the following structures will be installed: (1) a new trestle approach up to 25 ft wide; (2) an 80-ft aluminum gangway connecting to a 15-ft wide by 32-ft long small craft berthing float; (3) a dock face approximately 54-ft long and 35-ft wide; (4) two berthing dolphin fenders would be installed, one at each end section of the new dock with one 30-in. diameter plumb pile and two 18-in. diameter batter piles; and (5) a wave barrier, consisting of Z-sheet piles in between steel H piles, installed at the face of the dock (see Figure 2). Steel pipe pilings would support a precast concrete deck and trestle. Some piles would require internal tension anchors for increased support. Pile counts, sizes, and other details are shown in Table 2.

Table 2. Temporary Template and Permanent Pile Details for New Construction (ECO49 2017).

Component	Stage	Туре	Quantity	Size		
Pock ^{1,2} Vave Barrier Polphin Fenders ⁴	Template ³	Steel H or Pipe	60	12 in.		
DOCK *	Permanent	Steel Pipe	43	18 in.		
	Permanent	Sheet	32	NZ26		
Wave Barrier	Permanent	Steel H	16	W40x199		
	Template ³	Steel H or Pipe	4	12 in.		
Dolphin Fenders ⁴	Permanent	Steel Pipe	4	18 in.		
	Permanent	Steel Pipe	60 12 in. 43 18 in. 32 NZ26 16 W40x199 4 12 in.			
	Template ³	nt Sheet nt Steel H e³ Steel H or Pipe nt Steel Pipe nt Steel Pipe e³ Steel H or Pipe nt Steel Pipe nt Steel Pipe nt Steel H or Pipe nt Steel Pipe nt Steel Pipe	20	12 in.		
Barge Landing	Permanent	Steel Pipe	35	18 in.		
	Permanent	Sheet	34	NZ26		
T. 4.1	Template ³		84			
Total	Permanent		166			

¹Includes piles for the approach, end section, platform, and floating dock.

²Number of piles for dock is based on 25-ft approach trestle width.

³Noise from installation and removal of the template piles is considered in the analysis; therefore, template pile count equates to 2 x 84 or 168 but the actual number of piles to be installed is 84. Template piles were assumed to be 12-in. diameter for modeling.

⁴For two dolphin fender systems.

The current barge landing is located northwest of the existing dock and is comprised of gravel and cobbles with no formal structure. The uplands area on the west end of the trestle would be slightly graded into the existing terrestrial approach. The existing barge landing would be upgraded to a 30-ft by 90-ft precast concrete plank landing placed over fill, with a perimeter constructed of concrete, sheet piles, and 18-in. steel piles (see Table 2). Similar to the wave barrier, the sequence for installing the permanent barge ramp pipe piles would begin with advancement through overlying sediment with a vibratory hammer, followed by use of an impact hammer to drive the piles into bedrock.

Temporary Piles

Construction of the new dock would begin with the erection of a temporary template. The construction contractor would determine the specific type and size of template piles based on site conditions and availability of materials. For purposes of analysis, FAA has assumed template piles would be 12-inch diameter steel H or pipe piles (see Table 2). The template piles would be driven into the overburden by vibratory hammer and removed after the permanent piles are installed. A total of 84 template piles will be installed and then removed.

Permanent Piles

All permanent pipe piles would be installed using a combination of vibratory and impact hammering methods to drive the pile into the overburden. Pipe piles would then be drilled and socketed into the underlying bedrock using down-the-hole (DTH) hammering/drilling techniques. DTH equipment breaks up the rock below the pile while simultaneously installing the pile through rock formation. The pile is then set/confirmed with a few strikes of an impact hammer. Sheet piles would be driven into the overburden and set into the top layer of bedrock using a combination of vibratory and impact hammering.

Tension Anchors

Certain piles would require internal tension anchors. Up to eight of the dock piles and all six piles for the dolphins would require these internal tension anchors. Each pile with a tension anchor would first be drilled, socketed into bedrock, and proof driven with an impact hammer as described above for permanent piles. Then a separate smaller drill would be used to complete an approximately 5-in. diameter hole extending about 30 to 40 ft into bedrock below the tip of the pile. A steel bar would be grouted into this hole. Once the grout sets, a jack would be applied to the top of the bar and the tensioned rod would be locked off to plates at the top of the pile.

Permanent Wave Barrier

The wave barrier consisting of steel H piles with Z sheets in between is located at the face of the dock. The H piles and Z sheets would be initially driven through overlying sediment with a vibratory hammer, and set into the bedrock with an impact hammer. The wave barrier sheet piling would be driven either singly or in preassembled pairs.

2.1.3 Transport of Material and Equipment

FAA will employ the following number and types of vessels for construction operations (FAA 2016):

- (1) materials barge;
- (1) crane barge;

- (2) small push boats; and
- (1) small crew boat

Materials and equipment would be transported to the project site by barge. While work is conducted in the water, anchored barges will be used to stage construction materials and equipment. Vessels are anticipated to transit from Sitka, but may come from other communities following standard transit routes.

2.1.4 Dates and Duration of Activities

Construction would occur from approximately May through September 2018. However, the proposed authorization would be effective May 1, 2018 through April 30, 2019.

Based on the sequence of pile driving activities and the duration per method required to install piles, six construction scenarios were identified as representing the largest acoustic footprints of construction activities at the Biorka Island Project site over any 24-hour period (Table 5). These construction scenarios are described in detail in the underwater noise modeling report attached to the IHA application (Quijano and Austin 2017).

Table 3. Pile Driving and Removal Modeling Scenarios for the Biorka Island Project (ECO49 2017, Quijano and Austin 2017).

			Vib	ratory	DTH		Impact		
Scenario	Description	Piles installed*	Hrs per pile	Total hrs*	Hours per pile	Total hours*	Hours per pile	Total strikes*	Shift (hr)
81	Removal of existing piles and installation/removal of temporary piles	21		6.93	N	A	N	JA.	6.93
S2	Installation of 18-inch pipe piles (dock and dolphin)	3	0.22	0.99	2	6	0.17	15	7.49
S3	Installation of 18-inch pipe piles (barge landing)	4	0.33	1.32	NA		0.33	2,720	2.65
S4	Installation of 30-inch pipe piles (dolphins)	2		0.66	2	4	0.17	10	4.99
S5	Installation of H piles (dock wave barrier)	8		2.64	N	A	0.33	5,440	5.31
S6	Installation of sheet piles (dock wave barrier and barge landing)	12		3.96	N	A	0.25	6,120	6.96

Source: Appendix A (Quijano and Austin 2017)

NA indicates when a pile driving method was not required in a given scenario.

^{*}Refers to the number of piles, hours of operation of hammer strikes within a 24-hr period.

Pile installation would occur over a period of days. The estimated number of days required to complete each scenario is provided in Table 6. The total number of weighted days is not additive and does not represent the total duration of pile driving. There may be more than one pile driving method used over the course of a single construction day. However, the applicant does not anticipate more than two scenarios occurring over the course of a single day.

Table 4. Estimated Days to Complete Each Pile Driving Scenario (ECO49 2017).

		Weighted Totals (Number of Days Each Pile Driving Activity May Occur)							
Scenario	Number of Piles	Vibratory (Non-impulsive)	DTH (Non- impulsive)	Impact (Impulsive)					
S1	21	22	0	0					
S2	3	16	16	16					
S3	4	9	0	9					
S4	2	2	2	2					
S5	8	5	0	5					
S6 12		16	0	16					
Tot	al Days	70	18	48					

Note: The total number of weighted days are not additive because more than one activity may occur in a day. The days that each noise producing activity may occur are used in the Exposure Analysis see Section 6.2.

2.1.5 Acoustic Sources

There are a number of acoustic sources associated with removal and replacement of dock and barge landing structures including: vibratory pile driving, impact pile driving, and DTH hammering. Each of these elements generates in-water and in-air noise.

Three different pieces of pile driving equipment have been proposed for the construction of the dock at Biorka Island: the diesel impact hammer ICE I-36V2 for impact operations, the ICE-44B vibratory driver for vibratory, and the Numa Patriot 180 for DTH hydro-hammering. Table 5 lists relevant parameters for the acoustic modeling in this report. For impact pile driving, the rated energy was used to adjust the source levels obtained from measurements noted in the literature (for H and sheet piles), and as input to model pile driving source levels in the case of cylindrical piles. The rated striking frequency was used to estimate the total number of strikes in Scenarios S3 and S5 (see Table 3). The centrifugal force for the vibratory hammer was used to adjust the source levels obtained from measurements. Finally, although DTH hydro-hammering has impulsive source components, the high frequency of 1,100 blows/minute combined with long continuous operation intervals of several minutes make its signature noise more like a non-impulsive source and therefore we treat it as such in this opinion (Quijano and Austin 2017).

Table 5. Properties of pile driving equipment used for modeling the acoustic footprint from construction activities at Biorka Island (Quijano and Austin 2017).

Driving mechanism	Pile driver	Properties
Impact	ICE I-36V2	127.1 kNm rated energy, 34 strikes/minute
Vibratory	ICE-44B	1,789 kN centrifugal force
DTH hydro-hammering	Numa Patriot 180	1,100 blows/minute

Impact Hammer

Impact hammering is anticipated for scenarios 2-6 (see Table 4). An impact hammer is a steel device that works like a piston. The pile is first moved into position and set in the proper location using a choker cable or vibratory hammer. The impact hammer is held in place by a guide (lead) that aligns the hammer with the pile. A heavy piston moves up and down, striking the top of the pile and driving it into the substrate. Once the pile is set in place, pile installation with an impact hammer can take less than 15 minutes under good substrate conditions. However, under poor conditions, such as glacial till and bedrock or exceptionally loose material, piles can take longer to set. The proposed action anticipates using an ICE I-36V2 impact driver (ECO49 2017).

Sheet Piles

Spectral acoustic measurements from impact driving of sheet piles 24-in wide by 50 feet long (Buehler et al. 2015) were used to estimate source levels for the present study. Measurements in the frequency band 10–5,000 Hz were obtained at 10 m range in water depth of 13 m, when an ICE60S impact hammer (rated energy of 81.4 kJ) was in its final stage of driving these sheet piles.

The measured data were digitalized and processed in 1/3-octave-bands, yielding a broadband per-strike SEL reaching 180 dB re 1 μ Pa²·s. This can be considered high when compared to the typical broadband SEL for cylindrical pipes of similar dimensions (28-in diameter/ 50 feet long). Simulations were performed to correct suitably so that the measured broadband SEL would reflect a typical pile driving operation (versus the extreme). The simulated study used a cylindrical pile of 32.9 in (83.5 cm) diameter as a surrogate to determine how soil resistance impacts broadband SEL, SPL (rms), and peak pressure levels. This diameter was selected to approximate the width of an AZ25 sheet pile.

One-third-octave-band source levels for impact sheet pile driving have also been corrected to account for the higher energy of the ICE I-36V2 relative to the ICE60S; 20 dB were also added to account for the spherical spreading loss from the pile to the 10 m range where measurements were taken. No experimental data were available for frequencies higher than 5 kHz but the levels are expected to decrease with frequency, so the measurements were extrapolated to 25 kHz according to a decay rate of 14.1 dB/decade as suggested by the trend observed in the 1/3-octave-band levels from the measured data between 2000 and 5000 Hz. The anticipated source level broadband SEL for impact hammering sheet piles is 196 dB re 1 uPa²'s (Quijano and Austin 2017).

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¹ All installation methods for permanent piles propose the use of impact hammering. Only S1 pile removal and temporary pile installation does not propose to use impact hammering (ECO49 2017).

H-Piles

For impact pile driving of H piles, spectral measurements in the frequency band 10–4000 Hz for 12×12 in H piles were available from the literature (Buehler et al. 2015). The pile was driven using an APE D-19-42 diesel impact hammer (63.63 kN·m energy). These measurements were corrected for spherical spreading by +20.0 dB (corresponding to 10 m range of the measurement) and by +3 dB to account for the higher energy of the ICE I-36V2 hammer. One-third-octave-band SEL were obtained from the measured spectra up to 4000 Hz. The measurements were extrapolated to \sim 25 kHz according to a decay rate of 2.6 dB/decade as suggested by the trend observed in the 1/3-octave-band levels from the measured data between 1000-4000 Hz, resulting in the 1/3-octave-band SEL. The anticipated source level broadband SEL for impact hammering H piles is 189 dB re 1 uPa 2 's (Quijano and Austin 2017).

Cylindrical Piles

To model sounds resulting from impact pile driving of cylindrical pipes, JASCO's Pile Driving Source Model (PDSM) was used in conjunction with the GRLWEAP 2010 wave equation model. For consistency with the modeling approach used for all other piles (in which SEL 1/3-octave-band levels are input to MONM to determine received levels at given ranges and depths), the pile driving source signature obtained from PDSM was further converted into an equivalent single monopole source at the middle of the water column, suitable for MONM.

Impact pile driving in Scenarios 2 and 4 (see Table 3) is used to confirm that the pile is firmly positioned on the bedrock (i.e., pile near refusal). Pile refusal results in elevated SEL compared to normal pile driving conditions. For this reason, impact pile driving source levels for Scenarios 2 and 4 were obtained by specifying a high soil resistance input parameter in the PDSM model, which mimics the condition of low penetration rates during refusal. Figure 10 shows the two levels adjusted to account for the use of the impact hammer at pile refusal in Scenarios 2 and 4. Source level broadband SELs for 30-inch, and 18-inch at pile refusal for impact hammering are 202 dB and 194 dB re 1 uPa² 's respectively (Quijano and Austin 2017).

Vibratory Hammer

Vibratory hammering will be used during all of the scenarios analyzed (see Table 3). Thus, source level estimates were provided for installing sheet piles, H-piles, and cylindrical piles with a vibratory hammer. Generally, the pile is placed into position using a choker and crane, and then vibrated between 1,200 and 2,400 vibrations per minute. The vibrations liquefy the sediment surrounding the pile allowing it to penetrate to the required seating depth, or to be removed.

The pile driving equipment anticipated for vibratory hammering is the ICE-44B vibratory driver (ECO49 2017).

Sheet Piles

For vibratory pile driving of 24-in wide sheet piles, we used spectral measurements in the frequency band 10–4000 Hz from the literature (Buehler et al. 2015). In this case, the pile was driven using an APE 400 King Kong hydraulic vibratory hammer (3200 kN centrifugal force). These measurements were corrected for spherical spreading by +20 dB (corresponding to 10 m range) and by -2.5 dB (the logarithm of the ratio of the centrifugal force of the modeled to the measured hammer) to account for the smaller driving force of the ICE44B hammer. One-third-

octave-band SEL were obtained from the measured spectra up to 2500 Hz. At higher frequencies no experimental data were available, so the measurements were extrapolated to ~25 kHz according to a decay rate of 19 dB/decade as suggested by the trend observed in the 1/3-octave-band levels from the measured data between 1000-4000 Hz. The anticipated source level broadband SEL for vibratory hammering sheet piles is 180 dB re 1 uPa² 's (Quijano and Austin 2017).

H-Piles

For vibratory pile driving of H piles, spectral measurements in the frequency band 10–10,000 Hz for 14×14 in H piles were available from the literature (Burgess et al. 2005). In this case, the pile was driven using an APE 200 hydraulic vibratory hammer (1797 kN centrifugal force). These measurements were corrected for spherical spreading by +22.9 dB (corresponding to 14 m range) and by -0.02 dB to account for the slightly smaller driving force of the ICE44B hammer. One-third-octave-band SEL were obtained from the measured spectra up to 10,000 Hz. The measurements were extrapolated to ~ 25 kHz according to a decay rate of 20.2 dB/decade as suggested by the trend observed in the 1/3-octave-band levels from the measured data between 4000-10,000 Hz. The anticipated source level broadband SEL for vibratory hammering H piles is 178 dB re 1 uPa 2 's (Quijano and Austin 2017).

Cylindrical Piles

1/3-octave-band levels from measurements of vibratory pile driving of 30-in diameter piles was obtained at Ketchikan, AK (Warner and Austin 2016a), were adjusted to match the broadband SEL obtained by the preliminary model using GRLWEAP and PDSM. The resulting 1/3-octave-band levels exhibit peaks over multiple bands. Source level broadband SELs for 30-inch, 18-inch, and 12-inch piles for vibratory hammering are 193 dB, 187 dB, and 180dB re 1 uPa²'s respectively (Quijano and Austin 2017).

DTH Hydro-Hammering

In this project, the DTH hydro-hammer operates in vertical piles that have been partially driven by vibratory means. Before it begins operating, the DTH hydro-hammer is installed within the hollow pipe pile at the bottom of the pile. Piles are advanced by applying a pulsating mechanism to break the underlying bedrock while simultaneously removing broken rock fragments. We have assumed that the interaction between the rock and the DTH hydro-hammer is what generates noise, therefore sound levels do not depend on pile diameter (ECO49 2017). Only scenarios 2 and 4 propose to use hydro-hammering (see Table 3).

In May 2016, a Numa Patriot 180 hammer was used to drive 24 in diameter piles at a ferry terminal at Kodiak, AK (Warner and Austin 2016b). Acoustic signatures for DTH hydro-hammering were recorded at ranges of 10–30 m from the pile. The measured source levels at each 1/3-octave-band from these measurements were adjusted by $20\log_{10}(\text{range})$ (i.e., back propagated assuming spherical spreading) and averaged to provide the representative 1/3-octave-band SEL, which were used for acoustic modeling in Scenarios S2 and S4. The anticipated source level broadband SEL for DTH is 192 dB re 1 uPa² 's (Quijano and Austin 2017).

2.1.6 Mitigation Measures

FAA Proposed Mitigation Measures

The following measures will be incorporated by FAA to minimize potential impacts from project activities:

- 1) The dock would be maintained in a manner that does not introduce any pollutants or debris into the harbor or cause a migration barrier for fish;
- 2) Fuels, lubricants, and other hazardous substances would not be stored below the ordinary high water mark;
- 3) Properly sized equipment would be used to drive piles;
- 4) Oil booms would be readily available for containment should any releases occur;
- 5) The contractor would check for leaks regularly on any equipment, hoses, and fuel storage that occur at the project site;
- 6) All chemicals and petroleum products would be properly stored to prevent spills; and
- 7) No petroleum products, cement, chemicals, or other deleterious materials would be allowed to enter surface waters.
- 8) All vessels associated with project construction will avoid the 3,000 ft (914 m) designated aquatic zones surrounding any major rookery or haulout east of 144° W longitude.

NMFS PR1 Proposed Mitigation

The Permits Division (PR1) proposes to issue an IHA for non-lethal "takes" of marine mammals by Level A and Level B harassment (as defined by the MMPA) incidental to FAA's proposed action (82 FR 41229). When issued, the IHA will be valid from May 1, 2018, to April 30, 2019, and will authorize the incidental harassment of one ESA-listed whale species, and one listed sea lion species, as well as five non-ESA-listed whale and pinniped species. Table 6 shows the amount of proposed take for the two ESA-listed species in the proposed IHA.³

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² The MMPA defines "harassment" as "any act of pursuit, torment, or annoyance which has the potential to injure a marine mammal or marine mammal stock in the wild" (referred to as Level A harassment) or "has the potential to disturb a marine mammal or marine mammal stock in the wild by causing disruption of behavioral patterns, including but not limited to, migration, breathing, nursing, breeding, feeding, or sheltering" (referred to as Level B harassment). 16 U.S.C. 1362(18)(A) and (B). For the purposes of this consultation, NMFS considers that a take by

[&]quot;harassment" occurs when an animal is exposed to certain sound levels described below in Section 6 of this opinion.

³ Please see proposed IHA (<u>82 FR 41229</u>) for MMPA-authorized takes of marine mammal species not listed under the ESA.

Table 6. Amount of proposed incidental harassment (takes) of ESA-listed species in the proposed IHA (82 FR 41229).

Species	Proposed Authorized Level A Takes	Proposed Authorized Level B Takes		
Western DPS Steller sea lion (Eumatopias jubatus)	3	175 ⁴		
Mexico DPS Humpback whale (Megaptera noviaengliae)	0	22 ⁵		

The mitigation measures described below are required per the NMFS's IHA stipulations, and will be implemented by FAA to reduce potential impacts to marine mammals from pile removal and installation activities, DTH hydro-hammering operations, and vessel movements. Unless otherwise noted, these measures apply to all marine mammal species.

Establishing Exclusion and Disturbance Zones

Exclusion Zone (i.e., shutdown zone) – For all pile driving/removal and DTH hammering activities, the FAA will establish an exclusion zone intended to contain the area in which SPLs equal or exceed the auditory injury criteria for cetaceans and pinnipeds. The purpose of an exclusion zone is to define an area within which shutdown of activity would occur upon sighting of a marine mammal (or in anticipation of an animal entering the defined area), thus preventing injury (Level A harassment) of marine mammals (see *Response Analysis* Section 6.3, although serious injury or death are unlikely outcomes even in the absence of mitigation measures). Modeled radial distances for exclusion zones are shown in Table 7. However, a minimum shutdown zone of 10 m will be established during all pile driving activities, regardless of the estimated zone.

Disturbance Zone – Disturbance zones are the areas in which SPLs equal or exceed 160 and 120 dB rms (Level B harassment for impulse and continuous sound, respectively). Disturbance zones provide utility for monitoring conducted for mitigation purposes (*i.e.*, exclusion zone monitoring) by establishing monitoring protocols for areas adjacent to the exclusion zones. Monitoring of disturbance zones enables observers to be aware of and communicate the presence of marine mammals in the Project Area but outside the exclusion zone and thus prepare for potential shutdowns of activity. However, the primary purpose of disturbance zone monitoring is for documenting instances of Level B harassment; disturbance zone monitoring is discussed in greater detail later (see *Proposed Monitoring and Reporting*). Nominal radial distances for disturbance zones are shown in Table 7.

⁴ The proposed IHA (<u>82 FR 41229</u>) indicated a requested Level A take of 6 Steller sea lions, and a Level B take of 350 Steller sea lions. Of the proposed takes, 50% are anticipated to occur to ESA-listed western DPS animals. The basis for this apportionment is described below in Section 4.3.2

⁵ The proposed IHA (<u>82 FR 41229</u>) indicated a requested Level A take of 3 humpback whales, and a Level B take of 350 humpback whales. Humpback whales in southeast Alaska include individuals from two DPSs. Of the 3 level A exposures, we anticipate zero threatened Mexico DPS animals will be taken, and all 3 will occur from the Hawaii DPS, which is not listed under the ESA. The basis for this apportionment is described below in Section 4.3.1.

Given the size of the disturbance zone for vibratory pile driving and DTH drilling (~10 km; see Figure 4), it is impossible to guarantee that all animals would be observed or to make comprehensive observations of fine-scale behavioral reactions to sound, and only a portion of the zone (*e.g.*, what may be reasonably observed by visual observers stationed between Symonds Bay and Sitka Sound) would be observed. In order to document observed instances of harassment, monitors record all marine mammal observations, regardless of location. The observer's location, as well as the location of the pile being driven, is known from a GPS. The location of the animal is estimated as a distance from the observer, which is then compared to the location from the pile. It may then be estimated whether the animal was exposed to sound levels constituting incidental harassment on the basis of predicted distances to relevant thresholds in post-processing of observational and acoustic data, and a precise accounting of observed incidences of harassment created. This information may then be used to extrapolate observed takes to reach an approximate understanding of total takes.

Table 7. Distances to Level A Shutdown and Level B Exposure Zones (ECO49 2017).

		Distance to Level A Shutdown zone (m) ¹										
	Scen	ario 1 ²	Scenario 2 ³		Scenario 3		Scenario 4 ³		Scenario 5		Scenario 6	
	22	days	16 da	ays	9 days		2 days		5 days		16 days	
Species	Continuous	Impulse	Continuous	Impulse	Continuous	Impulse	Continuous	Impulse	Continuous	Impulse	Continuous	Impulse
Harbor Porpoise	10	10	600	50	10	800	400	100	10	1,6004	80	1,6004
Humpback whales	10	10	400	30	10	700	300	80	10	700	40	1,400
Harbor Seals	10	10	80	10	10	200	60	10	10	300	10	800
Killer whales	10	10	10	10	10	10	10	10	10	150	10	250
Steller sea lions	10	10	10	10	10	10	10	10	10	10	10	80
	Distance to Level B Exposure Zones (m) ¹											
All marine mammals	1,800	10	10,1005	800	5,000	600	10,1005	1,300	800	430	5,000	800

NOTE: Vibratory and impulse hammering will not happen simultaneously; there will be sufficient time tor PSOs to be notified and to adjust monitoring as needed. A PSO will be stationed at the mouth of the Symonds Bay about 800 m from the noise source. Pink shading indicates monitoring and potential shutdown for presence outside of the bay is required. Green indicates zones that apply to ESA-listed species.

¹From Table 6-3 in IHA application rounded up as appropriate (see Section 6.1.2 Table 10).

²Scenario 1 does not include impulse hammering

³Includes DTH drilling (continuous).

⁴Actual Level A zone is larger (see Table 10), but 1.6 km (1 mile) is considered to be a reasonable distance to monitor.

⁵Takes will be extrapolated due to these large monitoring zones.

Monitoring Protocols – Monitoring would be conducted before, during, and after pile driving and vibratory removal activities. In addition, observers shall record all instances of marine mammal occurrence, regardless of distance from activity, and shall document any behavioral reactions in concert with distance from piles being driven or removed. Observations made outside the exclusion zone will not result in shutdown; that pile segment would be completed without cessation, unless the animal approaches or enters the exclusion zone, at which point all pile driving or removal activities would be halted. Monitoring will take place from 30 minutes prior to initiation through 30 minutes post-completion of pile driving and removal activities. Pile driving activities include the time to install or remove a single pile or series of piles, as long as the time elapsed between uses of the pile driving equipment is no more than 30 minutes. Please see Section 11 of the FAA's application

(<u>www.nmfs.noaa.gov/pr/permits/incidental/construction.htm</u>), for the FAA's proposed monitoring protocols.

Protected Species Observers (PSOs)

- 1. Monitoring will be conducted by qualified observers, who will be placed at the best vantage point(s) practicable to monitor for marine mammals and implement shutdown/delay procedures when applicable by calling for the shutdown to the hammer operator. A minimum of two observers will be required for all pile driving/removal activities. PSO requirements for construction actions are as follows:
 - a. Independent observers (i.e., not construction personnel) are required;
 - b. For all pile driving and removal activities, a minimum of two observers shall be deployed, with one positioned to achieve optimal monitoring of the exclusion zone, and the second positioned to achieve optimal monitoring of surrounding waters of Biorka dock and portions of Symonds Bay and Sitka Sound. If practicable, the second observer should be deployed to an elevated position with clear sight lines to the Project area.
 - c. At least one observer must have prior experience working as an observer;
 - d. Other observers (that do not have prior experience) may substitute education (undergraduate degree in biological science or related field) or training for experience;
 - e. Where a team of three or more observers are required, one observer should be designated as lead observer or monitoring coordinator. The lead observer must have prior experience working as an observer; and
 - f. NMFS will require submission and approval of observer resumes.
 - g. Maximum of 4 consecutive hours on watch per PSO; and
 - h. Maximum of ~12 hours of watch time per day per PSO
- 2. Qualified PSOs are trained biologists, and need the following additional minimum qualifications:
 - Visual acuity in both eyes (correction is permissible) sufficient for discernment
 of moving targets at the water's surface with ability to estimate target size and
 distance;
 - b. Ability to conduct field observations and collect data according to assigned protocols;
 - c. Experience or training in the field identification of marine mammals, including the identification of behaviors:

- d. PSO training must be provided prior to project start and in accordance with the monitoring measures in the application and biological opinion, and shall include at a minimum:
 - i. Species identification (sufficient to distinguish the species listed in Table 10 of the proposed IHA)
- e. Sufficient training, orientation, or experience with the construction operations to provide for personal safety during observations;
- f. Writing skills sufficient to prepare a report of observations including but not limited to the number and species of marine mammals observed; dates and times when in-water construction activities were conducted; dates and times when in-water construction activities were suspended to avoid potential incidental injury from construction sound of marine mammals observed within a defined exclusion zone; and marine mammal behavior; and
- g. Ability to communicate orally, by radio or in person, with each other, and relevant project personnel (e.g., those necessary to effect activity delay or shutdown) to provide real-time information on marine mammals observed in the area as necessary.
- 3. PSOs will be located at the best vantage point(s) in order to properly see the entire exclusion zone and as much of the disturbance zone as possible;
- 4. The exclusion and disturbance zones around the pile will be monitored for the presence of marine mammals before, during, and after any pile driving or removal activity;
- 5. During all observation periods, observers will use binoculars and the naked eye to search continuously for marine mammals;
- 6. Prior to the start of pile driving activity, the exclusion zone will be monitored for 30 minutes to ensure that it is clear of marine mammals. Pile driving will only commence once observers have declared the exclusion zone clear of marine mammals; animals will be allowed to remain in the exclusion zone (*i.e.*, must leave of their own volition) and their behavior will be monitored and documented. The exclusion zone may only be declared clear, and pile driving started, when the entire exclusion zone is visible (*i.e.*, when not obscured by dark, rain, fog, *etc.*). Should such conditions arise while impact, vibratory, or DTH driving is underway, the activity would be halted; and
- 7. If a marine mammal approaches or enters the exclusion zone during the course of pile driving and removal operations, activity will be halted and delayed until either the animal has voluntarily left and been visually confirmed beyond the exclusion zone or 15 minutes have passed without re-detection of small cetaceans and pinnipeds, and 30 minutes for humpback whales. Monitoring will be conducted throughout the time required to drive or remove a pile.
- 8. Using delay and shut-down procedures, if a species for which authorization has not been granted or if a species for which authorization has been granted but the authorized takes are met, approaches or is observed within the Level B harassment zone, activities will shut down immediately and not restart until the animals have been confirmed to have left the area.

Soft Start Procedures

The use of a soft start procedure for impact pile driving is believed to provide additional protection to marine mammals by warning or providing a chance to leave the area prior to the hammer operating at full capacity, and typically involves a requirement to initiate sound from the hammer at reduced energy followed by a waiting period. This procedure is repeated two additional times. It is difficult to specify the reduction in energy for any given hammer because of variation across drivers and, for impact hammers, the actual number of strikes at reduced energy will vary because operating the hammer at less than full power results in "bouncing" of the hammer as it strikes the pile, resulting in multiple "strikes." For impact driving, we require an initial set of three strikes from the impact hammer at reduced energy, followed by a 30-second waiting period, then 2 subsequent 3 strike sets. Soft start will be required at the beginning of each day's impact pile driving work and at any time following a cessation of impact pile driving of 30 minutes or longer. Soft start procedures will not be required for vibratory hammering operations.

In-Water Work

For in-water heavy machinery work other than pile driving (*e.g.*, standard barges, tug boats, barge-mounted excavators, or clamshell equipment used to place or remove material), if a marine mammal comes within 10 meters, operations shall cease and vessels shall reduce speed to the minimum level required to maintain steerage and safe working conditions.

Timing Restrictions

The FAA will only conduct construction activities during daytime hours. Construction will also be restricted to the months of May through September to avoid overlap with times when marine mammals have higher densities in the project area.

Sound Source Verification

FAA will conduct in-situ sound source verification tests to verify the radii of the exclusion and harassment zones for two 18-inch piles and two 30-inch piles for vibratory, impact, and DTH hammering methods. These data will establish more accurate radii for exclusion and harassment zones. FAA will submit the results to NMFS within five days after completing measurements for each of the pile diameters and hammering methods tested, followed by a report submitted within 14 days after the completion of the measures.

The SSV is based on dual metric criteria that will establish source levels for impact pile driving, vibratory pile driving, and DTH drilling. The FAA will provide all monitoring data to NMFS. The reports would include the following information:

- 1. Size and type of piles;
- 2. A detailed description of the noise attenuation device, including design specifications;
- 3. The impact hammer energy rating used to drive the piles, and the make and model of the hammer and the output energy;
- 4. The physical characteristics of the bottom substrate into which the piles were driven;
- 5. The depth of water in which the pile was driven;
- 6. The depth into the substrate that the pile was driven;

- 7. A description of the sound monitoring equipment;
- 8. The distance between hydrophones and pile;
- 9. The depth of the hydrophones and depth of water at hydrophone locations;
- 10. The distance from the pile to the water's edge;
- 11. The total number of strikes to drive each pile and for all piles driven during a 24-hour period;
- 12. The results of the hydroacoustic monitoring;
- 13. Source levels for peak and RMS SPLs and single strike SEL at 10 m from the pile, and RMS pulse duration that contains 90% of pulse energy.
- 14. The distance at which peak, cumulative SEL, and RMS values exceed the respective threshold values:
- 15. For vibratory pile driving, SEL based on 30 second averaging of sound intensity;
- 16. The spectra graphs for each pile type; and
- 17. A description of any observable marine mammal behavior in the immediate area and, if possible, correlation to underwater sound levels occurring at that time.

A minimum of two piles of the 18-in and two piles of the 30-in piles for each construction type (i.e. impact and vibratory pile driving and DTH drilling) will be monitored. Piles chosen to be monitored will be representative of the different sizes and range of typical water depths at the project location where piles will be driven with an impact or vibratory hammer.

One bottom-mounted hydrophone will be placed at the nearest distance, approximately 10 meters, from each pile being monitored. An additional hydrophone will be placed at mid-water depth at a distance of 100 to 200 meters from the pile to provide two sound-level readings during ambient and pile driving conditions. A third hydrophone may be deployed at a greater distance (*e.g.*, 1-5 km or further) for the purpose of better defining the long-distance sound propagation. Underwater sound levels will be continuously monitored during the entire duration of each pile being driven. Sound levels will be measured in dB re: 1 μ Pa.

Least Practicable Impact

NMFS PR1 evaluated the FAA's proposed mitigation measures and considered their effectiveness in past implementation to preliminarily determine whether they are likely to effect the least practicable impact on the affected marine mammal species and stocks and their habitat.

Any mitigation measure(s) NMFS PR1 prescribes should be able to accomplish, have a reasonable likelihood of accomplishing (based on current science), or contribute to the accomplishment of one or more of the general goals listed below:

- 1. Avoidance or minimization of injury or death of marine mammals wherever possible (goals 2, 3, and 4 may contribute to this goal);
- 2. A reduction in the number (total number or number at biologically important time or location) of individual marine mammals exposed to stimuli expected to result in incidental take (this goal may contribute to 1, above, or to reducing takes by behavioral harassment only);
- 3. A reduction in the number (total number or number at biologically important time or location) of times any individual marine mammal would be exposed to stimuli expected to result in incidental take (this goal may contribute to 1, above, or to reducing takes by behavioral harassment only);

- 4. A reduction in the intensity of exposure to stimuli expected to result in incidental take (this goal may contribute to 1, above, or to reducing the severity of behavioral harassment only);
- 5. Avoidance or minimization of adverse effects to marine mammal habitat, paying particular attention to the prey base, blockage or limitation of passage to or from biologically important areas, permanent destruction of habitat, or temporary disturbance of habitat during a biologically important time; and
- 6. For monitoring directly related to mitigation, an increase in the probability of detecting marine mammals, thus allowing for more effective implementation of the mitigation.

Based on NMFS PR1's evaluation of the FAA's proposed measures, as well as any other potential measures considered by NMFS, NMFS PR1 preliminarily determined that the proposed mitigation measures provide the means of effecting the least practicable impact on marine mammal species or stocks and their habitat, paying particular attention to rookeries, mating grounds, and areas of similar significance.

Monitoring and Reporting

PR1 requires that observers use approved data forms. Among other pieces of information, the FAA will record detailed information about any implementation of shutdowns, including the distance of animals to the pile and description of specific actions that ensued and resulting behavior of the animal, if any. In addition, the FAA will attempt to distinguish between the number of individual animals taken and the number of incidences of take.

Data Collection

NMFS PR1 requires that, at a minimum, the following information be collected on the sighting forms:

- 1. Date and time monitoring activity begins or ends;
- 2. Construction activities occurring during each observation period;
- 3. Weather parameters (e.g., percent cover, visibility, and sun glare);
- 4. Water conditions (e.g., sea state, tide state);
- 5. Species, group size, age/size/sex categories (if determinable);
- 6. Description of any observable marine mammal behavior patterns, including bearing and direction of travel, and if possible, the correlation to SPLs;
- 7. An estimated total take estimate extrapolated from the number of marine mammals observed during the course of construction activities;
- 8. Distance from pile driving or removal activities to marine mammals and distance from the marine mammals to the observation point;
- 9. Description of implementation of mitigation measures (e.g., shutdown or delay);
- 10. Locations of all marine mammal observations; and
- 11. Other human activity in the area.

Reporting

Interim Reports

Brief, weekly summaries of PSO observations and recorded takes of ESA-listed species will be provided to NMFS AKR during construction. The frequent reports will allow NMFS AKR to track takes (including extrapolated takes) so that authorized take numbers are not exceeded.

End-of –Project Report

A draft report would be submitted to NMFS PR1 within 90 days of the completion of marine mammal monitoring, or 60 days prior to the requested date of issuance of any future IHA for projects at the same location, whichever comes first. The report will include marine mammal observations pre-activity, during-activity, and post-activity during pile driving and removal days, and will also provide descriptions of any behavioral responses to construction activities by marine mammals and a complete description of all mitigation shutdowns and the results of those actions and an extrapolated total take estimate based on the number of marine mammals observed during the course of construction. A final report must be submitted within 30 days following resolution of comments on the draft report.

Unauthorized Take Dead or Injured Marine Mammals

- 1. In the unanticipated event that survey operations clearly cause the take of a marine mammal in a manner prohibited by this IHA, such as a serious injury or mortality (e.g., ship-strike, gear interaction, and/or entanglement), the FAA will immediately the specified activities and immediately report the incident to the Office of Protected Resources, Permits Division, and the Alaska Regional Stranding Coordinator. The report must include the following information:
 - 1.1 Time, date, and location (latitude/longitude) of the incident;
 - 1.2 Description of the incident;
 - 1.3 Environmental conditions (e.g., wind speed and direction, Beaufort sea state, cloud cover, and visibility);
 - 1.4 Description of marine mammal observations in the 24 hours preceding the incident;
 - 1.5 Species identification or description of the animal(s) involved;
 - 1.6 The fate of the animal(s); and
 - 1.7 Photographs or video footage of the animal (if equipment is available).

Activities will not resume until NMFS is able to review the circumstances of the prohibited take. NMFS will work with FAA to determine what is necessary to minimize the likelihood of further prohibited take and ensure MMPA compliance. FAA may not resume their activities until notified by NMFS.

2. In the event that FAA discovers an injured or dead marine mammal, and the lead PSO determines that the cause of the injury or death is unknown and the death is relatively recent (i.e., in less than a moderate state of decomposition), the FAA will immediately report the incident to the Permits Division and the NMFS Alaska Stranding Hotline (1-877-925-7773). The report must include the same information identified above in Unauthorized Take items 1.1-1.7 of this list. Activities may continue while NMFS

- reviews the circumstances of the incident. NMFS will work with FAA to determine whether modifications in the activities are appropriate.
- 3. In the event that the FAA discovers an injured or dead marine mammal, and the lead observer determines that the injury or death is not associated with or related to the activities authorized in the IHA (e.g., previously wounded animal, carcass with moderate to advanced decomposition, scavenger damage), the FAA shall report the incident to the Office of Protected Resources, NMFS, and the Alaska Regional Stranding Coordinator, NMFS, within 24 hours of the discovery. The FAA shall provide photographs or video footage or other documentation of the stranded animal sighting to NMFS.

2.2 Action Area

"Action area" means all areas to be affected directly or indirectly by the Federal action and not merely the immediate area involved in the action (50 CFR 402.02). For this reason, the action area is typically larger than the project area and extends out to a point where no measurable effects from the proposed action occur.

The proposed dock replacement is located on Biorka Island within Symonds Bay near Sitka, Alaska. The action area includes: (1) the area in which construction activities will take place, (2) an ensonified area around the pile removal and installation activities⁶ (see Table 4), and (3) the transit route between Sitka and the project site (see Figure 3).

The action area encompasses approximately 32,000 square kilometers (Figure 3). Symonds Bay is approximately 0.4 miles (0.6 km) wide (east to west direction). Water depths are less than 66 ft (20 m) within 1,300 ft (400 m) of the dock.

Within this area, the loudest sound source with the greatest propagation distance is anticipated to be associated with DTH hydro-hammering. Received levels from DTH hydro-hammering with a source level of 192 dB re 1 uPa 2 's (Quijano and Austin 2017), may be expected on average to decline to 120 dB re 1 µPa (rms) within ~10 km of the pile assuming a sound speed profile with a pronounced surface channel and spherical spreading loss (20 Log R) (see Figures 3-4). The 120 dB isopleth was chosen because that is where we anticipate DTH hydro-hammering noise levels would approach ambient noise levels (i.e. the point where no measurable effect from the project would occur). While project noise may propagate beyond the 120 dB isopleth, we do not anticipate that marine mammals would respond in a biologically significant manner at these low levels and great distance from the source.

On shore at the project site, bedrock is exposed in many places. The overburden varies from 0 to about 15 ft (4.6 m) deep, and consists of highly fractured weathered bedrock and seams of very soft rock or soil. Due to the fractures and seams, it is possible to drive piles into this top layer "Category 1 intensely fractured bedrock." Beneath the top layer, the rock becomes more intact "Category II intensely to moderately fractured bedrock." The seabed composition is important in this project because it determines the pile-driving methods needed to achieve the required pile penetration.

⁶ See Section 3.1.2.3 of this opinion for additional details about how this radius was determined.

The action area includes transit areas for mobilization, and demobilization of construction equipment for vessels. Mobilization and demobilization is anticipated to occur out of Sitka, AK. However, considering that a contractor has not been selected at this point in time, staging areas for operations may vary (Figure 3). Regardless of staging area, the applicant has agreed that all vessels associated with project construction will avoid the 3,000 ft (914 m) aquatic zones surrounding any major Steller sea lion rookery or haulout.

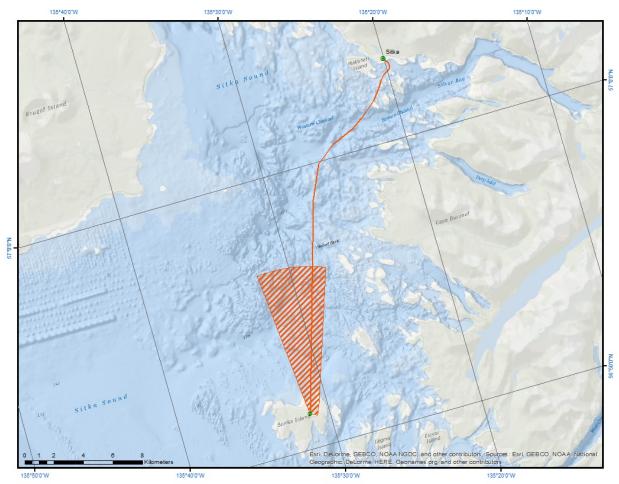
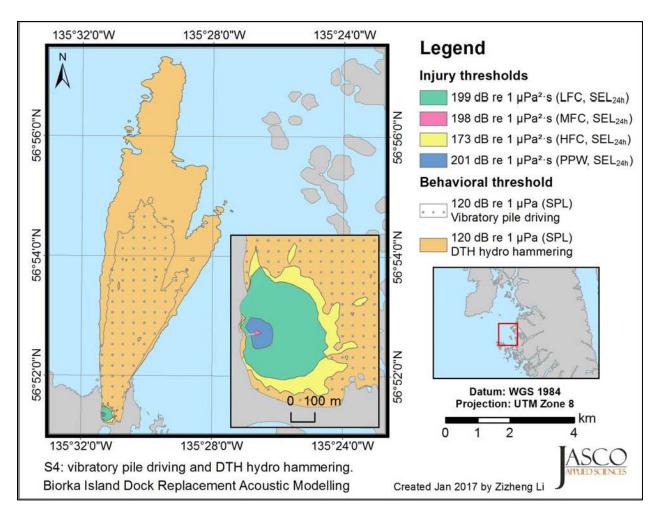


Figure 3. Generalized transit route from Sitka to Biorka Island and ensonified area associated with the loudest project operations (DTH Hammering).



Acoustic modeling results showing thresholds to selected impact criteria for non-impulsive sources in scenario 4. The inset shows a close-up of sound fields around the piles location. The 120 dB isopleth reaches approximately 10 km (Quijano and Austin 2017).

3 APPROACH TO THE ASSESSMENT

Section 7(a)(2) of the ESA requires Federal agencies, in consultation with NMFS, to ensure that their actions are not likely to jeopardize the continued existence of endangered or threatened species, or adversely modify or destroy their designated critical habitat. The jeopardy analysis considers both survival and recovery of the species. The adverse modification analysis considers the impacts to the conservation value of the designated critical habitat.

"To jeopardize the continued existence of a listed species" means to engage in an action that reasonably would be expected, directly or indirectly, to reduce appreciably the likelihood of both the survival and recovery of a listed species in the wild by reducing the reproduction, numbers, or distribution of that species (50 CFR § 402.02). As NMFS explained when it promulgated this definition, NMFS considers the likely impacts to a species' survival as well as likely impacts to its recovery. Further, it is possible that in certain, exceptional circumstances, injury to recovery alone may result in a jeopardy biological opinion [51 FR 19926, 19934 (June 2, 1986)].

Under NMFS's regulations, the destruction or adverse modification of critical habitat "means a direct or indirect alteration that appreciably diminishes the value of critical habitat for the conservation of a listed species. Such alterations may include, but are not limited to, those that alter the physical or biological features essential to the conservation of a species or that preclude or significantly delay development of such features" (50 CFR § 402.02).

The designation of critical habitat for Steller sea lions uses the term primary constituent element (PCE) or essential features. The new critical habitat regulations (81 FR 7414) replace this term with physical or biological features (PBFs). The shift in terminology does not change the approach used in conducting a "destruction or adverse modification" analysis, which is the same regardless of whether the original designation identified PCEs, PBFs, or essential features. In this biological opinion, we use the term PBF to mean PCE or essential feature, as appropriate for the specific critical habitat.

We use the following approach to determine whether the proposed action described in Section 2 is likely to jeopardize listed species or destroy or adversely modify critical habitat:

- Identify those aspects (or stressors) of the proposed action that are likely to have direct or indirect effects on listed species or critical habitat. As part of this step, we identify the action area the spatial and temporal extent of these direct and indirect effects.
- Identify the range-wide status of the species and critical habitat likely to be adversely affected by the proposed action. This section describes the current status of each listed species and its critical habitat relative to the conditions needed for recovery. We determine the range-wide status of critical habitat by examining the condition of its PBFs which were identified when the critical habitat was designated. Species and critical habitat status are discussed in Section 4 of this opinion.
- Describe the environmental baseline including: past and present impacts of Federal, state, or private actions and other human activities *in the action area*; anticipated impacts of

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proposed Federal projects that have already undergone formal or early section 7 consultation, and the impacts of state or private actions that are contemporaneous with the consultation in process. The environmental baseline is discussed in Section 5 of this opinion.

- Analyze the effects of the proposed actions. Identify the listed species that are likely to co-occur with these effects in space and time and the nature of that co-occurrence (these represent our *exposure analyses*). In this step of our analyses, we try to identify the number, age (or life stage), and gender of the individuals that are likely to be exposed to stressors and the populations or subpopulations those individuals represent. NMFS also evaluates the proposed action's effects on critical habitat features. The effects of the action are described in Section 6 of this opinion with the exposure analysis described in Section 6.2 of this opinion.
- Once we identify which listed species are likely to be exposed to an action's effects and the nature of that exposure, we examine the scientific and commercial data available to determine whether and how those listed species are likely to respond given their exposure (these represent our *response analyses*). Response analysis is considered in Section 6.3 of this opinion.
- Describe any cumulative effects. Cumulative effects, as defined in NMFS's implementing regulations (50 CFR § 402.02), are the effects of future state or private activities, not involving Federal activities, that are reasonably certain to occur within the action area. Future Federal actions that are unrelated to the proposed action are not considered because they require separate section 7 consultation. Cumulative effects are considered in Section 7 of this opinion.
- Integrate and synthesize the above factors to assess the risk that the proposed action poses to species and critical habitat. In this step, NMFS adds the effects of the action (Section 6) to the environmental baseline (Section 5) and the cumulative effects (Section 7) to assess whether the action could reasonably be expected to: (1) appreciably reduce the likelihood of both survival and recovery of the species in the wild by reducing its numbers, reproduction, or distribution; or (2) reduce the value of designated or proposed critical habitat for the conservation of the species. These assessments are made in full consideration of the status of the species and critical habitat (Section 4). Integration and synthesis with risk analyses occurs in Section 8 of this opinion.
- Reach jeopardy and adverse modification conclusions. Conclusions regarding jeopardy and the destruction or adverse modification of critical habitat are presented in Section 9. These conclusions flow from the logic and rationale presented in the Integration and Synthesis Section 8.
- If necessary, define a reasonable and prudent alternative to the proposed action. If, in completing the last step in the analysis, NMFS determines that the action under consultation is likely to jeopardize the continued existence of listed species or destroy or adversely modify designated critical habitat, NMFS must identify a reasonable and prudent alternative (RPA) to the action.

4 RANGEWIDE STATUS OF THE SPECIES AND CRITICAL HABITAT

Three species of marine mammals listed under the ESA under NMFS's jurisdiction may occur in the action area. This opinion considers the effects of the proposed action on these species (Table 8). The nearest designated critical habitat for Steller sea lions is Kaiuchali Island located less than one mile southwest of Biorka Island, and outside of the action area.

Table 8. Listing status and critical habitat designation for marine mammals considered in this opinion.

Species	Status	Listing	Critical Habitat
Humpback Whale, Mexico DPS Megaptera novaeangliae	Threatened	NMFS 1970, 35 FR 18319 NMFS 2016 81 FR 62260	Not designated
Sperm Whale Physeter microcephalus	Endangered	NMFS 1970 35 FR 18319	Not designated
Steller Sea Lion, Western DPS Eumetopias jubatus	Endangered	NMFS 1997, 62 FR 24345	1993 58 FR 45269

4.1 Species and Critical Habitats Not Considered Further in this Opinion

NMFS uses two criteria to identify those endangered or threatened species or critical habitat that are likely to be adversely affected. The first criterion is exposure or some reasonable expectation of a co-occurrence between one or more potential stressors associated with FAA's proposed activities and a listed species or designated critical habitat. The second criterion is the probability of a response given exposure.

We applied these criteria to the species and critical habitats listed above and determined that the following species and designated critical habitats are not likely to be adversely affected by the proposed action: sperm whales and Steller sea lion critical habitat.

4.1.1 Sperm Whales

Tagged sperm whales have recently been tracked within the Gulf of Alaska, with one whale tracked along the edge of Sitka Sound during August 2016 (SEASWAP 2017). Tagging studies primarily show that sperm whales use the deep water slope habitat extensively for foraging (Mathias et al. 2012). Interaction studies between sperm whales and the longline fishery have been focused along the continental slope of the eastern Gulf of Alaska in water depths between about 1,970 and 3,280 ft (600 and 1,000 m) (Straley et al. 2005, Straley et al. 2014). The shelf-edge/slope waters of the Gulf of Alaska are far outside of the action area.

Though we do not expect sperm whales will occur in the action area where pile driving activities will occur, it is possible these species may be encountered during transit from staging areas to the construction site off Biorka Island. Therefore, it is possible the species will be at-risk for

vessel strike. However, it is extremely unlikely that vessels will strike sperm whales for the following reasons:

- Few, if any, sperm whales are likely to be encountered because they are generally found in deeper waters than those in which the transit route will occur.
- Project duration is limited to 70 days.
- A limited number of vessels are associated with construction (~5 vessels).
- NMFS's guidelines for approaching marine mammals discourage vessels approaching within 100 yards of marine mammals

For these reasons, we conclude the majority of stressors associated with the proposed action would have no effect on sperm whales because they are not anticipated to overlap in time and space, and the effects of ship strike are discountable because they are extremely unlikely to occur. Therefore, sperm whales are not likely to be adversely affected by this action, and they are not discussed further in this opinion.

4.1.2 Steller Sea Lion Critical Habitat

NMFS designated critical habitat for Steller sea lions on August 27, 1993 (58 FR 45269). The following PBFs were identified at the time of listing:

- 1. Alaska rookeries, haulouts, and associated areas identified at 50 CFR 226.202(a), including:
 - 1.1. Terrestrial zones that extend 914 m (3,000 ft) landward
 - 1.2. Air zones that extend 914 m (3,000 ft) above the terrestrial zone
 - 1.3. Aquatic zones that extend 914 m (3,000 ft) seaward from each major rookery and major haulout east of 144° W. longitude
 - 1.4. Aquatic zones that extend 37 km (23 mi) seaward from each major rookery and major haulout west of 144° W. longitude
- 2. Three special aquatic foraging areas identified at 50 CFR 226.202(c):
 - 2.1. Shelikof Strait
 - 2.2. Bogoslof
 - 2.3. Seguam Pass

The action area does not overlap with designated critical habitat. The nearest critical habitat is Kaiuchali Island (blue dot near Biorka Island in Figure 4; see Figure 1 for zoomed view) located less than one mile southwest of Biorka Island, and outside of the action area. During transit from staging areas to the construction site off Biorka Island, mitigation measures require all vessels associated with construction operations to avoid the 3,000 ft (914 m) aquatic zone surrounding any designated critical habitat. In addition, Kaiuchali Island is located to the southwest of Biorka Island and is not in a direct access path from Sitka or other likely staging areas to the project site.

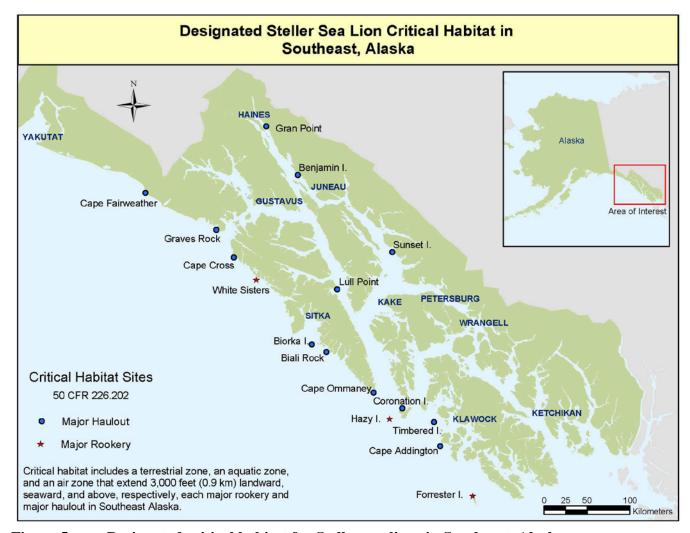


Figure 5. Designated critical habitat for Steller sea lions in Southeast, Alaska.

The transit route will not pass near enough to landmasses to encounter hauled-out pinnipeds; however, foraging sea lions may be encountered during vessel transit through critical habitat surrounding haulouts and rookeries. It is unlikely that vessel transit will impact critical habitat surrounding haulouts and rookeries to any measurable degree considering vessels will avoid designated aquatic zones. We conclude any impacts to these PBFs are likely to be insignificant. Therefore, we conclude Steller sea lion critical habitat is not likely to be adversely affected by this action, and we do not discuss it further in this opinion.

4.2 Climate Change

One threat is or will be common to all of the species we discuss in this opinion: global climate change. Because of this commonality, we present this narrative here rather than in each of the species-specific narratives that follow.

The timeframe for the proposed action is May 1, 2018 through April 30, 2019, which is a relatively short duration. We present potential climate change effects on listed species and their habitat below.

The average global surface temperature rose by 0.85° C from 1880 to 2012, and it continues to rise at an accelerating pace (IPCC 2014b); the 15 warmest years on record since 1880 have occurred in the 21st century, with 2015 being the warmest (NCEI 2016). The warmest year on record for average ocean temperature is also 2015 (NCEI 2016). Since 2000, the Arctic (latitudes between 60 and 90° N) has been warming at more than twice the rate of lower latitudes (Jeffries et al. 2014) due to "Arctic amplification," a characteristic of the global climate system influenced by changes in sea ice extent, atmospheric and oceanic heat transports, cloud cover, black carbon, and many other factors (Serreze and Barry 2011).

Direct effects of climate change include increases in atmospheric temperatures, decreases in sea ice, and changes in sea surface temperatures, oceanic pH, patterns of precipitation, and sea level. Indirect effects of climate change have impacted, are impacting, and will continue to impact marine species in the following ways (IPCC 2014b):

- Shifting abundances
- Changes in distribution
- Changes in timing of migration
- Changes in periodic life cycles of species

Further, ocean acidity has increased by 26 percent since the beginning of the industrial era (IPCC 2013) and this rise has been linked to climate change (Foreman and Yamanaka 2011, GAO 2014, Murray et al. 2014, Okey et al. 2014, Secretariat of the Convention on Biological Diversity 2014, Andersson et al. 2015). Climate change is also expected to increase the frequency of extreme weather and climate events including, but not limited to, cyclones, heat waves, and droughts (IPCC 2014a). Climate change has the potential to impact species abundance, geographic distribution, migration patterns, timing of seasonal activities (IPCC 2014a), and species viability into the future. Climate change is also expected to result in the expansion of low oxygen zones in the marine environment (Gilly et al. 2013). Though predicting the precise consequences of climate change on highly mobile marine species, such as many of those considered in this opinion, is difficult (Simmonds and Isaac 2007), recent research has indicated a range of consequences already occurring.

Climate change is likely to have its most pronounced effects on species whose populations are already in tenuous positions (Isaac 2009). Therefore, we expect the extinction risk of at least some ESA-listed species to rise with global warming. Marine species ranges are expected to shift as they align their distributions to match their physiological tolerances under changing environmental conditions (Doney et al. 2012). Cetaceans with restricted distributions linked to water temperature may be particularly exposed to range restriction (Learmonth et al. 2006, Isaac 2009). Hazen et al. (2012) examined top predator distribution and diversity in the Pacific Ocean in light of rising sea surface temperatures using a database of electronic tags and output from a global climate model. He predicted up to a 35 percent change in core habitat area for some key marine predators in the Pacific Ocean, with some species predicted to experience gains in available core habitat and some predicted to experience losses. MacLeod (2009) estimated, based upon expected shifts in water temperature, 88 percent of cetaceans would be affected by climate change, with 47 percent likely to be negatively affected.

For ESA-listed species that undergo long migrations, if either prey availability or habitat suitability is disrupted by changing ocean temperature regimes, the timing of migration can change or negatively impact population sustainability (Simmonds and Eliott. 2009). Low reproductive success and body condition in humpback whales may have resulted from the 1997/1998 El Niño (Cerchio et al. 2005).

The effects of these changes to the marine ecosystems of the Bering Sea, Aleutian Islands, and the Gulf of Alaska, and how they may affect Steller sea lions are uncertain. Warmer waters could favor productivity of some species of forage fish, but the impact on recruitment of important prey fish of Steller sea lions is unpredictable. Recruitment of large year-classes of gadids (e.g., pollock) and herring has occurred more often in warm than cool years, but the distribution and recruitment of other fish (e.g., osmerids) could be negatively affected (NMFS 2008b).

As temperatures in the Arctic and subarctic waters are warming and sea ice is diminishing, there is an increased potential for harmful algal blooms that produce toxins to affect marine life (see Figure 5). Biotoxins like domoic acid and saxitoxin may pose a risk to marine mammals in Alaska. In addition, increased temperatures can increase Brucella infections. In the Lefebvre et al. (2016) study of marine mammal tissues across Alaska, 905 individuals from 13 species were sampled including humpback whales, bowhead whales, beluga whales, harbor porpoises, northern fur seals, Steller sea lions, harbor seals, ringed seals, bearded seals, spotted seals, ribbon seals, Pacific walruses, and northern sea otters. Domoic acid was detected in all 13 species examined and had a 38% prevalence in humpback whales, and a 27% prevalence in Steller sea lions. Additionally, fetuses from a beluga whale, a harbor porpoise and a Steller sea lion contained detectable concentrations of domoic acid documenting maternal toxin transfer in these species. Saxitoxin was detected in 10 of the 13 species, with the highest prevalence in humpback whales (50%) and a 10% prevalence in Steller sea lions (Lefebvre et al. 2016).



Figure 6. Algal toxins detected in 13 species of marine mammals from southeast Alaska to the Arctic from 2004 to 2013 (Lefebvre et al. 2016).

4.3 Status of Listed Species

This opinion examines the status of each species that is likely to be adversely affected by the proposed action. The status is determined by the level of extinction risk that the listed species face, based on parameters considered in documents such as recovery plans, status reviews, and listing decisions. This informs the description of the species' likelihood of both survival and recovery. The species status section also helps to inform the description of the species' current "reproduction, numbers, or distribution" as described in 50 CFR § 402.02.

This section consists of narratives for each of the endangered and threatened species that occur in the action area and that may be adversely affected by the proposed action. In each narrative, we present a summary of information on the population structure and distribution of each species to provide a foundation for the exposure analyses that appear later in this opinion. Then we summarize information on the threats to the species and the species' status given those threats to provide points of reference for the jeopardy determinations we make later in this opinion. That is, we rely on a species' status and trend to determine whether or not an action's direct or indirect effects are likely to increase the species' probability of becoming extinct.

4.3.1 Mexico DPS Humpback Whale

We used information available in the status review (Bettridge et al. 2015), most recent stock assessment (Muto et al. 2017), NMFS species information (NMFS 2016a), report on estimated abundance and migratory destinations for North Pacific humpback whales (Wade et al. 2016), and recent biological opinions to summarize the status of the species, as follows.

Distribution

Humpback whales are widely distributed in the Atlantic, Indian, Pacific, and Southern Oceans. Individuals generally migrate seasonally between warmer, tropical and sub-tropical waters in winter months (where they reproduce and give birth to calves) and cooler, temperate and sub-Arctic waters in summer months (where they feed). In their summer foraging areas and winter calving areas, they tend to occupy shallower, coastal waters; though during seasonal migrations they disperse widely in deep, pelagic waters and tend to avoid shallower coastal waters (Winn and Reichley 1985).

Humpback whales have been known to occur within the Gulf of Alaska primarily in summer and fall, migrating to southerly breeding grounds in winter and returning to the north in spring (Calambokidis et al. 2008). However, based on recordings from moored hydrophones deployed in six locations in the Gulf of Alaska from October 1999 to May 2002, humpback calls were most commonly detected during the fall and winter (Stafford et al. 2007).

Humpback whales are present in Southeast Alaska in all months of the year. Most Southeast Alaska humpback whales winter in low latitudes, but some individuals have been documented over-wintering near Sitka and Juneau (NPS Fact Sheet available at http://www.nps.gov/glba). Humpback whales are the most commonly observed baleen whale in Sitka Sound and generally throughout Southeast Alaska (ECO49 2017). Late fall and winter whale habitat in Southeast Alaska appears to correlate with areas that have over-wintering herring such as lower Lynn Canal, Tenakee Inlet, Whale Bay, Ketchikan, and Sitka Sound area (Baker 1985, Straley 1990).

Humpback whale occurrence in the action area year round is considered likely. Ferguson et al. (2015) identified areas around Sitka and overlapping with the action area as a Biologically Important Area for humpback whale feeding during summer (June-August) and fall (September-November) (Figure 13).

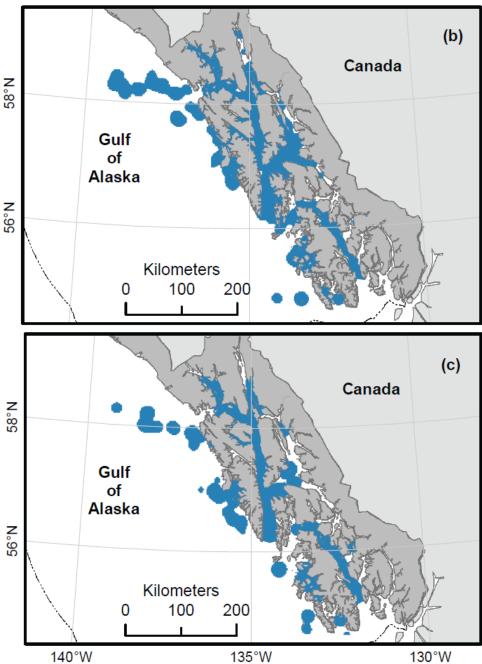


Figure 7. Seasonal humpback whale biologically important feeding areas in Southeast Alaska for (b) summer (June-Augst), and fall (September-November) (Ferguson et al. 2015), showing overlap with the action area.

Life History

Humpback whales are large baleen whales that are primarily dark grey in appearance, with variable areas of white on their fins, bellies, and flukes. The coloration of flukes is unique to individual whales. The lifespan of humpback whales is estimated to be 80 to 100 years. Sexual maturity is reached at five to 11 years of age. The gestation period of humpback whales is 11 months, and calves are nursed for 12 months. The average calving interval is two to three years. Birthing occurs in low latitudes during winter months.

Humpback whale feeding occurs in high latitudes during summer months. They exhibit a wide range of foraging behaviors and feed on a range of prey types, such as small schooling fishes, krill, and other large zooplankton.

Humpback whales produce a variety of vocalizations ranging from 20 Hz to 10 kHz (Winn et al. 1970b, Tyack and Whitehead 1983, Payne and Payne 1985, Silber 1986, Thompson et al. 1986, Richardson et al. 1995, Au 2000, Frazer and Mercado III 2000, Erbe 2002b, Au et al. 2006b, Vu et al. 2012). NMFS categorizes humpback whales in the low-frequency cetacean (i.e., baleen whale) functional hearing group. As a group, it is estimated that baleen whales applied frequency range is between 7 Hz and 35 kHz (NMFS 2016e).

During the breeding season males sing long, complex songs, with frequencies in the 20-5000 Hz range and intensities as high as 181 dB (Payne 1970, Winn et al. 1970a, Thompson et al. 1986). Source levels average 155 dB and range from 144 to 174 dB (Thompson et al. 1979). The songs appear to have an effective range of approximately 10 to 20 km. Animals in mating groups produce a variety of sounds (Tyack 1981).

Social sounds in breeding areas associated with aggressive behavior in male humpback whales are very different than songs and extend from 50 Hz to 10 kHz (or higher), with most energy in components below 3 kHz (Tyack and Whitehead 1983, Silber 1986). These sounds appear to have an effective range of up to 9 km (Tyack and Whitehead 1983).

Humpback whales produce sounds less frequently in their summer feeding areas. Feeding groups produce distinctive sounds ranging from 20 Hz to 2 kHz, with median durations of 0.2-0.8 seconds and source levels of 175-192 dB (Thompson et al. 1986). These sounds are attractive and appear to rally animals to the feeding activity (D'Vincent et al. 1985, Sharpe and Dill 1997).

In summary, humpback whales produce at least three kinds of sounds:

- 1. Complex songs with components ranging from at least 20 Hz–5 kHz with estimated source levels from 144–174 dB; these are mostly sung by males on the breeding grounds (Winn et al. 1970b, Richardson et al. 1995, Au 2000, Frazer and Mercado 2000, Au et al. 2006a);
- 2. Social sounds in the breeding areas that extend from 50Hz more than 10 kHz with most energy below 3kHz (Tyack and Whitehead 1983, Richardson et al. 1995); and
- 3. Feeding area vocalizations that are less frequent, but tend to be 20 Hz–2 kHz with estimated sources levels in excess of 175 dB re 1 Pa at 1m (Thompson et al. 1986, Richardson et al. 1995).

Additional information on humpback whales can be found at: http://www.nmfs.noaa.gov/pr/species/mammals/whales/humpback-whale.html.

Population Dynamics

NMFS recently conducted a global status review and changed the status of humpback whales under the ESA (81 FR 62260; September 8, 2016). Under the final rule, 14 DPSs of humpback whales are recognized worldwide:

- North Atlantic
 - West Indies
 - o Cape Verde Islands/Northwest Africa
- North Pacific
 - Western North Pacific (WNP)
 - o Hawaii
 - o Mexico
 - o Central America
- Northern Indian Ocean
 - o Arabian Sea
- Southern Hemisphere
 - o Brazil
 - o Gabon/Southwest Africa
 - Southeast Africa/Madagascar
 - West Australia
 - o East Australia
 - o Oceania
 - Southeastern Pacific

Humpback whales in the action area may belong to the Mexico or Hawaii DPSs (81 FR 62260).

The worldwide population of all humpback whales is estimated to be approximately 75,000 individuals. Population trends are not available for all humpback whale stocks or populations due to insufficient data, but growth appears to be positive in most areas. The most recent minimum population estimate of the central North Pacific stock is 7,890 whales with an estimated grown rate of 5.5 to 6.0 percent (Muto et al. 2017).

In the final rule changing the status of humpback whales under the ESA (81 FR 62260; September 8, 2016), the abundances of the Mexico and Hawaii DPSs throughout their range were estimated to be 3,264 (CV = 0.06) and 11,398 (CV = 0.04) whales, respectively. The Mexico DPS has an unknown trend. The growth rate of the Hawaii DPS was estimated to be between 5.5 and 6.0 percent.

Within Southeast Alaska and northern British Columbia, the abundance estimate for humpback whales is estimated to be 6,137 (CV= 0.07) animals which includes whales from the Hawaii DPS (93.9%) and Mexico DPS (6.1%) (NMFS 2016c, Wade et al. 2016).

Status

The humpback whale was listed as endangered under the Endangered Species Conservation Act (ESCA) on December 2, 1970 (35 FR 18319). Congress replaced the ESCA with the ESA in 1973, and humpback whales continued to be listed as endangered. NMFS recently conducted a global status review and changed the status of humpback whales under the ESA (81 FR 62260; September 8, 2016). As described in the Population Dynamics Section, humpback whales in the action area may belong to the Mexico or Hawaii DPSs. The Mexico DPS (which includes a small proportion of humpback whales found in the Aleutian Islands, Bering Sea, Chukchi Sea, Beaufort Sea, Gulf of Alaska, and Southeast Alaska) is listed as threatened, and the Hawaii DPS

(which includes most humpback whales found in the Aleutian Islands, Bering Sea, Chukchi Sea, Beaufort Sea, Gulf of Alaska, and Southeast Alaska) is not listed. Whales from these two DPSs overlap on feeding grounds off Alaska, and are not easily distinguishable. All waters off the coast of Alaska may contain ESA-listed humpbacks.

The humpback whale species was originally listed as endangered because of past commercial whaling. Subsistence hunters in Alaska are not authorized to take humpback whales.

Additional threats to the species include ship strikes, fisheries interactions (including entanglement) and noise. All threats to the species are discussed further in Section 5 of this opinion.

Critical Habitat

There is no critical habitat designated for the Mexico DPS of humpback whale.

4.3.2 Western DPS Steller Sea Lion

We used information available in the recent stock assessment report (Muto et al. 2017), recovery plan (NMFS 2008a), the status review (NMFS 1995), listing document (62 FR 24345), NMFS species information, and recent biological opinions to summarize the status of the species, as follows.

Distribution

Steller sea lions are distributed throughout the northern Pacific Ocean, including coastal and inland waters in Russia (Kuril Islands and the Sea of Okhotsk), east to Alaska, and south to central California (Año Nuevo Island) (Figure 8). Animals from the eastern DPS occur primarily east of Cape Suckling, Alaska (144° W) and animals from the endangered western DPS occur primarily west of Cape Suckling. The western DPS includes Steller sea lions that reside primarily in the central and western Gulf of Alaska, Aleutian Islands, and those that inhabit and breed in the coastal waters of Asia (e.g., Japan and Russia). The eastern DPS includes sea lions living primarily in southeast Alaska, British Columbia, California, and Oregon.

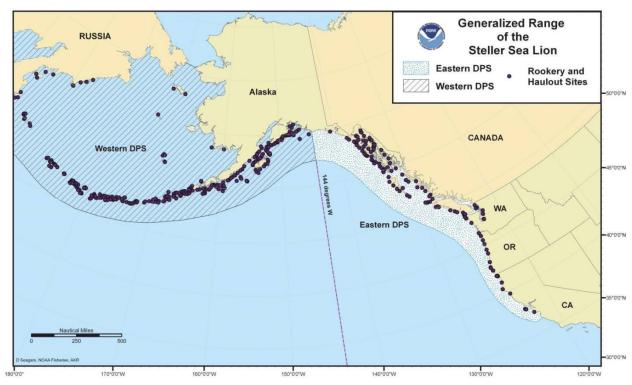


Figure 8. Generalized range of Steller sea lion, including rookery and haulout locations.

Southeast Alaska Distribution

Within the action area, Steller sea lions are anticipated to be predominantly from the eastern DPS. However, studies have confirmed movement of animals across the 144° W longitude boundary (Raum-Suryan et al. 2002, Pitcher et al. 2007, Fritz et al. 2013, Jemison et al. 2013). Jemison et al. (2013) found regularly occurring temporary movements of western DPS Steller sea lions across the 144° W longitude boundary, and some western DPS females have likely emigrated permanently and given birth at White Sisters and Graves rookeries. The vast majority of these sightings have been in northern Southeast Alaska, north of Frederick Sound (the action area is also in northern Southeast Alaska). Fritz et al. (2013) estimated an average annual breeding season movement of western DPS Steller sea lions to southeast Alaska of 917 animals. For this consultation we assume that 50 percent of the Steller sea lions observed in the action area may be from the western DPS (ECO49 2017).

Steller sea lions occur in coastal and nearshore habitats of Sitka Sound, and forage on herring and salmon throughout the Sound. Figure 9 depicts a likely seasonal foraging strategy for Steller sea lions in Southeast Alaska. These results suggest that seasonally aggregated high-energy prey species, such as eulachon and herring in late spring and salmon in summer and fall, influence the seasonal distribution of Steller sea lions in some areas of Southeast Alaska (Womble et al. 2009).

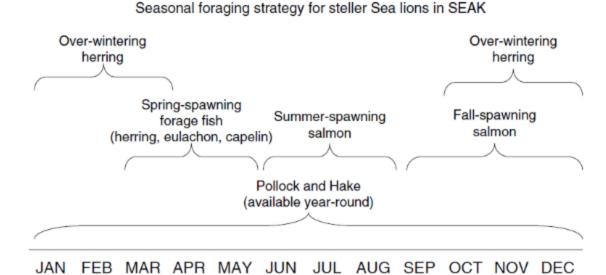


Figure 9. Seasonal foraging ecology of Steller sea lions in Southeast Alaska (Womble et al. 2009).

Steller sea lions are present in Sitka Sound in very low numbers over the summer months when construction is planned, during the interval between herring spawning and the return of adult herring to Sitka Sound (E. Majeski, pers. comm. 2017). Prey availability for Steller sea lions in Sitka Sound is limited during this period as compared to other seasons, and they are generally only observed as individuals or in small groups of three to five animals (E. Majeski, pers. comm. 2017).

Sea lions from both DPSs occur in the action area on a year-round basis. Kaiuchali Island is used as a sea lion rookery in spring-summer (late May to early July) and as a haulout during the non-breeding seasons (Fritz et al. 2016). Based on results of recent aerial surveys, there has been an increase of sea lions that use Kaiuchali Island during both the breeding and non-breeding seasons. In June 2013, Fritz et al. (2016) documented 22 individuals, none of which were pups. In June 2015, the same study recorded 77 Steller sea lions, including one pup. This limited information shows an increase in the numbers of animals at this location and indicates that the site has become a recently-established breeding site (ECO49 2017). The location of Kaiuchali Island is outside the action area, opposite Biorka Island (see Figures 1-2).

Kaiuchali Island is also used as a haulout during the non-breeding season. The utilization of adult, over- wintering herring in Sitka Sound by Steller sea lions has a substantial seasonal significance on the energy budget of sea lions at this time because herring are energy-rich (Iverson et al. 2002) and the overwintering aggregation occurs during a period of high-energetic demands for Steller sea lions (Winship and Trites 2003, Sigler et al. 2004). The late fall and overwintering aggregation of adult herring results in hundreds of animals using Kaiuchali Island as a haulout during this period. However, the construction period for the proposed action would not overlap with the overwintering aggregations of sea lions.

Life History

Steller sea lions are the largest of the eared seals (Otariidae), though there is significant difference in size between males and females: males reach lengths of 3.3 m (10.8 ft) and can weigh up to 1,120 kg (2,469 lb) and females reach lengths of 2.9 m (9.5 ft) and can weigh up to 350 kg (772 lb). Their fur is light buff to reddish brown and slightly darker on the chest and abdomen; their skin is black. Sexual maturity is reached and fist breeding occurs between 3 and 8 years of age. Pupping occurs on rookeries between May and June and females breed 11 days after giving birth. Implantation of the fertilized egg is delayed for about 3.5 months, and gestation occurs until the following May or June.

Most adult Steller sea lions occupy rookeries during pupping and breeding season (late Mayearly July). During the breeding season, most juvenile and non-breeding adults are at haulouts, though some occur at or near rookeries. Adult females and pups continue to stay on rookeries through August beginning a regular routine of alternating foraging trips at sea with nursing their pups on land. During the non-breeding season many Steller sea lions disperse from rookeries and increase their use of haulouts. Steller sea lions do not migrate, but they often disperse widely outside of the breeding season (Loughlin 1997). At sea, Steller sea lions commonly occur near the 200 m (656 ft) depth contour, but have been seen from near shore to well beyond the continental shelf (Kajimura and Loughlin 1988).

The ability to detect sound and communicate underwater and in-air is important for a variety of Steller sea lion life functions, including reproduction and predator avoidance. NMFS categorizes Steller sea lions in the otariid pinniped functional hearing group with an applied frequency range between 60 and 39 kHz in water (NMFS 2016e).

Additional information on Steller sea lions can be found at: https://alaskafisheries.noaa.gov/pr/steller-sea-lions.

Population Dynamics

The western DPS population declined approximately 75 % from 1976 to 1990 (the year of ESA-listing). Since 2000, the abundance of the western DPS has increased, but there has been considerable regional variation in trend (Muto et al. 2017). The minimum population estimate of western DPS Steller sea lions in Alaska is 50,983 individuals. Using data collected through 2015, there is strong evidence that non-pup and pup counts of western DPS Steller sea lions in Alaska increased at ~2% per year between 2000 and 2015 (Muto et al. 2017). Populations in the eastern Gulf of Alaska (closest to the action area) are increasing at an average rate of 5.07% for non-pups and 4.31% for pups annually (Muto et al. 2017).

Status

The Steller sea lion was listed as a threatened species under the ESA on November 26, 1990 (55 FR 49204). In 1997, NMFS reclassified Steller sea lions as two DPSs based on genetic studies and other information (62 FR 24345); at that time the eastern DPS was listed as threatened and the western DPS was listed as endangered. On November 4, 2013, the eastern DPS was removed from the endangered species list (78 FR 66139). Factors affecting the continued existence of the western DPS at the time of its listing included changes in the availability or quality of prey as a result of environmental changes or human activities and removals of Steller sea lions from the wild. Concern about possible adverse effects of contaminants was also noted.

Steller sea lions are hunted for subsistence purposes. As of 2009, data on community subsistence harvest are no longer being consistently collected; therefore, the most recent estimate of annual statewide (excluding St. Paul Island) harvest⁷ is 172.3 individuals from the 5-year period from 2004 to 2008. More recent data from St. Paul are available; the annual harvest is 29 sea lions from the 5-year period from 2010 to 2014. This results in a total take of 201 individuals (Muto et al. 2017). In addition, data were collected on Alaska Native harvest of Steller sea lions for 7 communities on Kodiak Island for 2011 and 15 communities in Southcentral Alaska in 2014; the Alaska Native Harbor Seal Commission and ADFG estimated a total of 20 adult sea lions were harvested on Kodiak Island in 2011, and 7.9 sea lions (CI = 6-15.3) were harvested in Southcentral Alaska in 2014, with adults comprising 84% of the harvest (Muto et al. 2017).

Additional threats to the species include environmental variability, competition with fisheries, predation by killer whales, toxic substances, incidental take due to interactions with active fishing gear, illegal shooting, entanglement in marine debris, disease and parasites, and disturbance from vessel traffic, tourism, and research activities. All threats to the species in the action area are discussed further in Section 5 of this opinion.

Critical Habitat

NMFS designated critical habitat for Steller sea lions on August 27, 1993 (58 FR 45269). More information about critical habitat can be found in Section 4.1.2 of this opinion.

5 ENVIRONMENTAL BASELINE

The "environmental baseline" includes the past and present impacts of all Federal, state, or private actions and other human activities in the action area, the anticipated impacts of all proposed Federal projects in the action area that have already undergone formal or early section 7 consultation, and the impact of state or private actions which are contemporaneous with the consultation in process (50 CFR § 402.02).

Focusing on the impacts of activities specifically within the action area allows us to assess the prior experience and condition of the animals that will be exposed to effects from the actions under consultation. This focus is important because individuals of ESA-listed species may commonly exhibit, or be more susceptible to, adverse responses to stressors in some life history states, stages, or areas within their distributions than in others. These localized stress responses or baseline stress conditions may increase the severity of the adverse effects expected from proposed actions.

5.1 Factors Affecting Species within the Action Area

A number of human activities have contributed to the current status of populations of ESA-listed species in the action area. The factors that have likely had the greatest impact are discussed in the sections below. For more information on all factors affecting the ESA-listed species considered in depth in this opinion, please refer to the following documents:

• "Alaska Marine Mammal Stock Assessments, 2016" (Muto et al. 2017).

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⁷ These numbers included both harvested and struck and lost sea lions.

- o Available online at http://www.nmfs.noaa.gov/pr/sars/pdf/ak_2016_final_sars_june.pdf
- "Recovery Plan for the Steller Sea Lion, Eastern and Western Distinct Population Segments (*Eumetopias jubatus*)" (NMFS 2008a)
 - o Available online at https://alaskafisheries.noaa.gov/sites/default/files/sslrpfinalrev030408.pdf
- "Final Recovery Plan for the Fin Whale (Balaenoptera physalus)" (NMFS 2010)
 - o Available online at http://www.nmfs.noaa.gov/pr/pdfs/recovery/finwhale.pdf
- "Status Review of the Humpback Whale (*Megaptera novaeangliae*)" (Bettridge et al. 2015)
 - Available online at http://www.nmfs.noaa.gov/pr/species/Status%20Reviews/humpback_whale_sr_2015.pdf

5.1.1 Climate Change

Overwhelming data indicate the planet is warming (IPCC 2014a), which poses a threat to most Arctic and Subarctic marine mammals.

Climate change has the potential to impact species abundance, geographic distribution, migration patterns, timing of seasonal activities (IPCC 2014a), and species viability into the future. Climate change is also expected to result in the expansion of low oxygen zones in the marine environment (Gilly et al. 2013). Though predicting the precise consequences of climate change on highly mobile marine species, such as many of those considered in this opinion, is difficult (Simmonds and Isaac 2007), recent research has indicated a range of consequences already occurring.

The indirect effects of climate change would result from changes in the distribution of temperatures suitable for the distribution and abundance of prey and the distribution and abundance of competitors or predators. For example, variations in the localized recruitment of herring in or near the action area caused by climate change could change the distribution and localized abundance of humpback whales. However, we have no information to indicate that this has happened to date. Warmer waters could favor productivity of some species of forage fish, but the impact on recruitment of important prey fish of Steller sea lions is unpredictable. Recruitment of large year-classes of gadids (e.g., pollock) and herring has occurred more often in warm than cool years, but the distribution and recruitment of other fish (e.g., osmerids) could be negatively affected (NMFS 2008a).

5.1.2 Fisheries

Worldwide, fisheries interactions have an impact on many marine mammal species. More than 97 percent of whale entanglement is caused by derelict fishing gear (Baulch and Perry 2014). There is also concern that mortality from entanglement may be underreported, as many marine mammals that die from entanglement tend to sink rather than strand ashore. Entanglement may also make marine mammals more vulnerable to additional dangers, such as predation and ship strikes, by restricting agility and swimming speed. The estimated mean annual mortality and serious injury rate for humpback whales in Southeast Alaska from salmon drift gillnet fishery is 5.5 whales in 2012-2013 (Muto et al. 2017).

Commercial fisheries may indirectly affect whales and pinnipeds by reducing the amount of available prey or affecting prey species composition. In Alaska, commercial fisheries target known prey species of ESA-listed whales and pinnipeds, such as pollock and cod.

The most recent minimum average annual estimated mortality and serious injury rate of western DPS Steller sea lions associated with observed commercial fisheries is 30 individuals (Muto et al. 2017). The minimum average annual mortality and serious injury rate for all fisheries, based on observer data (30 sea lions) for commercial fisheries and stranding data (1.6 sea lions) for unknown (commercial, recreational, or subsistence) fisheries is 32 western DPS Steller sea lions (Muto et al. 2017).

5.1.3 Harvest

Commercial whaling in the 19th and 20th centuries removed tens of thousands of whales from the North Pacific Ocean. As discussed in Section 4.3.1 of this opinion, commercial harvest was the primary factor for ESA-listing of humpback whales. This historical exploitation has impacted populations and distributions of humpback whales in the action area, and it is likely these impacts will continue to persist into the future.

Subsistence hunters in Alaska have reported one subsistence take of a humpback whale in South Norton Sound in 2006. There had not been any additional reported takes of humpback whales from this stock by subsistence hunters in Alaska or Russia until 2016 when hunters illegally harvested one near Toksook Bay in May (DeMarban and Demer 2016).

As of 2009, data on community subsistence harvest are no longer being collected for Steller sea lion; therefore, the most recent estimate of annual statewide (excluding St. Paul Island) harvest is 172 individuals from the 5-year period from 2004 to 2008. More recent data from St. Paul are available; the mean annual harvest is 29 sea lions from the 5-year period from 2010 to 2014 for a total of 201 Steller sea lions/year (Muto et al. 2017).

5.1.4 Natural and Anthropogenic Noise

ESA-listed species in the action area are exposed to several sources of natural and anthropogenic noise. Natural sources of underwater noise include sea ice, wind, waves, precipitation, and biological noise from marine mammals, fishes, and crustaceans. Anthropogenic sources of noise in the action area include:

- Vessels
 - Shipping
 - o Transportation
 - o Research
- Construction activities:
 - o Drilling
 - o Dredging
 - o Pile-driving
- Sonars
- Aircraft

⁸ These numbers included both harvested and struck and lost sea lions.

The combination of anthropogenic and natural noises contributes to the total noise at any one place and time.

Because responses to anthropogenic noise vary among species and individuals within species, it is difficult to determine long-term effects. Habitat abandonment due to anthropogenic noise exposure has been found in terrestrial species (Francis and Barber 2013). Clark et al. (2009a) identified increasing levels of anthropogenic noise as a habitat concern for whales because of its potential effect on their ability to communicate (i.e. masking). Some research (Parks 2003, McDonald et al. 2006, Parks 2009) suggests marine mammals compensate for masking by changing the frequency, source level, redundancy, and timing of their calls. However, the long-term implications of these adjustments, if any, are currently unknown.

5.1.4.1 *Noise Related to Construction Activities*

NMFS has conducted numerous ESA section 7 consultations related to construction activities in Southeast Alaskan waters. Many of the consultations have authorized the take (by harassment) of humpback whales and Steller sea lions from sounds produced during pile driving, drilling, and vessel operations.

In 2017, NMFS conducted three consultations with the Permits Division on the issuance of IHAs to take marine mammals incidental to dock and ferry terminal construction in Southeast Alaska (Sawmill Cove Dock, Gustavus Ferry Terminal, and Haines Ferry Terminal). The incidental take statements in the three biological opinions estimated 797 western DPS Steller sea lions and 45 Mexico DPS humpback whales, total, would be taken (by Level B harassment) as a result of exposure to continuous sounds at received levels at or above 120 dB re 1 μ Pa_{rms} and impulsive sounds at received levels at or above 160 dB re 1 μ Pa_{rms}. Only one Level A harassment of a Mexico DPS humpback whale was authorized.

Anticipated impacts by harassment from noise associated with construction activities generally include changes in behavioral state from low energy states (i.e., foraging, resting, and milling) to high energy states (i.e., traveling and avoidance).

5.1.5 Pollutants and Discharges

Previous development and discharges in portions of the action area are the source of multiple pollutants that may be bioavailable (i.e., may be taken up and absorbed by animals) to ESA-listed species or their prey items (NMFS 2013a).

The Clean Water Act of 1972 (CWA) has several sections or programs applicable to activities in offshore waters. Section 402 of the CWA authorizes the U.S. Environmental Protection Agency (EPA) to administer the National Pollutant Discharge Elimination System (NPDES) permit program to regulate point source discharges into waters of the United States. Section 403 of the CWA requires that EPA conduct an ocean discharge criteria evaluation for discharges to the territorial seas, contiguous zones, and the oceans. The Ocean Discharge Criteria (40 CFR Part 125, Subpart M) sets forth specific determinations of unreasonable degradation that must be made before permits may be issued.

The EPA issued a NPDES vessel general permit that authorizes several types of discharges incidental to the normal operation of vessels, such as grey water, black water, coolant, bilge water, ballast, and deck wash (EPA 2013). The permit is effective from December 19, 2013 to December 19, 2017, and applies to owners and operators of non-recreational vessels that are at least 24 m (79 ft) in length, as well as to owners and operators of commercial vessels less than 24 m that discharge ballast water.

The US Coast Guard has regulations related to pollution prevention and discharges for vessels carrying oil, noxious liquid substances, garbage, municipal or commercial waste, and ballast water (33 CFR Part 151). The State of Alaska regulates water quality standards within three miles of the shore.

5.1.6 Scientific Research

In the following sections, we describe the types of scientific research currently permitted for ESA-listed whales and sea lions in the action area. NMFS issues scientific research permits that are valid for five years for ESA-listed species. When permits expire, researchers often apply for a new permit to continue their research. Additionally, applications for new permits are issued on an on-going basis; therefore, the number of active research permits is subject to change in the period during which this opinion is valid.

Species considered in this opinion also occur in Canadian waters. Although we do not have specific information about any permitted research activities in Canadian waters, we assume they will be similar to those described below.

5.1.6.1 Whales

Humpback whales are exposed to research activities documenting their distribution and movements throughout their ranges. There are 16 active research permits authorizing takes of humpback whales in Alaskan waters (NMFS 2016b). Activities associated with these permits could occur in the action area, possibly at the same time as the proposed project activities. Currently permitted research activities include:

- Counting/surveying
- Opportunistic collection of sloughed skin and remains
- Behavioral and monitoring observations
- Various types of photography and videography
- Skin and blubber biopsy sampling
- Fecal sampling
- Suction-cup, dart/barb, satellite, and dorsal fin/ridge tagging

These research activities require close vessel approach. The permits also include incidental harassment takes to cover such activities as tagging, where the research vessel may come within 91 m (300 ft) of other whales while in pursuit of a target whale. These activities may cause stress to individual whales and cause behavioral responses, but harassment is not expected to rise to the level where injury or mortality is expected to occur.

5.1.6.2 *Pinnipeds*

Steller sea lions are exposed to research activities documenting their distribution and movements throughout their ranges.

Out of the 16 active research permits, two permits (Permit Nos. 15142 and 15324) include behavioral observations, counting/surveying, photo-identification, and capture and restraint (by hand, net, cage, or board), for the purposes of performing the following procedures:

- Collection of:
 - o Blood
 - o Clipped hair
 - o Urine and feces
 - o Nasal and oral swabs
 - o Vibrissae (pulled)
 - o Skin, blubber, or muscle biopsies
 - Weight and body measurements
- Injection of sedative
- Administration of drugs (intramuscular, subcutaneous, or topical)
- Attachment of instruments to hair or flippers, including flipper tagging
- Ultrasound

Permit Nos. 15142 and 15324 also include incidental harassment of non-target seals during the course of performing the permitted activities. Two additional permits (Permits Nos. 14610 and 18537) include harassment takes of bearded and ringed seals incidental to permitted research activities, and targeting bowhead whales and western DPS Steller sea lions respectively.

Activities may cause stress to individual seals, but, in most cases, harassment is not expected to rise to the level where injury or mortality is expected to occur.

5.1.7 Vessel Interactions

Ship strikes and other interactions with vessels unrelated to fisheries occur frequently with humpback whales. Neilson et al. (2012) summarized 108 large whale ship-strike events in Alaska from 1978 to 2011, 25 of which are known to have resulted in the whale's death. Eighty-six percent of these reports involved humpback whales. The minimum mean annual mortality and serious injury rate due to ship strikes reported in Alaska is 2.7 Central North Pacific humpback whales per year between 2010 and 2014. Most vessel collisions with humpbacks are reported from Southeast Alaska (Muto et al. 2017).

There are three documented occurrences of Steller sea lions being struck by vessels in Southeast Alaska; all were near Sitka. Although risk of ship strike has not been identified as a significant concern for Steller sea lions (Loughlin and York 2000), the recovery plan for this species states that Steller sea lions may be more susceptible to ship strike mortality or injury in harbors or in areas where animals are concentrated (e.g., near rookeries or haulouts) (NMFS 2008a).

The 3-mile no transit zones are established and enforced around rookeries in the area for further protection, and NMFS's guidelines for approaching marine mammals discourage vessels approaching within 100 yards of haulout locations. There is no designated critical habitat for Steller sea lions within the action area.

5.1.8 Environmental Baseline Summary

Historically, overexploitation of large whales caused declines in abundance to the point of near-extinction. There is no commercial whaling of humpback whales currently. Mexico DPS humpback whale trend is unknown.

The relationship between sound and marine mammal response to sound is the topic of extensive scientific research and public inquiry. Most observations report only short-term behavioral responses that include cessation of feeding, resting, or social interactions because study design precludes detection of difficult-to-detect long-term effects, if any exist. However, behavioral response could take the form of habitat abandonment, which could have implications at the population level.

Humpback whales and western DPS Steller sea lions in the action area appear to be increasing in population size – or, at least, their population sizes do not appear to be declining – despite their continued exposure to the direct and indirect effects of the activities discussed in the Environmental Baseline. While we do not have trend information for the Mexico DPS of humpback whales, they also do not appear to be declining as a result of the current stress regime.

6 EFFECTS OF THE ACTION

"Effects of the action" means the direct and indirect effects of an action on the species or critical habitat, together with the effects of other activities that are interrelated or interdependent with that action, that will be added to the environmental baseline (50 CFR 402.02). Indirect effects are those that are caused by the proposed action and are later in time, but still are reasonably certain to occur.

This biological opinion relies on the best scientific and commercial information available. We try to note areas of uncertainty, or situations where data is not available. In analyzing the effects of the action, NMFS gives the benefit of the doubt to the listed species by minimizing the likelihood of false negative conclusions (concluding that adverse effects are not likely when such effects are, in fact, likely to occur).

We organize our effects analysis using a stressor identification – exposure – response – risk assessment framework for the proposed activities.

We conclude this section with an Integration and Synthesis of Effects that integrates information presented in the Status of the Species and Environmental Baseline sections of this opinion with the results of our exposure and response analyses to estimate the probable risks the proposed action poses to endangered and threatened species.

6.1 Stressors

During the course of this consultation, we identified the following potential stressors from the proposed activities:

- Vessel strike
- Disturbance of seafloor
- Underwater sounds from:
 - o Vessels

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- o Pile Driving and Pile Removal
- o Down-the-Hole Hammering

Below we discuss each stressor's potential to affect ESA-listed species.

6.1.1 Stressors Not Likely to Adversely Affect ESA-listed Species

Based on a review of available information, we determined which of the possible stressors may occur, but for which the likely effects are discountable or insignificant and, therefore, need not be evaluated further in this opinion.

6.1.1.1 Vibratory and Impact Pile Driving Airborne Noise

Airborne noises could affect hauled out pinnipeds. However, noise generated during vibratory pile driving would reach the harbor seal in-air threshold (90 db) at approximately 112 ft (34 m) and is below the threshold for other pinnipeds. The in-air threshold for driving 30-in diameter pipe (110 dB at 50 ft) would reach the harbor seal threshold (90 dB) at approximately 500 ft (152 m), and the threshold for other pinnipeds (e.g., Steller sea lions) (100 dB) at approximately 158 ft (48 m).

Pile driving will not exceed in-air disturbance thresholds for hauled-out pinnipeds and there are no haulout sites close enough that sound at the site could be detected. Therefore, during impact pile driving, temporary in-air disturbance would be limited to harbor seals and sea lions swimming on the surface through the immediate action area near the dock (approximately 500 ft (152 m), and within 158 ft (48 m), respectively. At this distance, any animal swimming would already have been taken by the in- water noise levels; therefore, in-air disturbance is generally not considered for pinnipeds swimming near the project site. Further, proposed mitigation would prevent a take from occurring at these distances (see Section 2.1.5) or cause serious injury. For these reasons, effects from in-air noise are considered discountable (i.e., no haulouts nearby), and insignificant (i.e., shutdown mechanisms in place) for ESA-listed pinnipeds.

6.1.1.2 Tension Anchors

Up to eight of the dock piles and all six piles for the dolphins would require internal tension anchors. Noise associated with drilling a 5-in. diameter hole extending about 30 to 40 ft into bedrock below the tip of the pile is anticipated to be contained entirely within the piling and is not anticipated to reach or exceed the 120 dB threshold for continuous noise sources (McLean, pers. comm. 2017). Given the small size of the anchoring drill, the installation method within a pile, and the low anticipated source level, the effects of tension anchor installation are not anticipated to reach the level at which take could occur and are considered insignificant.

6.1.1.3 Vessel Strike

The possibility of vessel strike is extremely unlikely. While there have been a limited number of vessel strikes reported near Sitka, these involved vessels transiting at high speeds. During construction activities vessel speed will be very low (i.e., 2 km/hr [1 kt] or less), and the maximum transit speed for tug and barge vessels proposed for use is 18.5 km/hr (8-10 kts). Once vessels get to the construction site, they will be anchored. FAA considers potential effects from vessels to be negligible considering the size and use of the facilities on Biorka Island would not change as a result of dock reconstruction (FAA 2017a). Required mitigation measures described in Section 2.1.5 require all vessels associated with project construction to avoid the 3,000 ft (914

m) designated aquatic zones surrounding major rookeries or haulout locations east of 144°W longitude. In addition NMFS's guidelines for approaching marine mammals recommends that vessels not approach within 100 yards of marine mammals. There are approximately five vessels associated with construction, and operations will not increase vessel traffic. Finally, construction will only last ~70 days. All of these factors limit the risk of strike. We conclude the probability of strike occurring is extremely unlikely and therefore effects are discountable.

6.1.1.4 *Disturbance of Seafloor*

Short-term turbidity increases would likely occur during in-water construction work, including pile driving, and pile removal. The physical resuspension of sediments could produce localized turbidity plumes that could last from a few minutes to several hours. In general, turbidity associated with pile installation is expected to be localized to about a 25 ft radius around the pile (Everitt et al. 1980). Contaminated sediments are not expected at the project site but any that do occur would be tightly bound to the sediment matrix. Because of the relatively small work area, any increase in turbidity would be limited to the immediate vicinity of the project site and adjacent portion of the bay. There is little potential for pinnipeds to be exposed to increased turbidity during construction operations. Therefore, exposure to re-suspended contaminants is expected to be negligible since sediments would not be ingested and any contaminants would be tightly bound to them (ECO49 2017).

Considering local currents, tidal action and implementation of BMPs, any potential water quality exceedances would likely be temporary and highly localized. The local tides and currents would disperse suspended sediments from pile driving operations at a moderate to rapid rate depending on tidal stage (ECO49 2017).

Cetaceans are not expected to come close enough to the Biorka dock site to encounter increased turbidity from construction activities. Any pinnipeds would avoid the short-term, localized areas of turbidity. Therefore, the impact from increased turbidity levels would be negligible to marine mammals and would not cause a significant disruption of behavioral patterns that would rise to the level of harassment. Therefore, we conclude the effects from this stressor are insignificant.

6.1.1.5 Summary of Stressors Not Likely to Adversely Affect ESA-listed Species In conclusion, based on review of available information, we determined effects from in-air noise and vessel strike are extremely unlikely to occur. We consider the effects to ESA-listed whales and pinnipeds to be discountable.

We determined tension anchor installation and disturbance of seafloor are not likely to have measurable impact; therefore, we consider effects to ESA-listed whales and pinnipeds to be insignificant.

6.1.2 Stressors Likely to Adversely Affect ESA-listed Species

The following sections analyze the stressors likely to adversely affect ESA-listed species: underwater sounds from pile removal, pile installation, and DTH hydro-hammering. First, we present a brief explanation of the sound measurements used in the discussions of acoustic effects in this opinion.

6.1.2.1 Sound Measurements Used in this Document

"Sound pressure" is the sound force per unit micropascals (µPa), where 1 pascal (Pa) is the

pressure resulting from a force of one newton exerted over an area of one square meter. "Sound pressure level" is expressed as the ratio of a measured sound pressure and a reference level. The commonly used reference pressure in underwater acoustics is 1 μ Pa, and the units for sound pressure levels are decibels (dB) re 1 μ Pa. Sound pressure level (in dB) = 20 log (pressure/reference pressure).

Sound pressure level is an instantaneous measurement and can be expressed as "peak" (0-p), "peak-to-peak" (p-p), or "root mean square" (rms). Root mean square, which is the square root of the arithmetic average of the squared instantaneous pressure values, is typically used in discussions of the effects of sounds on vertebrates. All references to sound pressure level in this document are expressed as rms, unless otherwise indicated. In instances where sound pressure levels for airguns or other impulsive sources were originally expressed as 0-p or p-p, we used the following rough conversions in order to express those values in rms (Harris et al. 2001):

- rms is approximately 10 dB lower than 0-p
- rms is approximately 16 dB lower than p-p

Note that sound pressure level does not take the duration of a sound into account.

6.1.2.2 Acoustic Thresholds

As discussed in Section 2, *Description of the Proposed Action*, FAA intends to use a wide variety of acoustic systems in the action area (see Table 3).

Since 1997, NMFS has used generic sound exposure thresholds to determine whether an activity produces underwater and in-air sounds that might result in impacts to marine mammals (70 FR 1871). NMFS recently developed comprehensive guidance on sound levels likely to cause injury to marine mammals through onset of permanent and temporary thresholds shifts (PTS and TTS; Level A harassment) (81 FR 51693). NMFS is in the process of developing guidance for behavioral disruption (Level B harassment). However, until such guidance is available, NMFS uses the following conservative thresholds of underwater sound pressure levels, expressed in rms, from broadband sounds that cause behavioral disturbance, and referred to as Level B harassment under section 3(18)(A)(ii) of the Marine Mammal Protection Act (MMPA):

- impulsive sound: 160 dB re 1 μPa_{rms}
- continuous sound: 120 dB re 1μPa_{rms}

Under the PTS/TTS Technical Guidance, NMFS uses the following thresholds for underwater sounds that cause injury, referred to as Level A harassment under section 3(18)(A)(i) of the MMPA (NMFS 2016e). These acoustic thresholds are presented using dual metrics of cumulative sound exposure level (L_E) and peak sound level (PK) for impulsive sounds 9 and L_E for non-impulsive sounds (see Table 9):

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⁹ For the dual metric associated with impulsive sources, the applicant must consider whichever threshold results in the largest effect distance (isopleth) (NMFS 2016e).

Table 9. PTS Onset Acoustic Thresholds for Level A Harassment (NMFS 2016e).

	PTS Onset Acoustic Thresholds* (Received Level)				
Hearing Group	Impulsive	Non-impulsive			
Low-Frequency (LF) Cetaceans	<i>L</i> _{pk,flat} : 219 dB <i>L</i> _{E,LF,24h} : 183 dB	<i>L</i> _{E,LF,24h} ∶199 dB			
Mid-Frequency (MF) Cetaceans	<i>L</i> _{pk,flat} : 230 dB <i>L</i> _{E,MF,24h} : 185 dB	<i>L</i> е,мғ,24h: 198 dВ			
High-Frequency (HF) Cetaceans	<i>L</i> _{pk,flat} : 202 dB <i>L</i> _{E,HF,24h} : 155 dB	L е,нғ,24h: 173 dВ			
Phocid Pinnipeds (PW) (Underwater)	$L_{ m pk,flat}$: 218 dB $L_{ m E,PW,24h}$: 185 dB	<i>L</i> e,pw,24h: 201 dB			
Otariid Pinnipeds (OW) (Underwater)	$L_{ m pk,flat}$: 232 dB $L_{ m E,ow,24h}$: 203 dB	L E,0W,24h: 219 dB			

^{*} Dual metric acoustic thresholds for impulsive sounds: Use whichever results in the largest isopleth for calculating PTS onset. If a non-impulsive sound has the potential of exceeding the peak sound pressure level thresholds associated with impulsive sounds, these thresholds should also be considered.

Note: Peak sound pressure $(L_{\rm pk})$ has a reference value of 1 μ Pa, and cumulative sound exposure level $(L_{\rm E})$ has a reference value of 1 μ Pa's. The subscript "flat" is being included to indicate peak sound pressure should be flat weighted or unweighted within the generalized hearing range. The subscript associated with cumulative sound exposure level thresholds indicates the designated marine mammal auditory weighting function (LF, MF, and HF cetaceans, and PW and OW pinnipeds) and that the recommended accumulation period is 24 hours. The cumulative sound exposure level thresholds could be exceeded in a multitude of ways (i.e., varying exposure levels and durations, duty cycle). When possible, it is valuable for action proponents to indicate the conditions under which these acoustic thresholds will be exceeded.

Table 10 provides the modeled distances to Level A and Level B exposure thresholds from vibratory and impulsive noise sources used to estimate potential exposure to ESA-listed species (ECO49 2017).

The MMPA defines "harassment" as: any act of pursuit, torment, or annoyance which (i) has the potential to injure a marine mammal or marine mammal stock in the wild [Level A harassment]; or (ii) has the potential to disturb a marine mammal or marine mammal stock in the wild by causing disruption of behavioral patterns, including, but not limited to, migration, breathing, nursing, breeding, feeding, or sheltering [Level B harassment].

While the ESA does not define "harass," NMFS recently issued guidance interpreting the term "harass" under the ESA as to: "create the likelihood of injury to wildlife by annoying it to such an extent as to significantly disrupt normal behavioral patterns which include, but are not limited to, breeding, feeding, or sheltering" (Wieting 2016). For the purposes of this consultation, any

action that amounts to incidental harassment under the MMPA—whether Level A or Level B—constitutes an incidental "take" under the ESA and must be authorized by the ITS (Section 10).

As described below, we anticipate that exposures to listed marine mammals from noise associated with the proposed action may result in disturbance and potential injury. However, no mortalities or permanent impairment to hearing are anticipated.

6.2 Exposure

As discussed in the *Approach to the Assessment* section of this opinion, exposure analyses are designed to identify the ESA-listed resources that are likely to co-occur with the action's effects in space and time, as well as the nature of that co-occurrence. In this step of our analysis, we try to identify the number, age (or life stage), and sex of the individuals that are likely to be exposed to the action's effects and the population(s) or subpopulation(s) those individuals represent.

6.2.1 Exposure to Major Noise Sources

The potential for incidental take is estimated for each species by determining the likelihood that a listed marine mammal would be present within a Level A or Level B Zone of Influence (ZOI) during active pile driving/removal or DTH hammering for each of the construction scenarios shown in Table 3 and described below.

- Scenario 1: Removal of existing piles and installation/removal of temporary piles
- Scenario 2: Installation of 18-inch pipe piles with DTH hammering (dock and dolphin)
- Scenario 3: Installation of 18-inch pipe piles (barge landing)
- Scenario 4: Installation of 30-inch pipe piles with DTH hammering (dolphins)
- **Scenario 5**: Installation of H piles (dock wave barrier)
- Scenario 6: Installation of sheet piles (dock wave barrier and barge landing)

Assumptions

Acoustic Analysis Assumptions

All environmental and sound source parameters used to model sound fields, including references for each data set from which the parameters were derived, and all related assumptions made are provided in Appendix A of the IHA application (Quijano and Austin 2017).

Where uncertainties in operating conditions existed, the models were parameterized to yield realistically conservative noise levels. Several such conservative assumptions were applied to the methods used in this study so that the results would not underestimate potential effects on marine life including:

- All the distances to thresholds (Rmax, R95%) and the noise level contour maps represent the maximum sound levels over all depths because marine mammals swim through a wide depth range;
- The spectral content from pile driving activities of H piles, sheet piles, and cylindrical
 piles (impact driving) either modeled by PDSM or obtained from the literature, was
 extended in frequency to account for energy contributions up to 25 kHz, using
 measurement-based spectral decay factors;

- The location selected for modeling corresponded to the pile in the deepest water, for which coupling of sound into the water column is higher than shallower locations. In addition, the bathymetry used for modeling was also adjusted to represent mean high water;
- Modeling was carried out using the August sound speed profile, which exhibits the most pronounced surface channel over all other months for this geographic area;
- The spectral content for vibratory pile driving of cylindrical piles was obtained from measured levels, which exhibit large energy contributions at high-frequency harmonics.
- Source levels for impact pile driving testing in scenarios 2 and 4 were adjusted to reflect the higher SEL commonly observed in measurements obtained when piles are being driven to near refusal.

The reported radii for 24-hr SEL (Level A) thresholds are based on the assumption that marine mammals remain stationary or at a constant exposure range during the entire 24-hr period, which is an extremely unlikely scenario. These estimated distances for Level A represent an unlikely worst-case scenario.

Exposure Assumptions

- Animals occurring within the Level A and Level B ensonified zones are considered to be in each zone simultaneously, but would only be counted as one Level A take;
- Exposures are based on total number of days that pile driving could occur and that animals might occur in the ensonified action area;
- One day equates to any length of time that piles are driven whether it is a partial day or a 24-hour period;
- All listed marine mammals occurring in the action area are assumed to be incidentally taken:
- An individual animal can only be taken once during a 24-hour period;
- For animals that may occur in small groups, each individual in the group would be considered taken;
- Exposures to sound levels at or above the relevant thresholds equate to take, as defined by the MMPA; and
- Level B take estimates are unmitigated and do not take into account monitoring and mitigation efforts to reduce take as described in Section 2.1.5.

Finally, animals are assumed to be stationary and remain in the area of ensonification. This is unlikely as animals would be expected to move away from the noise source before the exposure would result in a meaningful impact that might affect the individual or populations (ECO49 2017).

Mitigation Measures to Minimize the Likelihood of Exposure to Major Noise Sources

Mitigation measures are described in detail in Section 2.1.5. The following mitigation measures will be required through the MMPA permitting process to reduce the adverse effects of exposure to major noise sources on marine mammals from the proposed construction activities.

1. PSOs are required on pile driving/removal and DTH hammering activities that may result in incidental take through acoustic exposure.

- 2. Establishment of radii associated with received sound level thresholds for Level A shutdown for marine mammals under NMFS's authority, including the use of PSOs to monitor for, and take steps to minimize occurrence of, marine mammals within Level A harm and Level B harassment zones for pile removal, pile driving, and DTH hammering activities that may result in incidental take through acoustic exposure.
- **3.** Use of soft start procedures for impact hammering.
- **4.** Timing restrictions for pile driving/removal and DTH hammering operations to avoid herring aggregations.

Approach to Estimating Exposure to Major Noise Sources

For this analysis we estimated take by considering: 1) acoustic thresholds above which the best available science indicates marine mammals will be behaviorally harassed or incur some degree of permanent hearing impairment; 2) the area or volume of water that will be ensonified above these levels in a day; 3) the density or occurrence of marine mammals within these ensonified areas; and 4) and the number of days of activities.

Distances to Level A and Level B Sound Thresholds

To assess potential underwater noise exposure of marine mammals during construction activities, Quijano and Austin (2017) determined source levels for six different construction scenarios (see Table 3). The source levels are frequency-dependent and suitable for modeling underwater acoustic propagation using JASCO's Marine Operations Noise Model (MONM). The modeling predicted the extent of ensonification and the acoustic footprint from construction activities, taking into account the effects of pile driving equipment, bathymetry, sound speed profile, and seabed geoacoustic parameters. Auditory weighting was applied to the modeled sound fields to estimate received levels relative to hearing sensitivities of five marine mammal hearing groups following NMFS 2016 guidance.

Quijano and Austin (2017) calculated the maximum extent (distance and ensonified areas) of the Level A and Level B exposure zones for each marine mammal functional hearing group. This was calculated for both impact and vibratory pile driving of 18- and 30-in piles and DTH hammering for each of the following six project scenarios. Source levels for sheet pile and H pile installation were obtained from literature, but the available measurements did not cover the full frequency spectrum of interest; data for vibratory installation of sheet and H piles were available to maximum frequencies of 4 kHz and 10 kHz, respectively.

Modeled results are presented in Table 10 with the distances in meters to marine mammal Level A and Level B thresholds for impulsive sources and vibratory sources (Table 10). The distances are based on NMFS PTS/TTS Technical Guidance (NMFS 2016e) for Level A and generic sound exposure thresholds for Level B (70 FR 1871). Acoustic contour maps, which show the directivity and range to various sound level isopleths, are presented in Appendix A of the IHA Application (Quijano and Austin 2017). The reported radii for 24-hr SEL (Level A) thresholds are based on the assumption that marine mammals remain stationary or at a constant exposure range during the entire 24-hr period, which is an extremely unlikely scenario.

Table 10. Modeled distances to Level A and Level B exposure thresholds from vibratory and impulsive noise sources (ECO49 2017).

			Distance to Level A Threshold (m) ¹											
			S1 ² S2		2	S3		S4		S5		S6		
Functional Hearing Group	Level A Vibratory Threshold (dBre 1µPa²·s)	Level A Impulse Threshold (dBre 1µPa²·s)	Vibratory	Impulse	Vibratory	Impulse	Vibratory	Impulse	Vibratory	Impulse	Vibratory	Impulse	Vibratory	Impulse
LFC ³	199	183	10	NA	350	30	10	630	260	80	10	670	40	1,360
MFC ⁴	198	185	10	NA	10	10	10	10	10	10	10	130	10	220
HFC ⁵	173	155	10	NA	510	50	10	770	360	90	10	2,050	10	2,930
PPW^6	201	185	10	NA	80	10	10	200	60	10	10	290	10	770
OPW ⁷	219	203	10	NA	10	10	10	10	10	10	10	10	10	80
	Level B Vibratory Threshold (dBre 1µPa)	Level B Impulse Threshold (dBre 1µPa)	Distance to Level B Threshold (m) ¹											
All marine mammals	120	160	1,800	NA	10,002	710	4,270	550	10,002	1,250	790	430	4,720	790

NOTE: Green shading indicates zones that apply to ESA-listed species.

 $^{^1}R_{95\%}$ as modeled in Appendix A (Quijano and Austin 2017). $^2Scenario\ 1$ does not include impulsive sources.

³Low-frequency cetaceans, humpback whales.

⁴Mid-frequency cetaceans, killer whales.

⁵High-frequency cetaceans, harbor porpoise.

⁶Phocid pinnipeds in water, harbor seals.

⁷Otariid pinnipeds in water, Steller sea lions.

Exposure Estimates

At-sea densities have not been determined for marine mammals in Sitka Sound; therefore, all estimates here are determined by using observational data from biologists, peer-reviewed literature, and information obtained from personal communication with researchers and state and Federal biologists, and from local charter boat operators (ECO49 2017).

Western DPS Steller Sea Lion

Steller sea lion abundance was estimated from NMFS Marine Mammal Lab surveys taken during pupping season, and from observations taken by charter whale-watching and fishing boats that operate from spring throughout the summer season in Sitka Sound (ECO49 2017).

Sea lions focus on seasonal prey resources which occur in the northcentral Sitka Sound area. The center of herring spawning distribution is north of calculated ZOIs for activities at the project site (see Table 10). Herring do not spawn in Symonds Bay and prey resources inside Symonds Bay are limited, especially as compared to the northern coastal areas of Sitka Sound. Therefore, while individual sea lions may occur outside Symonds Bay it is unlikely that they would be attracted to Symonds Bay to forage. During the construction period of May through September, sea lion numbers near Biorka Island are likely to be at their lowest. However, to be precautionary, it is assumed that Steller sea lions may be present outside Symonds Bay every day during construction, and exposure estimates include multiple harassments of the same individuals. Based on previous survey and whale-watching data, group size is anticipated to range from three to five animals per day (ECO49 2017). Using maximum group size, exposure of five animals to noise levels that may exceed Level B thresholds could occur each day of the proposed action (a total of 70 days):

Underwater Level B exposure estimate: 5 animals/day \times 70 days of pile driving activity = 350 potential Steller sea lion exposures to noise levels that may exceed the Level B threshold.

As described in Section 4.3.2, there has been an influx of western DPS animals into Southeast Alaska. For this application, it is assumed that 50 percent of the animals in the action area could be from the western DPS. Therefore, potential Level B takes are apportioned to each DPS (ECO49 2017). Out of a total 350 exposures, 175 animals are anticipated to be from the ESA-listed western DPS (Table 11).

No Level A takes are anticipated for Steller sea lions. The linear distance (from the noise source) to the threshold for a Level A take for sea lions (OPW) is less than 80 m (see Table 10). However, to be precautionary, it is assumed that a Steller sea lion could approach the construction site and become exposed to noise levels that exceed Level A thresholds during impact pile driving and DTH hammering. Level A takes were calculated considering that one group of five animals could be exposed. Assuming 50% of total exposures may be from the western DPS, and rounding up to the nearest whole number, we anticipate up to three Level A takes to western DPS animals may occur (Table 11).

Mexico DPS Humpback Whale

Humpback whales are most abundant in Sitka Sound from late fall through April when they forage on large densities of herring (Liddle et al. 2015b). The seasonal increase in whale abundance corresponds to increases in Pacific herring biomass during pre-spawning, spawning and overwintering periods (Liddle et al. 2015a). Whales feed on large schools of adult, overwintering herring throughout winter, and on pre-spawning and spawning aggregations of herring in spring. Sitka Sound is believed to be a last feeding stop for humpback whales as they migrate to winter breeding and calving waters in Hawaii. During winter months, groups of 30 to 40 humpback whales have been observed by the whale watching industry from the coastline of Sitka Sound (ECO49 2017).

The abundance of humpbacks in Sitka Sound changes by several orders of magnitude from one season to another in response to dense schools of herring in the sound (Liddle et al. 2015a). They are generally present in large numbers from late fall-early winter through mid- to late spring, but are infrequent to uncommon during the mid-summer months when herring are absent. During mid-summer, tour boat operators generally observe four to five whales per day near rocky islets in the middle of Sitka Sound.

Underwater Level B exposure estimate: 5 animals/day \times 70 days of pile driving activity = 350 potential humpback whale exposures to noise levels that may exceed the Level B threshold.

Within the Southeast Alaska and northern British Columbia, 93.9% of humpback whales are anticipated to be from the Hawaii DPS, with the remaining 6.1% from the Mexico DPS (NMFS 2016c, Wade et al. 2016). Therefore, potential Level B takes are apportioned to each DPS (ECO49 2017). Out of a total 350 Level B exposures, 22 animals are anticipated to be from the ESA-listed Mexico DPS (Table 11).

The maximum distance at which a humpback whale may be exposed to noise levels that exceed Level A thresholds is 1.4 km during Scenario 6 (see Table 10). Even though the ensonified area extends outside of the entrance to Symonds Bay, a PSO stationed near the mouth of the bay at Hanus Point would be able to see a humpback whale outside Symonds Bay before it enters the Level A zone and could shut down the noise producing activity to avoid Level A take. In the unlikely event a whale would go undetected and enter the Level A zone, the FAA requested authorization of three takes by Level A harassment for humpback whales. We estimate that all three humpback whales would be from the Hawaii DPS. No Level A takes for Mexico DPS humpback whales are anticipated (Table 11).

Table 11. Amount of proposed incidental harassment (takes) of ESA-listed species in the proposed IHA (82 FR 41229).

Species	Proposed Authorized Level A Takes	Proposed Authorized Level B Takes
Western DPS Steller sea lion (Eumatopias jubatus)	3	175 ¹⁰
Mexico DPS Humpback whale (Megaptera noviaengliae)	0	2211

Note: Take estimates are rounded up to the nearest whole number

In the *Response Analysis* (Section 6.3) we apply the best scientific and commercial data available to describe the species' expected responses to these exposures.

6.2.2 Exposure to Vessel Interactions

Mitigation Measures to Minimize the Likelihood of Exposure to Vessel Noise

As discussed in Section 2.1.5, the following mitigation measures will be required through FAA and PR1's permitting process to avoid or minimize exposure of marine mammals to vessel noise:

- 1. Vessels will not approach within 100 m of marine mammals
- 2. All vessels associated with project construction will avoid the 3,000 ft (914 m) zones surrounding any major rookery or haulout.
- **3.** If a marine mammal comes within 10 meters, operations shall cease and vessels shall reduce speed to the minimum level required to maintain steerage and safe working conditions.

Results of Vessel Noise Exposure

There are two phases of vessel noise and associated disturbance related to the proposed action. The first is vessel noise associated with the construction phase, and the second is vessel noise associated with operation of the Biorka dock.

These acoustic impacts will result from moving sources, and for individual marine mammals that are exposed to noise from transiting vessels, the effects from each exposure will be temporary in duration, on the order of minutes. For species such as humpback whales that prey upon food items that are not tied to a particular location in the way that salmon are seasonally tied to stream channels and stream mouths, effects of transient and temporary noise are expected to result in

¹⁰ The proposed IHA (<u>82 FR 41229</u>) indicated a requested Level A take of 6 Steller sea lion, and a Level B take of 350 Steller sea lion. Of the proposed takes, 50% are anticipated to occur to ESA-listed western DPS animals. The basis for this apportionment is described in Section 4.3.2

¹¹ The proposed IHA (<u>82 FR 41229</u>) indicated a requested Level A take of 3 humpback whales, and a Level B take of 350 humpback whales. Humpback whales in southeast Alaska include individuals from two DPSs. Of the 3 level A exposures, we anticipate zero threatened Mexico DPS animals will be taken, and all 3 will occur from the Hawaii DPS, which is not listed under the ESA. The basis for this apportionment is described in Section 4.3.1.

low levels of exposure and exposure that the animals can likely avoid without foregoing highly valuable foraging opportunities.

Vessel noise associated with this action will be transmitted through water and constitutes a continuous noise source. NMFS anticipates that whenever noise is produced from vessel operations, it may overlap with western DPS Steller sea lions and Mexico DPS humpback whales and that some individuals are likely to be exposed to these continuous noise sources. Broadband source levels for tugs and barges have been measured at 145 to 170 dB re: 1 µPa, and 170 to 180 dB re: 1 µPa for small ships and supply vessels (Richardson et al. 1995). Sound from vessels within this size range would reach the 120 dB threshold distances between 86 m and 233 m (282 and 764 feet) from the source (Richardson et al. 1995). Listed cetaceans and pinnipeds have the potential to overlap with vessel noise associated with the proposed construction activities. We anticipate low level exposure of short-term duration to listed marine mammals from vessel noise, and do not expect significant behavioral reactions. We will discuss potential responses of listed species to vessel noise in Section 6.3.2.

6.3 Response Analysis

As discussed in the *Approach to the Assessment* section of this opinion, response analyses determine how listed species are likely to respond after being exposed to an action's effects on the environment or directly on listed species themselves. Our assessments try to detect the probability of lethal responses, physical damage, physiological responses (particular stress responses), behavioral responses, and social responses that might result in reducing the fitness of listed individuals. Ideally, our response analyses consider and weigh evidence of adverse consequences, beneficial consequences, or the absence of such consequences.

Loud underwater noise can result in physical effects on the marine environment that can affect marine organisms. Possible responses by ESA-listed whales and pinnipeds to the impulsive and continuous sound produced by pile installation and removal, DTH hammering, and vessel noise include:

- Physical Response
 - o Threshold shifts
 - o Non-auditory physiological effects
- Behavioral responses
 - o Auditory interference (masking)
 - Tolerance or Habituation
 - o Change in dive, respiration, or feeding behavior
 - o Change in vocalizations
 - Avoidance or Displacement
 - o Vigilance

This analysis also considers information on the potential effects on prey of ESA-listed species in the action area.

6.3.1 Responses to Major Noise Sources (Pile Driving/Removal and DTH Hammering)

As described in the Sections 6.2.1, Mexico DPS humpback whales and western DPS Steller sea lions are anticipated to occur in the action area and are anticipated to overlap with noise associated with impact and vibratory pile driving/removal and DTH hammering activities. We assume that some individuals are likely to be exposed and respond to these impulsive and continuous noise sources.

Between May 1, 2018 and September 30, 2018, we estimated zero Mexico DPS humpback and three Western DPS Steller sea lions¹² may be exposed at noise levels loud enough, long enough, or at distances close enough to cause Level A harassment (see Section 6.2.1, *Exposure to Major Noise Sources*, Table 11). In addition, 22 Mexico DPS humpback¹³, and 175 western DPS Steller sea lions¹² may be exposed to noise levels sufficient to cause Level B harassment. All level B instances of take are anticipated to occur at received levels \geq 120 dB or 160 dB for continuous and impulsive noise sources respectively.

The effects of sounds from pile driving/removal and DTH hammering might result in one or more of the following: temporary or permanent hearing impairment, non-auditory physical or physiological effects, behavioral disturbance, and masking (Richardson et al. 1995, Nowacek et al. 2007, Southall et al. 2007). The effects of pile driving and DTH hammering on marine mammals are dependent on several factors, including the size, type, and depth of the animal; the depth, intensity, and duration of the pile driving and hammering sound; the depth of the water column; the substrate of the habitat; the distance between the pile and the animal; and the sound propagation properties of the environment. Impacts to marine mammals from pile driving/removal and DTH hammering activities are expected to result primarily from acoustic pathways. As such, the degree of effect is intrinsically related to the received level and duration of the sound exposure, which are in turn influenced by the distance between the animal and the source. The further away from the source, the less intense the exposure should be. The substrate and depth of the habitat affect the sound propagation properties of the environment. Shallow environments are typically more structurally complex, which leads to rapid sound attenuation. In addition, substrates that are soft (e.g., sand) absorb or attenuate the sound more readily than hard substrates (e.g., rock), which may reflect the acoustic wave. Soft porous substrates would also likely require less time to drive the pile, and possibly less forceful equipment, which would ultimately decrease the intensity of the acoustic source.

These instances of exposure assume a uniform distribution of animals and do not account for avoidance. The implementation of mitigation measures to reduce exposure to high levels of pile driving sound, the short duration of pile driving operations, and movement of animals reduce the likelihood that exposure to pile driving would cause a behavioral response that may affect vital

The proposed IHA (82 FR 41229) estimated a total of 6 Level A takes of Steller sea lions, and 350 level B takes. Of the proposed takes, 50% are anticipated to occur to ESA-listed western DPS animals.

¹³ The proposed IHA (82 FR 41229) indicated a requested Level A take of 3 humpback whales, and a Level B take of 350 humpback whales. Humpback whales in southeast Alaska include individuals from two DPSs. Of the proposed takes, 6.1% are anticipated to occur to ESA-listed Mexico DPS animals. The basis for this apportionment is described in Section 4.3.2.

functions (reproduction or survival), or would result in temporary threshold shift (TTS) or permanent threshold shift (PTS).

Cetacean Responses (Mexico DPS Humpback Whale)

As discussed in the *Status of the Species* section, we have no data on baleen whale hearing so we assume that baleen whale vocalizations are partially representative of their hearing sensitivities. While there is no direct data on hearing in low-frequency cetaceans, the applied frequency range is anticipated to be between 7 Hz to 35 kHz (NMFS 2016e)

Humpback whales produce a wide variety of sounds. During the breeding season males sing long, complex songs, with frequencies in the 20-5000 Hz range and intensities as high as 181 dB (Payne 1970, Winn et al. 1970a, Thompson et al. 1986). Source levels average 155 dB and range from 144 to 174 dB (Thompson et al. 1979). Social sounds in breeding areas associated with aggressive behavior in male humpback whales are very different than songs and extend from 50 Hz to 10 kHz (or higher), with most energy in components below 3 kHz (Tyack and Whitehead 1983, Silber 1986). These sounds appear to have an effective range of up to 9 km (Tyack and Whitehead 1983). Humpback whales produce sounds less frequently in their summer feeding areas. Feeding groups produce distinctive sounds ranging from 20 Hz to 2 kHz, with median durations of 0.2-0.8 seconds and source levels of 175-192 dB (Thompson et al. 1986). These sounds are attractive and appear to rally animals to the feeding activity (D'Vincent et al. 1985, Sharpe and Dill 1997).

This information leads us to conclude that humpback whales exposed to sounds produced by pile driving/removal and DTH hammering activities are likely to respond if they are exposed to low-frequency sounds. However, because whales are not likely to communicate at source levels that would damage the tissues of other members of their species, this evidence suggests that received levels of up to 175-192 dB are not likely to damage the tissues of humpback whales (Thompson et al. 1986).

Humpback whale seasonal distribution varies from infrequent (very low in number during summer), to common (very abundant during late fall through spring). Humpback whales are most abundant in Sitka Sound from late fall through April when they forage on large densities of herring (Liddle et al. 2015a). During the period of pile driving/removal and DTH hammering operations (May-Sept) lasting approximately 70 days, humpback whales are anticipated to be in low abundance in the action area when herring are minimal.

Pinniped Responses (Western DPS Steller Sea Lion)

The ability to detect sound and communicate underwater and in-air is important for a variety of Steller sea lion life functions, including reproduction and predator avoidance. NMFS categorizes Steller sea lions in the otariid pinniped functional hearing group with an applied frequency range between 60 and 39 kHz in water (NMFS 2016e).

The breeding season (late May to early July) for Steller sea lions overlaps with the proposed action (May-Sept). Kaiuchali Island is used as a sea lion rookery in spring-summer and as a haulout during the non-breeding seasons (Fritz et al. 2016). Steller sea lions are anticipated to be in very low numbers in Sitka Sound during the summer months when herring are minimal.

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Pile driving/removal and DTH hammering activities would likely impact Mexico DPS humpback whales and western DPS Steller sea lions, although the level of disturbance depends on whether the whales are feeding or migrating, as well as other factors such as the age of the animal, whether it tolerates the sound, etc. In addition to targeted studies in marine mammals indicating that frequency (beyond just differing sensitivities at different frequencies) can affect the likelihood of auditory impairment incurred, there is increasing evidence that contextual factors other than received sound level, including activity states of exposed animals, the nature and newness of the sound, and the relative spatial positions of sound and receiver, can strongly affect the probability of behavioral response (Ellison et al. 2012).

Physical Responses

Systemic stressors usually elicit direct physical or physiological responses and, therefore do not require high-level cognitive processing of sensory information (Herman and Cullinan 1997, Anisman and Merali 1999, de Kloet et al. 2005, Wright et al. 2007). These physical responses are not influenced by the animal's assessment of whether a potential stressor poses a threat or risk.

Threshold Shifts

Exposure of marine mammals to very loud noise can result in physical effects, such as changes to sensory hairs in the auditory system, which may temporarily or permanently impair hearing. TTS is a temporary hearing change and its severity is dependent upon the duration, frequency, sound pressure, and rise time of a sound (Finneran and Schlundt 2013). TTSs can last minutes to days. Full recovery is expected and this condition is not considered a physical injury. At higher received levels, or in frequency ranges where animals are more sensitive, PTS can occur. When PTS occurs, auditory sensitivity is unrecoverable (i.e., permanent hearing loss). Both TTS and PTS can result from a single pulse or from accumulated effects of multiple pulses from an impulsive sound source (i.e., impact pile driving) or from accumulated effects of non-pulsed sound from a continuous sound source (i.e., vibratory or DTH hammering). In the case of exposure to multiple pulses, each pulse need not be as loud as a single pulse to have the same accumulated effect.

Few data are available to define the hearing range, frequency sensitivities, or sound levels necessary to induce TTS or PTS in whales and pinnipeds. The best available information for whales and pinnipeds comes from captive studies of toothed whales and California sea lions, studies of terrestrial mammal hearing, and extensive modeling (Finneran et al. 2000, Schlundt et al. 2000, Finneran et al. 2002, Finneran et al. 2003, Nachtigall et al. 2003, Nachtigall et al. 2004, Finneran et al. 2005, Finneran et al. 2007, Lucke et al. 2009, Mooney et al. 2009a, Mooney et al. 2009b, Finneran et al. 2010a, Finneran et al. 2010b, Finneran and Schlundt 2010, Popov et al. 2011a, Popov et al. 2011b, Kastelein et al. 2012a, Kastelein et al. 2012b). Finneran et al. (2003) exposed two California sea lions to single underwater pulses up to 183 dB re 1 μ Pa_{p-p} and found no measurable TTS following exposure. Southall et al. (2007) estimated TTS will occur in pinnipeds exposed to a single pulse of sound at 212 dB re 1 μ Pa_{0-p} and PTS will occur at 218 dB re 1 μ Pa_{0-p}. Based on this information, NMFS established the following Level A impulsive sound thresholds for low-frequency cetaceans and otarrid pinnipeds in the water as 183 dB re1 μ Pa²·s, and 203 dB re1 μ Pa²·s respectively (NMFS 2016e).

Both duration and pressure level of a sound are factors in inducement of threshold shift. Exposure to non-pulsed sound (i.e., vibratory or DTH hammering) may induce more threshold shift than exposure to a pulsed sound with the same energy; however, this is dependent on the duty cycle of the pulsed source (because some recovery may occur between exposures) (Kryter et al. 1966, Ward 1997). For example, the impairment caused by exposure to one a high SPL pulse may equal the exposure of a lower SPL continuous sound. The low level continuous sound may also cause more impairment than a series of intermittent lower SPL sounds (Ward 1997). TTS was reported in toothed whales after exposure to relatively short, continuous sounds (ranging from 1 to 64 sec) at relatively high sound pressure levels ranging from 185 to 201 dB re 1 μPa_{rms} (Ridgway et al. 1997, Schlundt et al. 2000, Finneran et al. 2005, Finneran et al. 2007); however, toothed whales experienced TTS at lower sound pressure levels (160 to 179 dB re 1 μPa_{rms}) when exposed to continuous sounds of relatively long duration ranging from 30 to 54 min (Nachtigall et al. 2003, Nachtigall et al. 2004). Kastak et al. (2005) indicated pinnipeds exposed to continuous sounds in water experienced the onset of TTS from 152 to 174 dB re 1 μPa_{rms}. ¹⁴ Southall et al. (2007) estimated PTS will occur in pinnipeds exposed to continuous sound pressure levels of 218 dB re: 1 µPa_{0-p}.

Based on this information NMFS established the following Level A continuous sound thresholds for low-frequency cetaceans and otariid pinnipeds in the water as 199 dB re1 μ Pa²·s, and 219 dB re1 μ Pa²·s respectively (NMFS 2016e).

To experience TTS from a continuous source, a humpback whale will have to remain in the respective 10 m, 350 m, 260 m, or 40 m radius ZOIs for an extended period of time and will need to remain in the ZOI even longer to experience PTS. For Steller sea lions continuous source levels never reach TTS or PTS onset (see Table 10). The reported radii for 24-hr SEL (Level A) thresholds are based on the assumption that marine mammals remain stationary or at a constant exposure range during the entire 24-hr period, which is an extremely unlikely scenario, though it is possible they may remain in the area if highly motivated by the presence of a food source. In this instance, it is possible that a whale could experience TTS if it chooses to remain in the ensonified area for an extended period. Though the exact time a whale will need to remain in the ensonified area to experience threshold shift is not known, based on the findings from Nachtigall et al. (2003) and Nachtigall et al. (2004), we estimate a whale will need to remain in the ensonified zone for tens of minutes to experience low-level TTS and likely several to tens of hours to experience PTS, if at all. Considering the applicant has agreed to shutdown if marine mammals approach or occur within the Level A zones, TTS and PTS is unlikely to occur.

Non-auditory Physiological Effects

Non-auditory physiological effects or injuries that theoretically might occur in marine mammals exposed to strong underwater sound include stress, neurological effects, bubble formation, resonance effects, and other types of organ or tissue damage (Cox et al. 2006, Southall et al. 2007). Studies examining such effects are limited. In general, little is known about the potential for pile driving/removal or DTH hammering to cause auditory impairment or other physical effects in marine mammals. Available data suggest that such effects, if they occur at all, would presumably be limited to short distances from the sound source and to activities that extend over a prolonged period. The available data do not allow identification of a specific exposure level above which

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 $^{^{14}}$ Values originally reported as sound exposure level of 183 to 206 dB re 1 μ Pa 2 -s.

non-auditory effects can be expected (Southall et al. 2007) or any meaningful quantitative predictions of the numbers (if any) of marine mammals that might be affected in those ways. Marine mammals that show behavioral avoidance of pile driving, including some cetaceans and some pinnipeds, are especially unlikely to incur auditory impairment or non-auditory physical effects (82 FR 41229).

An animal's perception of a threat may be sufficient to trigger stress responses consisting of some combination of behavioral responses, autonomic nervous system responses, neuroendocrine responses, or immune responses (Moberg 2000). In many cases, an animal's first and sometimes most economical (in terms of energetic costs) response is behavioral avoidance of the potential stressor. Autonomic nervous system responses to stress typically involve changes in heart rate, blood pressure, and gastrointestinal activity. These responses have a relatively short duration and may or may not have a significant long-term effect on an animal's fitness.

Neuroendocrine stress responses often involve the hypothalamus-pituitary-adrenal system. Virtually all neuroendocrine functions that are affected by stress – including immune competence, reproduction, metabolism, and behavior – are regulated by pituitary hormones. Stress-induced changes in the secretion of pituitary hormones have been implicated in failed reproduction, altered metabolism, reduced immune competence, and behavioral disturbance (Blecha 2000). Increases in the circulation of glucocorticoids are also equated with stress (Romano et al. 2004).

The primary distinction between stress (which is adaptive and does not normally place an animal at risk) and "distress" is the cost of the response. During a stress response, an animal uses glycogen stores that can be quickly replenished once the stress is alleviated. In such circumstances, the cost of the stress response would not pose serious fitness consequences. However, when an animal does not have sufficient energy reserves to satisfy the energetic costs of a stress response, energy resources must be diverted from other functions. This state of distress will last until the animal replenishes its energetic reserves sufficient to restore normal function.

Relationships between these physiological mechanisms, animal behavior, and the costs of stress responses are well-studied through controlled experiments and for both laboratory and free-ranging animals (Jessop et al. 2003, Lankford et al. 2005, Crespi et al. 2013). Stress responses due to exposure to anthropogenic sounds or other stressors and their effects on marine mammals have also been reviewed (Fair and Becker 2000, Romano et al. 2002) and, more rarely, studied in wild populations (Romano et al. 2002). For example, Rolland et al. (2012) found that noise reduction from reduced ship traffic in the Bay of Fundy was associated with decreased stress in North Atlantic right whales. During the time following September 11, 2001, shipping traffic and associated ocean noise decreased along the northeastern U.S. This decrease in ocean noise was associated with a significant decline in fecal stress hormones in North Atlantic right whales, suggesting that chronic exposure to increased noise levels, although not acutely injurious, can produce stress (Rolland et al. 2012). These levels returned to their previous level within 24 hrs after the resumption of shipping traffic. Exposure to loud noise can also adversely affect reproductive and metabolic physiology (Kight and Swaddle 2011). In a variety of factors, including behavioral and physiological responses, females appear to be more sensitive or respond more strongly than males (Kight and Swaddle 2011).

These and other studies lead to a reasonable expectation that some marine mammals will experience physiological stress responses upon exposure to acoustic stressors and that it is possible that some of these would be classified as "distress." In addition, any animal experiencing TTS would likely also experience stress responses (NRC 2003).

As discussed throughout the *Response Analysis* of this opinion, we expect individuals may experience TTS (but are not likely to experience PTS), may experience masking, and may exhibit behavioral responses from project activities. Therefore, we expect ESA-listed whales and pinnipeds may experience stress responses. If whales and pinnipeds are not displaced and remain in a stressful environment (i.e. within the ZOI pile driving activities), we expect the stress response will dissipate shortly after the cessation of pile driving. Similarly, if whales or pinnipeds are exposed to sounds from the DTH hammer, we expect a stress response will accompany a brief startle response. However, in any of the above scenarios, we do not expect significant or long-term harm to individuals from a stress response.

Behavioral Responses

Processive stressors require high-level cognitive processing of sensory information (Herman and Cullinan 1997, Anisman and Merali 1999, de Kloet et al. 2005, Wright et al. 2007). Behavioral responses are influences by an animal's assessment of whether a potential stressor poses a threat or risk. Behavioral responses may include: changing durations of surfacing and dives, number of blows per surfacing, or moving direction and/or speed; reduced/increased vocal activities; changing/cessation of certain behavioral activities (such as socializing or feeding); visible startle response or aggressive behavior (such as tail/fluke slapping or jaw clapping); avoidance of areas where sound sources are located; and/or flight responses (*e.g.*, pinnipeds flushing into water from haul-outs or rookeries).

The biological significance of many of these behavioral disturbances is difficult to predict, especially if the detected disturbances appear minor. However, the consequences of behavioral modification could be expected to be biologically significant if the change affects growth, survival, or reproduction. Significant behavioral modifications that could potentially lead to effects on growth, survival, or reproduction include:

- Drastic changes in diving/surfacing patterns (such as those thought to cause beaked whale stranding due to exposure to military mid-frequency tactical sonar);
- Longer-term habitat abandonment due to loss of desirable acoustic environment; and
- Longer-term cessation of feeding or social interaction.

The onset of behavioral disturbance from anthropogenic sound depends on both external factors (characteristics of sound sources and their paths) and the specific characteristics of the receiving animals (hearing, motivation, experience, demography) and is difficult to predict (Southall et al. 2007). Below we describe some of the anticipated behavioral responses to major noise sources associated with the proposed action.

Tolerance, Habituation, and Sensitization

While numerous studies have shown that underwater sounds from industry activities are often readily detectable by marine mammals in the water at distances of many kilometers, few studies have attempted to address habituation, sensitization, or tolerance (Nowacek et al. 2007). Tolerance is defined as 'the intensity of disturbance that an individual tolerates without responding in a defined way' (Nisbet 2000). Tolerance levels can be measured instantaneously and are, therefore, more readily demonstrated than the longer-term processes of habituation or sensitization. In fact, habituation and sensitization are identified, and distinguished from each other, by the direction of change indicated by repeated measures of tolerance taken over time. Thus, over the course of a habituation process, individual tolerance levels will increase, whereas tolerance levels will conversely decrease as individuals become sensitized to specific stimuli (Bejder et al. 2009).

Despite industry activities occurring at distances of only a few kilometers away, oftentimes marine mammals show no apparent response or tolerance to industry activities of various types (Miller et al. 2005, Bain and Williams 2006). This is often true even in cases when the sounds must be readily audible to the animals based on measured received levels and the hearing sensitivity of that mammal group. Weir (2008) observed marine mammal responses to seismic pulses from a 24 airgun array firing a total volume of either 5,085 in³ or 3,147 in³ in Angolan waters between August 2004 and May 2005. Weir recorded a total of 207 sightings of humpback whales (n = 66), sperm whales (n = 124), and Atlantic spotted dolphins (n = 17) and reported that there were no significant differences in encounter rates (sightings/hr) for humpback and sperm whales according to the airgun array's operational status (i.e., active versus silent). Based on the available information on pinnipeds in water exposed to multiple noise pulses, exposures in the ~150-180 dB re 1µ Pa range (rms values over the pulse duration) generally have limited potential to induce avoidance behavior in pinnipeds (Southall et al. 2007). This information indicates marine mammal tolerance of underwater sounds, and we anticipate that some humpback whales and Steller sea lions exposed to low frequency underwater sounds from impulsive construction activities may tolerate pile driving noise and show no apparent response. More information is needed in order to determine if the learned processes of habituation or sensitization are occurring over time as animals experience repeated exposures.

Masking

Masking occurs when anthropogenic sounds and marine mammal signals overlap at both spectral and temporal scales. It is important to distinguish TTS and PTS, which persist after the sound exposure, from masking, which occurs during the sound exposure. Because masking (without resulting in threshold shift) is not associated with abnormal physiological function, it is not considered a physiological effect, but rather a potential behavioral effect.

For the pile driving/removal sound generated from the proposed construction activities, sound will consist of low frequency impulsive and continuous noise depending on if they are using an impact or vibratory hammer, or DTH hammer. Lower frequency anthropogenic sounds are more likely to affect detection of communication calls and other potentially important natural sounds such as surf and prey noise. This could affect communication signals used by low frequency mysticetes when they occur near the noise band and thus reduce the communication space of animals (Clark et al. 2009b) and cause increased stress levels (Foote et al. 2004, Holt et al.

2009). However, marine mammals are thought to be able to compensate for masking by adjusting their acoustic behavior by shifting call frequencies, and/or increasing call volume and vocalization rates. For example, blue whales are found to increase call rates when exposed to seismic survey noise in the St. Lawrence Estuary (Di Lorio and Clark. 2010). In addition, the sound localization abilities of marine mammals suggest that, if signal and noise come from different directions, masking would not be as severe as the usual types of masking studies might suggest (Richardson et al. 1995).

Noise from pile driving/removal and DTH hammering is relatively short-term. It is possible that pile driving/removal and DTH hammering noise resulting from this proposed action may mask acoustic signals important to WDPS Steller sea lions and Mexico DPS humpback whales, but the short-term duration (up to 70 days) limited affected area, and pauses between operations would limits the impacts from masking. Any masking event that could possibly rise to Level B harassment under the MMPA would occur concurrently within the zones of behavioral harassment already estimated for vibratory pile driving and DTH hammering, and which have already been taken into account in the exposure analysis.

Changes in Vocalization

Marine mammals vocalize for different purposes and across multiple modes, such as whistling, echolocation click production, calling, and singing. Changes in vocalization behavior in response to anthropogenic noise can occur for any of these modes and may result from a need to compete with an increase in background noise or may reflect increased vigilance or a startle response.

In addition to these behavioral responses, whales alter their vocal communications when exposed to anthropogenic sounds. Communication is an important component of the daily activity of animals and ultimately contributes to their survival and reproductive success. Animals communicate to find food (Marler et al. 1986, Elowson et al. 1991), acquire mates (Ryan 1985), assess other members of their species (Parker 1974, Owings et al. 2002), evade predators (Greigsmith 1980), and defend resources (Zuberbuhler et al. 1997). Human activities that impair an animal's ability to communicate effectively might have significant effects on the survival and reproductive performance of animals experiencing the impairment.

At the same time, most animals that vocalize have evolved with an ability to make adjustments to their vocalizations to increase the signal-to-noise ratio, active space, and recognizability of their vocalizations in the face of temporary changes in background noise (Cody and Brown 1969, Brumm 2004, Patricelli and Blickley 2006). A few studies have demonstrated that marine mammals make the same kind of vocal adjustments in the face of high levels of background noise. For example, two studies reported that some mysticete whales stopped vocalizing – that is, adjust the temporal delivery of their vocalizations – when exposed to active sonar (Miller et al. 2000, Melcon et al. 2012). Melcón et al. (2012) reported that during 110 of the 395 d-calls (associated with foraging behavior) they recorded during mid-frequency active sonar transmissions, blue whales stopped vocalizing at received levels ranging from 85 to 145 dB, presumably in response to the sonar transmissions. These d-calls are believed to attract other individuals to feeding grounds or maintain cohesion within foraging groups (Oleson et al. 2007). It should also be noted that mid-frequency sonar is not in the frequency range of most baleen whale calls, and a response by blue whales to mid-frequency sonar suggests that they have the ability to perceive and respond to these sounds (Erbe 2002a, Southall et al. 2007, Melcon et al. 2012).

Humpback whales have been observed to increase the length of their songs in the presence of potentially masking signals (Miller et al. 2000, Fristrup et al. 2003).

Pirotta et al. (2014) showed that the probability of recording a harbor porpoise "buzz" (interclick interval associated with attempted prey captures or social communication) declined by 15% in the ensonified area of a 2D seismic operation. The probability of occurrence of buzzes increased significantly with distance from the seismic source. This suggests that the likelihood of buzzing was dependent upon received noise intensity. Observed changes in buzzing occurrence could reflect disruption of either foraging or social activities. These effects may result from prey reactions to noise, leading to reduced porpoise foraging rates. Alternatively, foraging effort may change if porpoises adjust time budgets or diving behavior to avoid noise (Pirotta et al. 2014).

The effect of seismic airgun pulses on bowhead whale calling behavior has been extensively studied in the Beaufort Sea and is similar to the patterns reported in other whales. During the autumn in 2007 and 2008, calling rates decreased significantly in the presence (<30 km [<18.6 mi]) of airgun pulses (Blackwell et al. 2010).

In birds, song diversity is an important index for population viability, and is influenced by anthropogenic noise (Laiolo et al. 2008, Slabbekorn and Ripmeester 2008), and in bowhead whales song diversity and complexity may serve as a barometer of the impact of encroaching Arctic oil and gas development (Johnson et al. 2014). The same may be true for song diversity and complexity in humpback whales.

Responses While Feeding

The absence of changes in the behavior of foraging humpback whales or Steller sea lions should not be interpreted to mean that the marine mammals were not affected by the noise. Animals that are faced with human disturbance must evaluate the costs and benefits of relocating to alternative locations; those decisions would be influenced by the availability of alternative locations, the distance to the alternative locations, the quality of the resources at the alternative locations, the conditions of the animals faced with the decision, and their ability to cope with or "escape" the disturbance (Lima and Dill 1990, Gill and Sutherland 2001, Frid and Dill. 2002, Beale and Monaghan 2004a, b, Bejder et al. 2006, Bejder et al. 2009). Specifically, animals delay their decision to flee from predatory stimuli they detect until they decide that the benefits of abandoning a location are greater than the costs of remaining at the location or, conversely, until the costs of remaining at a location are greater than the benefits of fleeing (Ydenberg and Dills 1986). Ydenberg and Dill (1986) and Blumstein (2003) presented an economic model that recognized that animals will almost always choose to flee a site over some short distance to a predator; at a greater distance, animals will make an economic decision that weighs the costs and benefits of fleeing or remaining; and at an even greater distance, animals will almost always choose not to flee. For example, in a review of observations of the behavioral responses of 122 minke whales, 2,259 fin whales, 833 right whales, and 603 humpback whales to various sources of human disturbance, Watkins (1986) reported that fin, humpback, minke, and North Atlantic right whales tolerated sounds that occurred at relatively low received levels, had most of their energy at frequencies below or above the hearing capacities of these species, or were from distant human activities and received levels were below ambient levels. Most of the negative reactions that were observed occurred within 100 m of a sound source or when sudden increases in received sound levels were judged to be in excess of 12 dB, relative to previous ambient sounds.

As a result of using this kind of economic model to consider whales' behavioral decisions, we would expect whales to continue foraging in the face of moderate levels of disturbance. For example, humpback whales, which only feed during part of the year and must satisfy their annual energetic needs during the foraging season, may continue foraging in the face of disturbance. Similarly, a humpback cow accompanied by her calf is less likely to flee or abandon an area at the cost of her calf's survival. By extension, we assume that animals that choose to continue their pre-disturbance behavior would have to cope with the costs of doing so, which will usually involve physiological stress responses and the energetic costs of stress physiology (Frid and Dill 2002, MMS 2008).

Responses While Migrating and Resting

Migrating whales respond more strongly to noise than do feeding whales. Avoidance responses of migrating humpback whales to impulsive airgun noise appear consistent with bowhead and gray whale avoidance at received levels between 150-180 dB (Richardson et al. 1995). Migrating humpbacks showed localized avoidance of operating airguns in the range of received levels 157-164 dB. In addition, humpback whales seemed more sensitive to seismic airgun noise while exhibiting resting behavior (McCauley et al. 2000). For resting humpback pods that contained cow-calf pairs, the mean airgun noise level for avoidance was 140 dB re 1 μ Pa rms, and a startle response was observed at 112 dB re 1 μ Pa rms (McCauley et al. 2000). When calves are small, comparatively weak and possibly vulnerable to predation and exhaustion, the potential continual dislocation of these animals in a confined area would interrupt this resting and feeding stage, with potentially more serious consequences than any localized avoidance response to an operating seismic vessel as seen during their migratory swimming behavior (McCauley et al. 2000). For comparison with the proposed action, impact pile driving (also an impulsive source) it anticipated to reach the 160 dB re 1 μ Pa rms threshold between 430-1,250 m from the source.

Approach Responses

In 9 of the 16 trials (McCauley et al. 2000), mostly single, large mature humpbacks approached the operating airgun within 100-400m to investigate before swimming off. These whales would have received maximum air gun signals at 100m of 179 dB re 1 µPa rms (or 195 dB re 1 µPa peak—peak). This level is equivalent to the high peak to peak 192 dB re 1µPa humpback whale sounds recorded in Alaska (Thompson et al. 1986). The underwater signals produced by humpback whale breaching were audibly similar to air gun signals. McCauley et al. (2000) speculate that given the similarities between airgun and breaching signals, male humpback whales may identify airgun signals as a "competitor." Humpback whales on feeding grounds did not alter short-term behavior or distribution in response to explosions with received levels of about 150dB re 1µ Pa/Hz at 350Hz (Lien et al. 1993, Todd et al. 1996). However, at least two individuals were probably killed by the high-intensity, impulse blasts and had extensive mechanical injuries in their ears (Ketten et al. 1993, Todd et al. 1996). Frankel and Clark (1998) showed that breeding humpbacks showed only a slight statistical reaction to playbacks of 60 - 90 Hz sounds with a received level of up to 190 dB re 1 µPa rms. Although these studies demonstrated that humpback whales may exhibit short-term behavioral reactions to playbacks of industrial noise, the long-term effects of these disturbances on the individuals exposed to them are not known.

Avoidance

Avoidance is the displacement of an individual from an area or migration path as a result of the presence of a sound or other stressors, and is one of the most obvious manifestations of disturbance in marine mammals (Richardson et al. 1995).

Studies of bowhead, gray, and humpback whales have determined that received levels of pulses in the 160-170 dB re 1 μ Pa rms range seem to cause obvious avoidance behavior in a substantial fraction of the animals exposed.

Marine mammal sighting data during the Apache seismic surveys in Cook Inlet reported the most common behavior of harbor seals during non-seismic periods was "look/sink" followed by "travel," whereas during periods of active seismic shooting, "travel" was more common than "look/sink" (Lomac-MacNair 2014). Received levels exceeding 190 dB re 1μ Pa are likely to elicit avoidance responses, at least in some pinnipeds (Harris et al. 2001, Blackwell et al. 2004, Miller et al. 2005).

Avoidance is one of many behavioral responses whales may exhibit when exposed to pile driving/removal and DTH hammering noise. Other behavioral responses include evasive behavior to escape exposure or continued exposure to a sound that is painful, noxious, or that they perceive as threatening, which we would assume would be accompanied by acute stress physiology; increased vigilance of an acoustic stimulus, which would alter their time budget (that is, during the time they are vigilant, they are not engaged in other behavior); and continued predisturbance behavior with the physiological consequences of continued exposure.

Responses of Prey Resources

As described in the *Status of Listed Species*, in Southeast Alaska, marine mammal distributions and seasonal increases in their abundance are strongly influenced by seasonal pre-spawning and spawning aggregations of forage fish, particularly Pacific herring (*Clupea pallasii*), eulachon (*Thaleichthys pacificus*) and Pacific salmon (*Onchorynchus spp.*) (Marston et al. 2002, Sigler et al. 2004, Womble et al. 2005). All five species of salmon are found in Sitka Sound and are preyed upon by Steller sea lions. However, there are no salmon spawning streams in the vicinity of the project or presence of eulachon or herring during the construction time period that would tend to aggregate foraging marine mammals.

Herring are the keystone species in Southeast Alaska, especially Sitka Sound, serving as a vital link between lower trophic levels, including crustaceans and small fish, and higher trophic levels. Foraging studies of Steller sea lions suggest that during their non-breeding season, they forage on seasonally densely aggregated prey (Sinclair and Zeppelin 2002). In southeast Alaska, Pacific herring typically spawn from March to May and attract large numbers of predators (Marston et al. 2002, Womble et al. 2009). The relationship between humpback whales and Steller sea lions and these ephemeral fish runs is so strong in Sitka Sound that the seasonal abundance and distribution of marine mammals reflects the distribution of pre-spawning and spawning herring, and overwintering aggregations of adult herring in Sitka Sound.

Of all known Steller sea lion prey species, only Chinook and coho salmon have been studied for effects of exposure to pile driving noise (Halvorsen et al. 2012). These studies defined very high

noise level exposures (SEL $_{cum}\,$ of 210 dB re $1\mu Pa^2$ s) as threshold for onset of injury, and supported the hypothesis that one or two mild injuries resulting from pile driving exposure at these or higher levels are unlikely to affect the survival of the exposed animals, at least in a laboratory environment. Hart Crowser Inc. et al. (2009) studied the effects on juvenile coho salmon from pile driving of sheet piles at the Port of Anchorage in Knik Arm of Cook Inlet. The fish were exposed in-situ (in that location) to noise from vibratory or impact pile driving at distances ranging from less than 1 meter to over 30 meters. The results of this studied showed no mortality of any of the test fish within 48 hours of exposure to the pile driving activities, and for the necropsied fish, no effects or injuries were observed as a result of the noise exposure (NMFS 2016d). Noise generated from pile driving can reduce the fitness and survival of fish in areas used by foraging marine mammals; however, given the small area of pile driving within the action area relative to known feeding areas in Steller sea lions, and the fact that any physical changes to this habitat would not be likely to reduce the localized availability of fish (Fay and Popper 2012), it is unlikely that Western DPS Steller sea lion' prey would be affected. In general, we expect fish will be capable of moving away from project activities if they experience discomfort. We expect the area in which stress, injury, TTS, or changes in balance, of prey species may occur (if at all) will be limited to a few meters directly around the pile driving and DTH hammering operations. We consider potential adverse impacts to prey resources from piledriving and DTH hammering in the action area to be unlikely.

Studies on euphausiids and copepods, which are some of the more abundant and biologically important groups of zookplankton, have documented the use of hearing receptors to maintain schooling structures (Wiese 1996) and detection of predators (Chu et al. 1996) respectively, and therefore have some sensitivity to sound; however any effects of pile driving/removal and DTH hammering on zooplankton would be expected to be restricted to the area within a few feet or meters of the project and would likely be sub-lethal.

No appreciable adverse impact on zooplankton populations will occur due in part to large reproductive capacities and naturally high levels of predation and mortality of these populations. Any mortality or impacts on zooplankton as a result of construction operations is immaterial as compared to the naturally-occurring reproductive and mortality rates of these species. This is consistent with previous conclusions that crustaceans (such as zooplankton) are not particularly sensitive to sound produced by even louder impulsive sounds such as seismic operations (Wiese 1996).

Finally, the proposed action is outside critical herring spring spawning and over-wintering time periods, and the action area is approximately 24 km to the southwest of the nearest spawning sites near Sitka (ECO49 2017).

6.3.2 Responses to Vessel Noise

As described in the *Exposure to Vessel Interactions* Section 6.2.2, Mexico DPS humpback whales and western DPS Steller sea lions are all anticipated to occur in the action area and are anticipated to overlap with noise associated with vessel transit. We assume that some individuals are likely to be exposed and respond to these continuous noise sources.

Reactions of marine mammals to vessels often include changes in general activity (e.g. from resting or feeding to active avoidance), changes in surfacing-respiration-dive cycles, and changes in speed and direction of movement (NMFS 2013b). Past experiences of the animals with vessels are important in determining the degree and type of response elicited from an animal-vessel encounter. Whale reactions to slow-moving vessels are less dramatic than their reactions to faster and/or erratic vessel movements. Some species have been noted to tolerate slow-moving vessels within several hundred meters, especially when the vessel is not directed toward the animal and when there are no sudden changes in direction or engine speed (Wartzok et al. 1989, Richardson et al. 1995, Heide-Jorgensen et al. 2003).

Humpback whale reactions to approaching boats are variable, ranging from approach to avoidance (Payne 1978, Salden 1993). On rare occasions humpbacks "charge" towards a boat and "scream" underwater, apparently as a threat (Payne 1978). Baker et al. (1983) reported that humpbacks in Hawaii responded to vessels at distances of 2 to 4 km. Bauer and Herman (1986) concluded that reactions to vessels are probably stressful to humpbacks, but that the biological significance of that stress is unknown. Humpbacks seem less likely to react to vessels when actively feeding than when resting or engaged in other activities (Krieger and Wing 1984). Mothers with newborn calves seem most sensitive to vessel disturbance (Clapham and Mattila 1993). Marine mammals that have been disturbed by anthropogenic noise and vessel approaches are commonly reported to shift from resting behavioral states to active behavioral states, which would imply that the incur an energy cost. Morete et al. (2007) reported that undisturbed humpback whale cows that were accompanied by their calves were frequently observed resting while their calves circled them (milling) and rolling interspersed with dives. When vessels approached, the amount of time cows and calves spent resting and milling respectively declined significantly. There is the potential for interactions between vessels and cow calf pairs in Southeast Alaska.

In general, baleen whales react strongly and rather consistently to approaching vessels of a wide variety of types and sizes. Whales are anticipated to interrupt their normal behavior and swim rapidly away if approached by a vessel. Surfacing, respiration, and diving cycles can be affected. The flight response often subsides by the time the vessel has moved a few kilometers away. After single disturbance incidents, at least some whales are expected to return to their original locations. Vessels moving slowly and in directions not toward the whales usually do not elicit such strong reactions (Richardson and Malme 1993).

Few authors have specifically described the responses of pinnipeds to boats, and most of the available information on reactions to boats concerns pinnipeds hauled out on land or ice. However, the mere presence and movements of ships in the vicinity of seals and sea lions can cause disturbance to their normal behaviors (Calkins and Pitcher 1982, Kucey 2005, Jansen et al. 2006), Disturbances from vessels may motivate seals and sea lions to leave haulout locations and enter the water (Richardson 1998, Kucey 2005). The possible impact of vessel disturbance on Steller sea lions has not been well studied, yet the response by sea lions to disturbance will likely depend on the season and life stage in the reproductive cycle (NMFS 2008a).

The action area does not include Steller sea lion critical habitat, and the FAA required all vessels associated with project construction will avoid the 3,000 ft (914 m) designated aquatic zones surrounding any major rookery or haulout as they transit to and from the project site.

Out of the limited number of vessels associated with the proposed action, transiting speeds are anticipated to be 10 knots or less, and vessels will primarily be anchored at the construction site unless deploying people or supplies.

We anticipate that noise associated with transiting vessels would drop to 120 dB within 233 meters (or less) of most vessels associated with the proposed action (Richardson et al. 1995). Considering that regulations restrict approaching humpback whales within 100 yards, and FAA is applying an exclusion zone for vessel transit out to 100 m (~109 yards) around any marine mammal, a Steller sea lion or humpback whale that perceived the vessel noise is likely to ignore such a signal and devote its attentional resources to stimuli in its local environment. If animals do respond, they may exhibit slight deflection from the noise source, engage in low-level avoidance behavior, short-term vigilance behavior, or short-term masking behavior, but these behaviors are not likely to result in adverse consequences for the animals. The nature and duration of response is not anticipated to be a significant disruption of important behavioral patterns such as feeding or resting. During the operational period of the action (May-Sept), the action area is not considered high quality habitat for western DPS Steller sea lions, and while it does overlap with humpback whale feeding BIAs, it is outside spring spawning sites for herring. Temporary avoidance of the action area is not likely to adversely affect these species. Therefore, the impact of vessel transit on Mexico DPS humpback whales and western DPS Steller sea lion is not anticipated to reach the level of harassment under the ESA, and is considered insignificant.

6.3.3 Response Summary

No Level A take of Mexico DPS humpback whales is anticipated. The maximum distance at which a humpback whale may be exposed to noise levels that exceed Level A thresholds is 1.4 km during Scenario 6 (see Table 10). Even though the ensonified area extends outside of the entrance to Symonds Bay, a PSO stationed near the mouth of the bay at Hanus Point would be able to see a humpback whale outside Symonds Bay before it enters the Level A zone and could shut down the noise producing activity to avoid Level A take. In the unlikely event a whale would go undetected and enter the Level A zone, the FAA has requested authorization of three takes by Level A harassment for humpback whales. However, based on the probability of occurrence, all three humpback whales are likely to be from the Hawaii DPS, which is not listed.

We anticipate up to 3 western DPS Steller sea lions may be exposed to Level A harassment from impulsive noise at distances from 10 to 80 meters depending on the Scenario (see Table 10).

It is anticipated that for the major noise sources associated with the proposed action (impact pile driving, vibratory pile removal and driving, and DTH hammering), the distances to the Level B isopleth (120 dB for continuous noise sources, and 160 dB for impulsive noise sources) range from 430m-10 km depending on the source and threshold of concern (Quijano and Austin 2017).

Based on this information, we would not anticipate humpback whales or Steller sea lions to devote attention to a noise stimulus beyond the 120 dB isopleth (for continuous noise sources), which may be more than 10 kilometers from the source (Quijano and Austin 2017), and beyond the 160 dB isopleth (for impulsive noise sources) which may reach more than 430 meters. At these distances, a marine mammal that perceives a signal is likely to ignore such a signal and devote its attention to stimuli in its local environment (that is, they would filter the sound out as background noise or ignore it) (Miller et al. 1999, Richardson 1999). Because of their distance

from the noise source, we would also not anticipate humpback whales or Steller sea lions would change their behavior or experience physiological stress responses at received levels < 120 dB or <160 dB for continuous and impulsive sources respectively; these animals may exhibit slight deflection from the noise source, but this behavior is not likely to result in adverse consequences for the animals exhibiting that behavior.

Feeding humpbacks, however, may cease calling or alter vocalization at significantly lower received levels. Feeding bowhead whales have been known to respond to seismic noise at low received levels (85 dB-145 dB). Ferguson et al. (2015) identified areas around Sitka and overlapping with the action area as a Biologically Important Area (BIA) for humpback whale feeding during summer (June-August) and fall (September-November) (Figure 13). The proposed action activities will occur between May and September and would overlap with these BIAs for humpback whales. However, the predominant spawning areas for herring are over 24 km northeast of the project site (ECO49 2017).

Those animals that are closer to the source and not engaged in activities that would compete for their attentional resources (for example, foraging) might engage in low-level avoidance behavior (changing the direction or their movement to take them away from or tangential to the source of the disturbance) possibly accompanied by short-term vigilance behavior, but they are not likely to change their behavioral state (that is, animals that are foraging or migrating would continue to do so). We do not anticipate that low-level avoidance or short-term vigilance would occur until impulsive noise levels are >140 dB for humpback whales (McCauley et al. 2000). Females and females with calves may avoid sound sources \geq 140 dB. However, we would not anticipate the majority of individuals to show low-level avoidance until impulsive noise levels are \geq 150 dB (Lien et al. 1993, Richardson et al. 1995, Todd et al. 1996). Again, neither low level avoidance nor short-term vigilance is likely to result in adverse consequences for the animals exhibiting the behavior.

At some distance that is closer still, these species are likely to engage in more active avoidance behavior. Of the humpback whales and Steller sea lions that may be exposed to received levels ≥120 or ≥160 dB (for continuous and impulsive noise sources respectively) during the 197 instances of exposure from the proposed action (Level B harassment), some whales and sea lions are likely to reduce the amount of time they spend at the ocean's surface, increase their swimming speed, change their swimming direction to avoid construction operations, change their respiration rates, increase dive times, increase vigilance, reduce feeding behavior, or alter vocalizations and social interactions (Richardson et al. 1986, Ljungblad et al. 1988, Richardson and Malme 1993, Greene et al. 1999, Frid and Dill. 2002, Christie et al. 2009, Koski et al. 2009, Blackwell et al. 2010, Funk et al. 2010, Melcon et al. 2012). Based on the proposed action, we would expect these kind of responses at maximum distances out to 10 km for vibratory and DTH hammering (Scenarios 2 and 4), and distances out to 790 m for impact pile driving (Scenario 6) (Quijano and Austin 2017). However, while these exposures may all occur within the same year, they are anticipated to be separated in time and space considering the applicant does not anticipate more than two scenarios occurring over the course of a single day (ECO49 2017).

Many animals perform vital functions, such as feeding, resting, traveling, and socializing, on a diel cycle (24-hour cycle). Disruption of such functions resulting from reactions to stressors such

as sound exposure are more likely to be significant if they last more than one diel cycle or recur on subsequent days (Southall et al. 2007). Consequently, a behavioral response lasting less than one day and not recurring on subsequent days is not considered particularly severe unless it could directly affect reproduction or survival (Southall et al. 2007). Note that there is a difference between multi-day substantive behavioral reactions and multi-day anthropogenic activities.

Some whales or sea lions may be less likely to respond because they are feeding. The whales and seals that are exposed to these sounds probably would have prior experience with similar construction stressors resulting from their exposure during previous years; that experience will make some animals more likely to avoid the construction activities while others would be less likely to avoid those activities. In addition, standard mitigation measures (ramp ups and shut downs) will be in place along with monitoring measures. Some Mexico DPS humpback whales and western DPS Steller sea lions might experience physiological stress (but not distress) responses if they attempt to avoid one construction operation and encounter another construction operations (i.e., Sawmill Cove Dock replacement) while they are engaged in avoidance behavior.

Of the responses considered above, we do not expect PTS will occur. Nor do we consider exposure to vessel noise to rise to the level of take. We expect TTS, masking, behavioral responses, and physical and physiological effects may occur in Mexico DPS humpback whales and western DPS Steller sea lions. Though project activities may cause TTS, interruptions in communications (masking), avoidance of the action area, and stress associated with these disruptions in exposed individual whales and pinnipeds, we expect all effects will be temporary. Prey species may experience stress, injury, TTS, or changes in balance in a small radius directly around the pile driving or DTH hammering activities or startle and disperse when exposed to sounds from project activities. We do not expect effects to prey species will be sufficient to affect ESA-listed whales or pinnipeds.

7 CUMULATIVE EFFECTS

"Cumulative effects" are those effects of future state or private activities, not involving Federal activities, that are reasonably certain to occur within the action area (50 CFR 402.02). Future Federal actions that are unrelated to the proposed action are not considered in this section because they require separate consultation, per section 7 of the ESA.

We searched for information on non-Federal actions reasonably certain to occur in the action area. We did not find any information about non-Federal actions other than what has already been described in the Environmental Baseline (Section 5 of this opinion). We expect climate change, fisheries, harvest, noise, pollutants and discharges, scientific research, and ship strike will continue into the future. We expect moratoria on commercial whaling and bans on commercial sealing will remain in place, aiding in the recovery of ESA-listed whales and pinnipeds.

8 INTEGRATION AND SYNTHESIS OF EFFECTS

The Integration and Synthesis section is the final step of NMFS's assessment of the risk posed to species and critical habitat as a result of implementing the proposed action. In this section, we add the effects of the action (Section 6) to the environmental baseline (Section 5) and the

cumulative effects (Section 7) to formulate the agency's biological opinion as to whether the proposed action is likely to: (1) result in appreciable reductions in the likelihood of the survival or recovery of the species in the wild by reducing its numbers, reproduction, or distribution; or (2) result in the adverse modification or destruction of critical habitat as measured through potential reductions in the value of designated critical habitat for the conservation of the species. These assessments are made in full consideration of the status of the species (Section 4).

As we discussed in the *Approach to the Assessment* section of this opinion, we begin our risk analyses by asking whether the probable physical, physiological, behavioral, or social responses of endangered or threatened species are likely to reduce the fitness of endangered or threatened individuals or the growth, annual survival or reproductive success, or lifetime reproductive success of those individuals.

The survival and recovery of Mexico DPS humpback whales and western DPS Steller sea lion in the action area may be affected by:

- Climate change
 - o Prey distribution
 - o Habitat quality
- Fisheries interactions
- Subsistence harvests
- Natural and anthropogenic noise
- Pollutants and discharges
- Scientific research
- Ship strike

Despite these pressures, available trend information indicates western DPS Steller sea lion populations are increasing. Population trends for Mexico DPS humpbacks are not known; however, Hawaii DPS humpback which are also in the action area are growing at a rate of nearly 6 percent (Muto et al. 2017).

We concluded in the *Effects of the Action* (Section 6 of this opinion) that ESA-listed whales and pinnipeds may be harassed by the proposed activities. We expect the following number of whales and sea lions to represent the maximum number of individuals that will be exposed to Level A and Level B harassment associated with the proposed action:

- 3 (Level A) and 175 (Level B) takes of Western DPS Steller sea lions
- 0 (Level A) and 22 (Level B) takes of Mexico DPS humpback whales

We expect these exposures may cause TTS and interruptions in communication (i.e., masking) and could elicit the following behavioral responses:

- Temporary displacement from feeding areas
- Avoidance of the ensonified area

We expect low-level, brief stress responses will accompany these responses. We do not expect whales or pinnipeds exposed to these sounds will experience a continued reduction in fitness.

Prey species may experience stress, injury, TTS, changes in balance, or may be displaced when exposed to sounds from project activities. We do not expect these effects will limit the prey available to ESA-listed whales or pinnipeds.

In summary, we do not expect exposure to any of the stressors related to the proposed project to reduce fitness in any individual whale or pinnipeds. Therefore, we do not expect fitness consequences to ESA-listed whale or pinniped populations or species.

9 CONCLUSION

After reviewing the current status of ESA-listed species, the environmental baseline for the action area, the anticipated effects of the proposed activities, and the possible cumulative effects, it is NMFS's biological opinion that the FAA's proposed action, and PR1's proposed issuance of an IHA to FAA for the proposed Biorka Dock Reconstruction Project near Sitka, Alaska is not likely to jeopardize the continued existence of the following species:

- Mexico DPS Humpback whale
- Western DPS Steller sea lion

In addition, the proposed action is not likely to adversely affect the following species or critical habitat:

- Sperm whale
- Steller sea lion critical habitat

10 INCIDENTAL TAKE STATEMENT

Section 9 of the ESA prohibits the "take" of endangered species without special exemption. "Take" is defined as to harass, harm, pursue, hunt, shoot, wound, kill, trap, capture or collect, or to attempt to engage in any such conduct. "Incidental take" is defined as take that is incidental to, and not the purpose of, the carrying out of an otherwise lawful activity (50 CFR 402.02). Based on recent NMFS guidance, the term "harass" under the ESA means to: "create the likelihood of injury to wildlife by annoying it to such an extent as to significantly disrupt normal behavioral patterns which include, but are not limited to, breeding, feeding, or sheltering" (Wieting 2016). The MMPA defines "harassment" as: any act of pursuit, torment, or annoyance which (i) has the potential to injure a marine mammal or marine mammal stock in the wild [Level A harassment]; or (ii) has the potential to disturb a marine mammal or marine mammal stock in the wild by causing disruption of behavioral patterns, including, but not limited to, migration, breathing, nursing, breeding, feeding, or sheltering [Level B harassment] (16 U.S.C. §1362(18)(A)(i) and (ii)). Both Level A and Level B takes are anticipated and authorized for the proposed action.

Under the terms of sections 7(b)(4) and 7(o)(2), taking that is incidental and not intended as part of the agency action is not considered to be prohibited taking under the ESA, provided that such taking is in compliance with the terms and conditions of this Incidental Take Statement (ITS).

Section 7(b)(4)(C) of the ESA specifies that in order to provide an Incidental Take Statement for an endangered or threatened species of marine mammal, the taking must first be authorized under section 101(a)(5) of the MMPA. Accordingly, the terms of this Incidental Take Statement and the exemption from Section 9 of the ESA become effective only upon the issuance of MMPA authorization to take the marine mammals identified here. Absent such authorization, this Incidental Take Statement is inoperative.

The terms and conditions described below are nondiscretionary. FAA and NMFS PR1 have a continuing duty to regulate the activities covered by this ITS. In order to monitor the impact of incidental take, the FAA and PR1 must monitor the progress of the action and its impact on the species as specified in the ITS (50 CFR 402.14(i)(3)). If the FAA or PR1 (1) fail to require the authorization holder to adhere to the terms and conditions of the ITS through enforceable terms that are added to the authorization, and/or (2) fail to retain oversight to ensure compliance with these terms and conditions, the protective coverage of section 7(o)(2) may lapse.

10.1 Amount or Extent of Take

Section 7 regulations require NMFS to estimate the number of individuals that may be taken by proposed actions or utilize a surrogate (e.g., other species, habitat, or ecological conditions) if we cannot assign numerical limits for animals that could be incidentally taken during the course of an action (50 CFR § 402.14 (i)(1); see also 80 FR 26832 (May 11, 2015).

NMFS anticipates the proposed Biorka Island Dock Replacement Project, between May 1, 2018 and April 30, 2019, is likely to result in the incidental take of ESA-listed species by Level A and Level B harassment. As discussed in Section 6.2 of this opinion, the proposed action is expected to take the following number of ESA-listed individuals described in Table 12.

Table 12. Summary of instances of exposure associated with the proposed pile driving/removal and DTH hammering resulting in incidental take of ESA-listed species by Level A and Level B harassment.

Species	Proposed Authorized Level A Takes	Proposed Authorized Level B Takes	Anticipated Temporal Extent of Take
Western DPS Steller sea lion (Eumetopias jubatus)	3	175 ¹⁵	May 1, 2018 through April 30, 2019 ¹⁶
Mexico DPS Humpback whale (Megaptera novaeangliae)	0	2217	

Level B harassment of these individuals will occur by exposure to received sound from continuous sound sources with received sound levels of least 120 dB re 1 μ Pa_{rms} (i.e., vibratory or DTH hammering), or exposure to received sound from impulsive sound sources with received sound levels of least 160 dB re 1 μ Pa_{rms} (i.e., impact hammering). Level A takes are limited to western DPS Steller sea lion and will occur by exposure from impulsive sound sources (i.e.,

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¹⁵ The proposed IHA (<u>82 FR 41229</u>) indicated a requested Level A take of 6 Steller sea lion, and a Level B take of 350 Steller sea lion. Of the proposed takes, 50% are anticipated to occur to ESA-listed western DPS animals. The basis for this apportionment is described in Section 4.3.2

¹⁶ While the proposed IHA would be valid from May 1, 2018, through April 30, 2019, the allowed construction window within that timeframe is limited to May 1, 2018 through September 30, 2018.

¹⁷ The proposed IHA (<u>82 FR 41229</u>) indicated a requested Level A take of 3 humpback whales, and a Level B take of 350 humpback whales. Humpback whales in southeast Alaska include individuals from two DPSs. Of the 3 level A exposures, we anticipate zero threatened Mexico DPS animals will be taken, and all 3 will occur from the Hawaii DPS, which is not listed under the ESA. The basis for this apportionment is described in Section 4.3.1.

impact hammering) with received levels of at least 203 dB re1 μ Pa²·s. The take estimate is based on the best available information of whale and pinniped surveys and sightings in the area that will be ensonified from the proposed activities. Death or injury is not expected for any individual whales or pinnipeds that are exposed to these sounds.

ESA-listed whales and pinnipeds observed within the ZOI during pile removal/installation or DTH hammering will be considered to be taken, even if they exhibit no overt behavioral reactions due to the potential for unobservable physiological responses.

Any incidental take of ESA-listed whales and pinnipeds considered in this consultation is restricted to the permitted action as proposed. If the actual incidental take exceeds the predicted level or type, the FAA and PR1 must reinitiate consultation. Likewise, if the action deviates from what is described in Section 2 of this opinion, the FAA and PR1 must reinitiate consultation.

10.2 Effect of the Take

In Section 9 of this opinion, NMFS determined that the level of incidental take, coupled with other effects of the proposed action, is not likely to jeopardize the continued existence western DPS Steller sea lions or Mexico DPS humpback whales.

10.3 Reasonable and Prudent Measures

"Reasonable and prudent measures" are nondiscretionary measures to minimize the amount or extent of incidental take (50 CFR 402.02). NMFS concludes the reasonable and prudent measures described below, along with implementing terms and conditions, is necessary and appropriate to minimize the amount of incidental take of ESA-listed whales and pinnipeds resulting from the proposed actions. These measure is non-discretionary and must be binding conditions of the Permits Division's authorization for the exemption in ESA section 7(o)(2) to apply.

- 1. This ITS is valid only for the activities described in this opinion, and which have been authorized under section 101(a)(5) of the MMPA
- 2. The taking of western DPS Steller sea lion and Mexico DPS humpback whales shall be by incidental harassment only. The taking by serious injury or death is prohibited by and will result in the modification, suspension, or revocation of the ITS.
- 3. The FAA and PRA must implement and monitor the effectiveness of mitigation measures incorporated as part of the proposed authorization for the incidental taking of ESA-listed marine mammals pursuant to section 101(a)(5)(D) of the MMPA, as specified below. In addition, they must submit reports to NMFS AKR that evaluate the mitigation measures and report the results of the monitoring program, as specified below.

This Reasonable and Prudent Measure, along with its implementing terms and conditions, are designed to minimize the impact of incidental take that might otherwise result from the proposed action.

10.4 Terms and Conditions

To be exempt from the prohibitions of section 9 of the ESA, the FAA and PR1 must require any applicant or contractor to comply with the following terms and conditions, which implement the reasonable and prudent measure described above and outline the mitigation, monitoring, and

reporting measures required by section 7 regulations (50 CFR 402.14(i)). These terms and conditions are non-discretionary. If the FAA or PR1 to ensure compliance with these terms and conditions and their implementing reasonable and prudent measures, the protective coverage of section 7(o)(2) may lapse.

Partial compliance with these terms and conditions may result in more take than anticipated, and may invalidate this take exemption. These terms and conditions constitute no more than a minor change to the proposed action because they are consistent with the basic design of the proposed action.

To implement the reasonable and prudent measure, the FAA and PR1 must ensure that any applicant or contractor adheres to all portions of the description of the action (Section 2.1 Proposed Action), especially mitigation and monitoring measures described in Section 2.1.5 of this opinion. The FAA and PR1 must also ensure that any applicant or contractor adheres to the following Terms and Conditions: 18

- 1. FAA and PR1 shall require their permitted operators to possess a current and valid Incidental Harassment Authorization issued by NMFS under section 101(a)(5) of the MMPA, and any take must occur in compliance with all terms, conditions, and requirements included in such authorizations.
- 2. Monthly PSO reports, a final PSO report, and completed marine mammal observation record forms (developed by applicant) will be provided during the project. Items 2.1 through 2.4, below, provide details about what must be included in the reports.
 - 2.1. The reporting period for each monthly PSO report will be the entire calendar month, and reports will be submitted by close of business on the 5th business day of the month following the end of the reporting period (e.g., The monthly report covering May 1 through 31, 2018, will be submitted to NMFS Alaska Region by close of business [i.e., 5:00 pm, AKDT] on June 7, 2018).
 - 2.1.1. Completed marine mammal observation record forms, in electronic format, will be provided to NMFS Alaska Region in monthly reports.
 - 2.1.2. Observer report data will include the following for each listed marine mammal observation (or "sighting event" if repeated sightings are made of the same animal[s]):
 - 2.1.2.1. Species, date, and time for each sighting event
 - 2.1.2.2. Number of animals per sighting event and number of adults/juveniles/calves/pups per sighting event
 - 2.1.2.3. Primary, and, if observed, secondary behaviors of the marine mammals in each sighting event
 - 2.1.2.4. Geographic coordinates for the observed animals, with the position recorded by using the most precise coordinates practicable (coordinates must be recorded in decimal degrees, or similar standard, and defined coordinate system)
 - 2.1.2.5. Time and description of most recent project activity prior to marine mammal observation

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¹⁸ These terms and conditions are in addition to reporting required by PR1.

- 2.1.2.6. Environmental conditions as they existed during each sighting event, including, but not limited to:
 - 2.1.2.6.1. Beaufort Sea State
 - 2.1.2.6.2. Weather conditions
 - 2.1.2.6.3. Visibility (km/mi)
 - 2.1.2.6.4. Lighting conditions
- 2.1.3. Observer report data will also include the following for each take of a marine mammal that occurs in the manner and extent as described in Section 10 of this opinion:
 - 2.1.3.1. All information listed under Item 2.1.2, above
 - 2.1.3.2. Cause of the take (e.g., humpback within Level B zone during impact pile driving)
 - 2.1.3.3. Time the animal(s) entered the zone, and, if known, the time it exited the zone
 - 2.1.3.4. For takes of humpback whales and Steller sea lion, the observer report will estimate the probability of occurrence of ESA-listed DPSs out of the total estimated takes (e.g., Out of a total 350 Steller sea lions estimated to be taken by Level B harassment, western DPS 0.5 (350) = 175 sea lions may have been taken been taken)
- 2.2. A final technical report will be submitted to NMFS Alaska Region within 90 days after the proposed action has completed, and all vessels have left the action area. The report will summarize all project activities and results of marine mammal monitoring conducted during project activities. The final technical report will include all elements from Item 2.1, above, as well as:
 - 2.2.1. Summaries that include monitoring effort (e.g., total hours, total distances, and marine mammal distribution through the study period, accounting for sea state and other factors that affect visibility and detectability of marine mammals)
 - 2.2.2. Analyses on the effects from various factors that influences detectability of marine mammals (e.g., sea state, number of observers, fog, glare, etc.)
 - 2.2.3. Species composition, occurrence, and distribution of marine mammal sightings, including date, water depth, numbers, age/size/sex categories (if determinable), group sizes, and ice cover
 - 2.2.4. Species composition, occurrence, and distribution of marine mammal takes, including date, water depth, numbers, age/size/sex categories (if determinable), group sizes, and ice cover
 - 2.2.4.1. Humpback whale and Steller sea lion take estimates will be broken out by DPS.
 - 2.2.5. Analyses of effects of project activities on listed marine mammals
 - 2.2.6. Number of marine mammals observed and taken (by DPS) during periods with and without project activities (and other variables that could affect detectability), such as:
 - 2.2.6.1. Initial sighting distances versus project activity at time of sighting
 - 2.2.6.2. Observed behaviors and movement types versus project activity at time of sighting
 - 2.2.6.3. Numbers of sightings/individuals seen versus project activity at time of sighting

- 2.2.6.4. Distribution around the action area versus project activity at time of sighting
- 2.3. If unauthorized take occurs, (i.e., take of ESA-listed species not included in table 12), it must be reported to NMFS Alaska Region within one business day to the contact listed in Item 2.4, below. Observation records for ESA-listed marine mammals taken in a manner or to the extent other than described in Section 10.1 of this opinion must include:
 - 2.3.1. All information listed under Item 2.1, above
 - 2.3.2. Number of listed animals taken
 - 2.3.3. Date and time of each take
 - 2.3.4. Cause of the take (e.g., sperm whale observed within Level B zone or ship-strike of a humpback whale)
 - 2.3.5. Time the animal(s) entered the zone, and, if known, the time it exited the zone, if applicable
 - 2.3.6. Mitigation measures implemented prior to and after the animal entered the zone, if applicable
- 2.4. NMFS Contact:

Monthly and final reports and reports of unauthorized take will be submitted to: NMFS Alaska Region, Protected Resources Division

Alicia Bishop

alicia.bishop@noaa.gov

907-586-7224

11 CONSERVATION RECOMMENDATIONS

Section 7(a)(1) of the ESA directs Federal agencies to use their authorities to further the purposes of the ESA by carrying out conservation programs for the benefit of endangered and threatened species. Conservation recommendations are discretionary agency activities to minimize or avoid adverse effects of a proposed action on ESA-listed species or critical habitat, help implement recovery plans, or develop information (50 CFR 402.02).

We offer the following conservation recommendation, which will provide information for future consultations involving the issuance of permits that may affect ESA-listed whales and pinnipeds:

• **Behavioral responses of marine mammals**: We recommend that PR1 summarize findings from past IHA holders about behavioral responses of ESA-listed species to sounds from DTH hammering. Better understanding of how ESA-listed species have responded to sounds from past projects will inform our exposure and response analyses in the future.

In order for the NMFS Alaska Region to be kept informed of actions minimizing or avoiding adverse effects on, or benefiting, ESA-listed species or their habitats, PR1 should notify the NMFS Alaska Region of any conservation recommendations it implements.

12 REINITIATION NOTICE

This concludes formal consultation on the proposed Biorka Island Dock Replacement Project and PR1's issuance of an IHA to FAA. As provided in 50 CFR § 402.16, reinitiation of formal consultation is required where discretionary Federal agency involvement or control over the action has been retained (or is authorized by law) and if:

- The amount or extent of proposed take is exceeded.
- New information reveals effects of the agency action that may affect ESA-listed species or critical habitat in a manner, or to an extent, not considered in this opinion.
- The agency action is subsequently modified in a manner that causes an effect to the ESA-listed species, or critical habitat not considered in this opinion.
- A new species is ESA-listed or critical habitat designated that may be affected by the action.

In instances where the amount or extent of authorized take is exceeded, the FAA and PR1 must immediately request reinitiation of section 7 consultation.

13 DATA QUALITY ACT DOCUMENTATION AND PRE-DISSEMINATION REVIEW

Section 515 of the Treasury and General Government Appropriations Act of 2001 (Public Law 106-554) (Data Quality Act (DQA)) specifies three components contributing to the quality of a document. They are utility, integrity, and objectivity. This section of the opinion addresses these DQA components, documents compliance with the DQA, and certifies that this opinion has undergone pre-dissemination review.

13.1 Utility

This document records the results of an interagency consultation. The information presented in this document is useful to NMFS AKR, PR1, FAA and the general public. These consultations help to fulfill multiple legal obligations of the named agencies. The information is also useful and of interest to the general public as it describes the manner in which public trust resources are being managed and conserved. The information presented in these documents and used in the underlying consultations represents the best available scientific and commercial information and has been improved through interaction with the consulting agency.

This consultation will be posted on the NMFS Alaska Region website http://alaskafisheries.noaa.gov/pr/biological-opinions/. The format and name adhere to conventional standards for style.

13.2 Integrity

This consultation was completed on a computer system managed by NMFS in accordance with relevant information technology security policies and standards set out in Appendix III, 'Security of Automated Information Resources,' Office of Management and Budget Circular A-130; the Computer Security Act; and the Government Information Security Reform Act.

13.3 Objectivity

Standards: This consultation and supporting documents are clear, concise, complete, and unbiased; and were developed using commonly accepted scientific research methods. They adhere to published standards including the ESA Consultation Handbook, ESA Regulations, 50 CFR § 402.01 et seq.

Best Available Information: This consultation and supporting documents use the best available information, as referenced in the literature cited section. The analyses in this opinion contain more background on information sources and quality.

Referencing: All supporting materials, information, data and analyses are properly referenced, consistent with standard scientific referencing style.

Review Process: This consultation was drafted by NMFS staff with training in ESA implementation, and reviewed in accordance with Alaska Region ESA quality control and assurance processes.

14 REFERENCES

- Andersson, A. J., D. I. Kline, P. J. Edmunds, S. D. Archer, N. Bednarsek, R. C. Carpenter, M. Chadsey, P. Goldstein, A. G. Grottoli, T. P. Hurst, A. L. King, J. E. Kubler, I. B. Kuffner, K. R. M. Mackey, B. A. Menge, A. Paytan, U. Riebesell, A. Schnetzer, M. E. Warner, and R. C. Zimmerman. 2015. Understanding ocean acidification impacts on organismalto ecological scales. Oceanography 28:16-27.
- Anisman, H., and Z. Merali. 1999. Understanding stress: Characteristics and caveats. Alcohol Research and Health **23**:241-249.
- Au, W. W., A. A. Pack, M. O. Lammers, L. M. Herman, M. H. Deakos, and K. Andrews. 2006a. Acoustic properties of humpback whale songs. The Journal of the Acoustical Society of America **120**:1103-1110.
- Au, W. W. L. 2000. Hearing in whales and dolphins: an overview. Pages 1-42 *in* W. W. L. Au, A. N. Popper, and R. R. Fay, editors. Hearing by Whales and Dolphins. Springer-Verlag, New York.
- Au, W. W. L., A. A. Pack, M. O. Lammers, L. M. Herman, M. H. Deakos, and K. Andrews. 2006b. Acoustic properties of humpback whale songs. Journal of the Acoustical Society of America **120**:1103-1110.
- Bain, D. E., and R. Williams. 2006. Long-range effects of airgun noise on marine mammals: Responses as a function of received sound level and distance. IWC Scientific Committee, St. Kitts and Nevis, West Indies.
- Baker, C. S. 1985. The behavioral ecology and populations structure of the humpback whale (*Megaptera novaeangliae*) in the central and eastern Pacific. University of Hawaii at Manoa.
- Baker, C. S., L. M. Herman, B. G. Bays, and G. B. Bauer. 1983. The impact of vessel traffic on the behavior of humpback whales in southeast Alaska: 1982 season. Report submitted to the National Marine Mammal Laboratory, Seattle, Washington, Kewalo Basin Marine Mammal Laboratory, University of Hawaii, Honolulu, HI.
- Bauer, G. B., and L. M. Herman. 1986. Effects of vessel traffic on the behavior of humpback whales in Hawai'i. Report Submitted to NMFS Southwest Region, U.S. Department of Commerce, National Oceanic and Atmospheric Administration, National Marine Fisheries Service, Southwest Fisheries Science Center, Western Pacific Program Office; Honolulu, Hawai'i.
- Baulch, S., and C. Perry. 2014. Evaluating the impacts of marine debris on cetaceans. Marine Pollution Bulletin **80**:210-221.
- Beale, C. M., and P. Monaghan. 2004a. Behavioural responses to human disturbance: a matter of choice? Animal Behaviour **68**:1065-1069.
- Beale, C. M., and P. Monaghan. 2004b. Human disturbance: people as predation-free predators? Journal of Applied Ecology **41**:335-343.
- Bejder, L., A. Samuels, H. Whitehead, H. Finn, and S. Allen. 2009. Impact assessment research: use and misuse of habituation, sensitisation and tolerance to describe wildlife responses to anthropogenic stimuli. Marine Ecology Progress Series **395**:177-185.
- Bejder, L., A. Samuels, H. Whitehead, N. Gales, J. Mann, R. Connor, M. Heithaus, J. Watson-Capps, C. Flaherty, and M. Krutzen. 2006. Decline in relative abundance of bottlenose dolphins exposed to long-term disturbance. Conservation Biology **20**:1791-1798.
- Bettridge, S., C. S. Baker, J. Barlow, P. J. Clapham, M. Ford, D. Gouveia, D. K. Mattila, R. M.

- Pace III, P. E. Rosel, G. K. Silber, and P. R. Wade. 2015. Status review of the humpback whale (*Megaptera novaeangliae*) under the Endangered Species Act. NOAA-TM-NMFS-SWFSC-540, Southwest Fisheries Science Center, National Marine Fisheries Service, National Oceanic and Atmospheric Administration, U.S. Department of Commerce, La Jolla, California.
- Blackwell, S. B., K. H. Kim, W. C. Burgess, R. G. Norman, and C. R. Greene. 2010. Underwater sounds near Northstar during late summer and autumn of 2005-2009. Pages 4-1 to 4-57 *in* W. J. Richardson, editor. Monitoring of industrial sounds, seals, and bowhead whales near BP's Northstar Oil Development, Alaskan Beaufort Sea: Comprehensive report for 2005–2009.
- Blackwell, S. B., J. W. Lawson, and M. T. Williams. 2004. Tolerance by ringed seals (*Phoca hispida*) to impact pipe-driving and construction sounds at an oil production island. Journal of the Acoustical Society of America **115**:2346-2357.
- Blackwell, S. B., C. S. Nations, T. L. McDonald, C. R. Greene, A. M. Thode, M. Guerra, and A. Michael Macrander. 2013. Effects of airgun sounds on bowhead whale calling rates in the Alaskan Beaufort Sea. Marine Mammal Science **29**:E342-E365.
- Blecha, F. 2000. Immune system response to stress. Pages 111-122 *in* G. P. Moberg and J. A. Mench, editors. The biology of animal stress. CAB International, Oxon.
- Blumstein, D. T. 2003. Flight-initiation distance in birds is dependent on intruder starting distance. The Journal of Wildlife Management **67**:852-857.
- Brumm, H. 2004. The impact of environmental noise on song amplitude in a territorial bird. Journal of Animal Ecology **73**:434-440.
- Calambokidis, J., E. A. Falcone, T. J. Quinn, A. M. Burdin, P. J. Clapham, J. K. B. Ford, C. M. Gabriele, R. LeDuc, D. Mattila, L. Rojas-Bracho, J. M. Straley, B. L. Taylor, J. U. R., D. Weller, B. H. Witteveen, M. Yamaguchi, A. Bendlin, D. Camacho, K. Flynn, A. Havron, J. Huggins, and N. Maloney. 2008. SPLASH: Structure of populations, levels of abundance and status of humpback whales in the North Pacific U.S. Department of Commerce, Western Administrative Center, Seattle, Washington.
- Calkins, D. G., and K. W. Pitcher. 1982. Population assessment, ecology and trophic relationships of Steller sea lions in the Gulf of Alaska. Outer Continental Shelf Environmental Assessment Program, U. S. Department of the Interior.
- Cerchio, S., J. K. Jacobsen, D. M. Cholewiak, and E. A. Falcone. 2005. Reproduction of female humpback whales off the Revillagigedo Archipelago during a severe El Niño event. Page 55 Sixteenth Biennial Conference on the Biology of Marine Mammals, San Diego, California.
- Christie, K., C. Lyons, W. R. Koski, D. S. Ireland, and D. W. Funk. 2009. Patterns of bowhead whale occurrence and distribution during marine seismic operations in the Alaskan Beaufort Sea. Page 55 Eighteenth Biennial Conference on the Biology of Marine Mammals, Quebec City, Canada.
- Chu, K., C. Sze, and C. Wong. 1996. Swimming behaviour during the larval development of the shrimp Metapenaeus ensis (De Haan, 1844)(Decapoda, Penaeidae). Crustaceana **69**:368-378.
- Clapham, P. J., and D. K. Mattila. 1993. Reactions of Humpback Whales to Skin Biopsy Sampling on a West-Indies Breeding Ground. Marine Mammal Science **9**:382-391.
- Clark, C., W. T. Ellison, B. Southall, L. Hatch, S. M. Van Parijs, A. S. Frankel, D. Ponirakis, and G. C. Gagnon. 2009a. Acoustic masking of baleen whale communications: potential

- impacts from anthropogenic sources. Page 56 Eighteenth Biennial Conference on the Biology of Marine Mammals, Quebec City, Canada.
- Clark, C. W., W. T. Ellison, B. L. Southall, L. Hatch, S. M. Van Parijs, A. Frankel, and D. Ponirakis. 2009b. Acoustic masking in marine ecosystems: Intuitions, analysis, and implication. Marine Ecology Progress Series **395**:201-222.
- Cody, M. L., and J. H. Brown. 1969. Song Asynchrony in Neighbouring Bird Species. Nature **222**:778-781.
- Cox, T. M., T. Ragen, A. Read, E. Vos, R. Baird, K. Balcomb, J. Barlow, J. Caldwell, T. Cranford, and L. Crum. 2006. Understanding the impacts of anthropogenic sound on beaked whales. SPACE AND NAVAL WARFARE SYSTEMS CENTER SAN DIEGO CA.
- Crespi, E. J., T. D. Williams, T. S. Jessop, and B. Delehanty. 2013. Life history and the ecology of stress: how do glucocorticoid hormones influence life-history variation in animals? Functional Ecology **27**:93-106.
- D'Vincent, C. G., R. M. Nilson, and R. E. Hanna. 1985. Vocalization and coordinated feeding behavior of the humpback whale in southeastern Alaska. Scientific Reports of the Whales Research Institute **36**:41–47.
- de Kloet, E. R., M. Joels, and F. Holsboer. 2005. Stress and the brain: From adaptation to disease. Nature Reviews Neuroscience **6**:463-475.
- DeMarban, A., and L. Demer. 2016. Western Alaska hunters may be in trouble after landing offlimits whale. Alaska Dispatch News. Alaska Dispatch News, Anchorage, Alaska
- Di Lorio, L., and C. W. Clark. 2010. Exposure to seismic survey alters blue whale acoustic communication. Biology Letters **6**:51-54.
- Doney, S. C., M. Ruckelshaus, J. E. Duffy, J. P. Barry, F. Chan, C. A. English, H. M. Galindo, J. M. Grebmeier, A. B. Hollowed, and N. Knowlton. 2012. Climate change impacts on marine ecosystems. Marine Science 4.
- ECO49. 2017. Request for Incidental Harassment Authorization Biorka Island Dock Replacement Sitka, Alaska. Prepared for Federal Aviation Administration (FAA) Alaska Region. Received August 4, 2017.
- Ellison, W., B. Southall, C. Clark, and A. Frankel. 2012. A New Context-Based Approach to Assess Marine Mammal Behavioral Responses to Anthropogenic Sounds. Conservation Biology **26**:21-28.
- Elowson, A. M., P. L. Tannenbaum, and C. T. Snowdon. 1991. Food-associated calls correlate with food preferences in cotton-top tamarins. Animal Behaviour **42**:931-937.
- EPA. 2013. Vessel general permit for discharges incidental to the normal operation of vessels (VGP): authorization to discharge under the National Pollutant Discharge Elimination System. U.S. Environmental Protection Agency.
- Erbe, C. 2002a. Hearing Abilities of Baleen Whales. Defense Research and Development Canada, Ottawa, Ont.
- Erbe, C. 2002b. Hearing abilities of baleen whales. Atlantic report CR 2002-065. Contract Number: W7707-01-0828. Defence R&D Canada.
- Everitt, R., C. Fiscus, and R. DeLong. 1980. Northern Puget Sound marine mammals. Interagency Energy. Environment R & D Program Report, US EPA, EPA-600/7-80-139. US EPA, Washington, DC.
- FAA. 2016. Request for Informal Section 7 Endangered Species Act Consultation for Biorka Island Dock Replacement Project, Biorka Island, Sitka, Alaska. Federal Aviation

- Administration (FAA), Alaska Region. Received May 23, 2016.
- FAA. 2017a. Finding of No Significant Impact and Record of Decision for the Biorka Dock Replacement Project Environmental Assessment. Received September 2017. 107 pages.
- FAA. 2017b. Request for Initiation of Formal Consultation for Biorka Island Dock Replacement Project, Sitka, Alaska. Received August 15, 2017.
- Fair, P. A., and P. R. Becker. 2000. Review of stress in marine mammals. Journal of Aquatic Ecosystem Stress and Recovery **7**:335-354.
- Fay, R. R., and A. N. Popper. 2012. Fish hearing: New perspectives from two senior bioacousticians. Brain, Behavior and Evolution **79**:215-217.
- Ferguson, M. C., C. Curtice, and J. Harrison. 2015. Biologically important areas for cetaceans within U.S. waters Gulf of Alaska region. Aquatic Mammals **41**:65-78.
- Finneran, J. J., D. A. Carder, and S. H. Ridgway. 2003. Temporary threshold shift (TTS) measurements in bottlenose dolphins (*Tursiops truncatus*), belugas (*Delphinapterus leucas*), and California sea lions (*Zalophus californianus*). Environmental Consequences of Underwater Sound (ECOUS) Symposium, San Antonio, Texas
- Finneran, J. J., D. A. Carder, C. E. Schlundt, and R. L. Dear. 2010a. Temporary threshold shift in a bottlenose dolphin (*Tursiops truncatus*) exposed to intermittent tones. Journal of the Acoustical Society of America **127**:3267-3272.
- Finneran, J. J., D. A. Carder, C. E. Schlundt, and S. H. Ridgway. 2005. Temporary threshold shift in bottlenose dolphins (*Tursiops truncatus*) exposed to mid-frequency tones. Journal of the Acoustical Society of America **118**:2696-2705.
- Finneran, J. J., D. S. Houser, P. W. Moore, B. K. Branstetter, J. S. Trickey, and S. H. Ridgway. 2010b. A method to enable a bottlenose dolphin (*Tursiops truncatus*) to echolocate while out of water. Journal of the Acoustical Society of America **128**:1483-1489.
- Finneran, J. J., and C. E. Schlundt. 2010. Frequency-dependent and longitudinal changes in noise-induced hearing loss in a bottlenose dolphin (*Tursiops truncatus*). Journal of the Acoustical Society of America **128**:567-570.
- Finneran, J. J., and C. E. Schlundt. 2013. Effects of fatiguing tone frequency on temporary threshold shift in bottlenose dolphins (*Tursiops truncatus*). Journal of the Acoustical Society of America **133**:1819-1826.
- Finneran, J. J., C. E. Schlundt, B. Branstetter, and R. L. Dear. 2007. Assessing temporary threshold shift in a bottlenose dolphin (*Tursiops truncatus*) using multiple simultaneous auditory evoked potentials. Journal of the Acoustical Society of America **122**:1249-1264.
- Finneran, J. J., C. E. Schlundt, R. Dear, D. A. Carder, and S. H. Ridgway. 2000. Masked temporary threshold shift (MTTS) in odontocetes after exposure to single underwater impulses from a seismic watergun. Journal of the Acoustical Society of America 108:2515.
- Finneran, J. J., C. E. Schlundt, R. Dear, D. A. Carder, and S. H. Ridgway. 2002. Temporary shift in masked hearing thresholds in odontocetes after exposure to single underwater impulses from a seismic watergun. Journal of the Acoustical Society of America 111:2929-2940.
- Foote, A. D., R. W. Osborne, and A. R. Hoelzel. 2004. Environment Whale-call response to masking boat noise. Nature **428**:910-910.
- Foreman, M. G., and Y. E. Yamanaka. 2011. Report of Working Group 20 on Evaluations of Climate Change Projections. PICES Sci. Rep. No. 40, 165 pp.
- Francis, C. D., and J. R. Barber. 2013. A framework for understanding noise impacts on wildlife: An urgent conservation priority. Frontiers in Ecology and the Environment **11**:305-313.

- Frankel, A. S., and C. W. Clark. 1998. Results of low-frequency playback of M-sequence noise to humpback whales, *Megaptera novaeangliae*, in Hawai'i. Canadian Journal of Zoology-Revue Canadienne De Zoologie **76**:521-535.
- Frazer, L. N., and E. Mercado. 2000. A sonar model for humpback whale song. IEEE Journal of Oceanic Engineering **25**:160-182.
- Frazer, L. N., and E. Mercado III. 2000. A sonar model for humpback whale song. Ieee Journal of Oceanic Engineering **25**:160-182.
- Frid, A., and L. M. Dill. 2002. Human-caused disturbance stimuli as a form of predation risk. 6(1): 11. [online] URL: . Conservation Ecology **6**:1-16.
- Frid, A., and L. M. Dill. 2002. Human-caused disturbance stimuli as a form of predation risk. 6(1): 11. [online] URL: . Conservation Ecology **6**:1-16.
- Fristrup, K. M., L. T. Hatch, and C. W. Clark. 2003. Variation in humpback whale (Megaptera novaeangliae) song length in relation to low-frequency sound broadcasts. The Journal of the Acoustical Society of America 113:3411-3424.
- Fritz, L., K. Sweeney, R. Towell, and T. Gelatt. 2016. Aerial and ship-based surveys of Steller sea lions (*Eumetopias jubatus*) conducted in Alaska in June-July 2013 through 2015, and an update on the status and trend of the western distinct population segment in Alaska, U.S. Dept. Commerce, NOAA Tech. Memo NMFS-AFSC-321. 72 p.
- Fritz, L., K. Sweeney, D. Johnson, M. Lynn, T. Gelatt, and J. Gilpatrick. 2013. Aerial and ship-based surveys of Steller sea lions (*Eumetopias jubatus*) conducted in Alaska in June-July 2008 through 2012, and an update on the status and trend of the western distinct population segment in Alaska. National Oceanic and Atmospheric Administration, National Marine Fisheries Service, Alaska Fisheries Science Center.
- Funk, D. W., R. Rodrigues, D. S. Ireland, and W. R. Koski. 2010. Summary and assessment of potential effects on marine mammals. Pages 11-11 11-59 *in* I. D. Funk DW, Rodrigues R, and Koski WR, editor. Joint Monitoring Program in the Chukchi and Beaufort seas, open water seasons, 2006–2008.
- GAO. 2014. Ocean Acidification. Report to Congress. . September 2014. United States Government Accountability Office GAO 14-736. 41 p.
- Gill, J. A., and W. J. Sutherland. 2001. Predicting the consequences of human disturbance from behavioral decisions. Pages 51-64 *in* L. M. Gosling and W. J. Sutherland, editors. Behavior and Conservation. Cambridge University Press, Cambridge.
- Gilly, W. F., J. M. Beman, S. Y. Litvin, and B. H. Robison. 2013. Oceanographic and Biological Effects of Shoaling of the Oxygen Minimum Zone. Annual Review of Marine Science 5:393.
- Greene, C. R., N. S. Altman, W. J. Richardson, and R. W. Blaylock. 1999. Bowhead Whale Calls. Page 23 Marine Mammal and Acoustical Monitoring of Western Geophysical's Open-Water Seismic Program in the Alaskan Beaufort Sea, 1998. LGL Ecological Research Associates, Inc, King City, Ontario, Canada.
- Greig-smith, P. W. 1980. Parental investment in nest defense by stonechats (*Saxicola torquata*). Animal Behaviour **28**:604-619.
- Halvorsen, M. B., B. M. Casper, C. M. Woodley, T. J. Carlson, and A. N. Popper. 2012. Threshold for onset of injury in Chinook salmon from exposure to impulsive pile driving sounds. PLoS ONE 7:e38968.
- Harris, R. E., G. W. Miller, and W. J. Richardson. 2001. Seal responses to airgun sounds during summer seismic surveys in the Alaskan Beaufort Sea. Marine Mammal Science 17:795-

812.

- Hart Crowser. 2009. Acoustic Monitoring and In-situ Exposures of Juvenile Coho Salmon to Pile Driving at the Port of Anchorage Marine Terminal Redevelopment Project Knik Arm, Anchorage, Alaska. October 2009.
- Hazen, E. L., S. Jorgensen, R. R. Rykaczewski, S. J. Bograd, D. G. Foley, I. D. Jonsen, S. A. Shaffer, J. P. Dunne, D. P. Costa, L. B. Crowder, and B. A. Block. 2012. Predicted habitat shifts of Pacific top predators in a changing climate. Nature Climate Change Letters.
- Heide-Jorgensen, M. P., K. L. Laidre, O. Wiig, M. V. Jensen, L. Dueck, L. D. Maiers, H. C. Schmidt, and R. C. Hobbs. 2003. From Greenland to Canada in ten days: Tracks of bowhead whales, *Balaena mysticetus*, across Baffin Bay. Arctic **56**:21-31.
- Herman, J. P., and W. E. Cullinan. 1997. Neurocircuitry of stress: Central control of the hypothalamo-pituitary-adrenocortical axis. Trends in Neurosciences **20**:78-84.
- Holt, M. M., D. P. Noren, V. Veirs, C. K. Emmons, and S. Veirs. 2009. Speaking up: Killer whales (*Orcinus orca*) increase their call amplitude in response to vessel noise. Journal of the Acoustical Society of America **125**:EL27-EL32.
- IPCC. 2013. Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change. New York, NY.
- IPCC. 2014a. Climate change 2014: Impacts, adaptation, and vulnerability. IPCC Working Group II contribution to AR5. Intergovernmental Panel on Climate Change.
- IPCC, editor. 2014b. Summary for policymakers. Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA.
- Iverson, S. J., K. J. Frost, and S. L. Lang. 2002. Fat content and fatty acid composition of forage fish and invertebrates in Prince William Sound, Alaska: factors contributing to among and within species variability. Marine Ecology Progress Series **241**:161-181.
- Jansen, J. K., J. L. Bengtson, P. L. Boveng, S. P. Dahle, and J. M. Ver Hoef. 2006. Disturbance of harbor seals by cruise ships in Disenchantment Bay, Alaska: an investigation at three spatial and temporal scales. Alaska Fish. Sci. Cent., Natl. Mar. Fish. Serv., Seattle WA.
- Jeffries, M. O., J. Richter-Menge, and J. E. Overland, Eds. 2014. Arctic report card 2014. Available online at: http://www.arctic.noaa.gov/reportcard.
- Jemison, L. A., G. W. Pendleton, L. W. Fritz, K. K. Hastings, J. M. Maniscalco, A. W. Trites, and T. S. Gelatt. 2013. Inter-population movements of Steller sea lions in Alaska with implications for population separation. PLoS ONE 8:e70167.
- Jessop, T. S., A. D. Tucker, C. J. Limpus, and J. M. Whittier. 2003. Interactions between ecology, demography, capture stress, and profiles of corticosterone and glucose in a free-living population of Australian freshwater crocodiles. General and comparative endocrinology **132**:161-170.
- Johnson, H. D., K. M. Stafford, J. C. George, W. G. Ambrose, and C. W. Clark. 2014. Song sharing and diversity in the Bering-Chukchi-Beaufort population of bowhead whales (Balaena mysticetus), spring 2011. Marine Mammal Science:n/a-n/a.
- Kajimura, H., and T. R. Loughlin. 1988. Marine mammals in the oceanic food web of the eastern subarctic Pacific. Bulletin of the Ocean Research Institute, University of Tokyo **26**:187-223.
- Kastak, D., B. L. Southall, R. J. Schusterman, and C. R. Kastak. 2005. Underwater temporary threshold shift in pinnipeds: effects of noise level and duration. Journal of the Acoustical

- Society of America **118**:3154-3163.
- Kastelein, R. A., R. Gransier, L. Hoek, and C. A. F. d. Jong. 2012a. The hearing threshold of a harbor porpoise (*Phocoena phocoena*) for impulsive sounds (L). Journal of the Acoustical Society of America **132**:607-610.
- Kastelein, R. A., R. Gransier, L. Hoek, and J. Olthuis. 2012b. Temporary threshold shifts and recovery in a harbor porpoise (*Phocoena phocoena*) after octave-band noise at 4 kHz. Journal of the Acoustical Society of America **132**:3525-3537.
- Kenyon, K. W., and D. W. Rice. 1961. Abundance and distribution of the Steller sea lion. (Eumetopias jubatus). Journal of Mammalogy **42**:223-234.
- Ketten, D. R., J. Lien, and S. Todd. 1993. Blast injury in humpback whale ears: evidence and implications (Abstract). Journal of the Acoustical Society of America **94**:1849-1850.
- Kight, C. R., and J. P. Swaddle. 2011. How and why environmental noise impacts animals: an integrative, mechanistic review. Ecology Letters.
- Koski, W. R., D. W. Funk, D. S. Ireland, C. Lyons, K. Christie, A. M. Macrander, and S. B. Blackwell. 2009. An update on feeding by bowhead whales near an offshore seismic survey in the central Beaufort Sea.
- Krieger, K. J., and B. L. Wing. 1984. Hydroacoustic Surveys and Identification of Humpback Whale Forage in Glacier Bay, Stephens Passage, and Frederick Sound, Southeastern Alaska Summer 1983. NMFS; Auke Bay Lab., Auke Bay, AK.
- Kryter, K. D., W. D. Ward, J. D. Miller, and D. H. Eldredge. 1966. Hazardous exposure to intermittent and steady-state noise. Journal of the Acoustical Society of America **39**:451-464.
- Kucey, L. 2005. Human disturbance and the hauling out behavior of steller sea lions (*Eumetopias jubatus*). University of British Columbia, British Columbia.
- Laiolo, P., M. Vögeli, D. Serrano, and J. L. Tella. 2008. Song Diversity Predicts the Viability of Fragmented Bird Populations. PLoS ONE **3**:e1822.
- Lankford, S., T. Adams, R. Miller, and J. Cech Jr. 2005. The cost of chronic stress: impacts of a nonhabituating stress response on metabolic variables and swimming performance in sturgeon. Physiological and Biochemical Zoology **78**:599-609.
- Lefebvre, K. A., L. Quakenbush, E. Frame, K. B. Huntington, G. Sheffield, R. Stimmelmayr, A. Bryan, P. Kendrick, H. Ziel, T. Goldstein, J. A. Snyder, T. Gelatt, F. Gulland, B. Dickerson, and V. Gill. 2016. Prevalence of algal toxins in Alaskan marine mammals foraging in a changing arctic and subarctic environment. Harmful Algae **55**:13-24.
- Liddle, J. B., T.J. Quinn II, J.M. Straley, and M. Adkinson. 2015a. Humpback whale predation on Pacific Herring in Sitka Sound, Chapter 3, pp 59-103 in PhD. Dissertation. Univ. of Alaska Fairbanks, Fairbanks, Alaska. August 2015.
- Liddle, J. B., T.J. Quinn II, J. M. Straley, and M. Adkinson. 2015b. Humpback whale abundance associated with Sitka Sound, 1989-2008, Chapter 2, pp. 33-58 in PhD. Dissertation. Univ. of Alaska Fairbanks, Fairbanks, Alaska. August 2015.
- Lien, J., S. Todd, P. Stevick, F. Marques, and D. Ketten. 1993. The reaction of humpback whales to underwater explosions: orientation, movements, and behavior. Journal of the Acoustical Society of America **94**:1849.
- Lima, S. L., and L. M. Dill. 1990. Behavioral decisions made under the risk of predation a review and prospectus. Canadian Journal of Zoology-Revue Canadienne De Zoologie **68**:619-640.
- Ljungblad, D. K., B. Wursig, S. L. Swartz, and J. M. Keene. 1988. Observations on the

- behavioral responses of bowhead whales (Balaena mysticetus) to active geophysical vessels in the Alaskan Beaufort Sea. Arctic **41**:183-194.
- Lomac-MacNair, K., C. Thissen, and M.A. Smultea. 2014. Draft NMFS 90-Day Report for Marine Mammal Monitoring and Mitigation during SAExploration's Colville River Delta 3D Seismic Survey, Beaufort Sea, Alaska, August to September 2014. Submitted to SAE, Prepared by Smultea Environmental Sciences, P.O. Box 256, Preston, WA 98050. December 2, 2014., Preston, WA.
- Loughlin, T. R. 1997. Using the phylogeographic method to identify Steller sea lion stocks. Molecular Genetics of Marine Mammals **Spec. Pub. 3**:159-171.
- Loughlin, T. R., and A. E. York. 2000. An accounting of the sources of Steller sea lion, *Eumetopias jubatus*, mortality. Marine Fisheries Review **62**:40-45.
- Lucke, K., U. Siebert, P. A. Lepper, and M.-A. Blanchet. 2009. Temporary shift in masked hearing thresholds in a harbor porpoise (*Phocoena phocoena*) after exposure to seismic airgun stimuli. Journal of the Acoustical Society of America **125**:4060-4070.
- Macleod, C. D. 2009. Global climate change, range changes and potential implications for the conservation of marine cetaceans: A review and synthesis. Endangered Species Research 7:125-136.
- Marler, P., A. Dufty, and R. Pickert. 1986. Vocal communication in the domestic chicken. 1. Does a sender communicate information about the quality of a food referent to a receiver. Animal Behaviour **34**:188-193.
- Marston, B. H., M. F. Willson, and S. M. Gende. 2002. Predator aggregations during eulachon Thaleichthys pacificus spawning runs. Marine Ecology Progress Series **231**:229-236.
- Mathias, D., A. M. Thode, J. Straley, J. Calambokidis, G. S. Schorr, and K. Folkert. 2012. Acoustic and diving behavior of sperm whales (Physeter macrocephalus) during natural and depredation foraging in the Gulf of Alaska. The Journal of the Acoustical Society of America 132:518-532.
- McCauley, R. D., J. Fewtrell, A. J. Duncan, C. Jenner, M.-N. Jenner, J. D. Penrose, R. I. T. Prince, A. Adhitya, J. Murdoch, and K. McCabe. 2000. Marine seismic surveys a study of environmental implications. APPEA Journal **40**:692-708.
- McDonald, M. A., J. A. Hildebrand, and S. M. Wiggins. 2006. Increases in deep ocean ambient noise in the Northeast Pacific west of San Nicolas Island, California. Journal of the Acoustical Society of America **120**:711-718.
- McLean, K., pers. comm. 2017. Response to Additional Information Request Biorka Island Dock Reconstruction Project. From Kristi McLean (R&M consultations, INC) to Alicia Bishop (NMFS) regarding tension anchor installation and vessel use. Received Monday October 9, 2017.
- Melcon, M. L., A. J. Cummins, S. M. Kerosky, L. K. Roche, S. M. Wiggins, and J. A. Hildebrand. 2012. Blue whales respond to anthropogenic noise. PLoS ONE 7:e32681.
- Miller, G., V. Moulton, R. Davis, M. Holst, P. Millman, A. MacGillivray, and D. Hannay. 2005. Monitoring seismic effects on marine mammals—southeastern Beaufort Sea, 2001-2002. Offshore oil and gas environmental effects monitoring/Approaches and technologies. Battelle Press, Columbus, OH:511-542.
- Miller, G. W., R. E. Elliot, W. R. Koski, V. D. Moulton, and W. J. Richardson. 1999. Marine mammal and acoustical monitoring of Western Geophysical's open-water seismic program in the Alaskan Beaufort Sea, 1998. LGL, Ltd.
- Miller, P. J. O., N. Biassoni, A. Samuels, and P. L. Tyack. 2000. Whale songs lengthen in

- response to sonar. Nature 405:903-903.
- MMS (Mineral Management Service). 2008. Beaufort Sea and Chukchi Sea Planning Areas Oil and Gas Lease Sales 209, 212, 217, and 221 Draft environmental impact statement Alaska OCS Region, Anchorage, AK.
- Moberg, G. P. 2000. Biological response to stress: implications for animal welfare. The biology of animal stress: basic principles and implications for animal welfare:1-21.
- Mooney, T. A., P. E. Nachtigall, M. Breese, S. Vlachos, and W. W. L. Au. 2009a. Predicting temporary threshold shifts in a bottlenose dolphin (*Tursiops truncatus*): The effects of noise level and duration. Journal of the Acoustical Society of America **125**:1816-1826.
- Mooney, T. A., P. E. Nachtigall, and S. Vlachos. 2009b. Sonar-induced temporary hearing loss in dolphins. Biology Letters **5**:565-567.
- Morete, M. E., T. L. Bisi, and S. Rosso. 2007. Mother and calf humpback whale responses to vessels around the Abrolhos Archipelago, Bahia, Brazil. Journal of Cetacean Research and Management 9:241-248.
- Murray, C. S., A. Malvezzi, C. J. Gobler, and H. Baumann. 2014. Offspring sensitivity to ocean acidification changes seasonally in a coastal marine fish. Marine Ecology Progress Series **504**:1-11.
- Muto, M. M., V. T. Helker, R. P. Angliss, B. A. Allen, P. L. Boveng, J. M. Breiwick, M. F.
 Cameron, P. J. Clapham, S. P. Dahle, M. E. Dahlheim, B. S. Fadely, M. C. Ferguson, L.
 W. Fritz, R. C. Hobbs, Y. V. Ivashchenko, A. S. Kennedy, J. M. London, S. A. Mizroch,
 R. R. Ream, E. L. Richmond, K. E. W. Shelden, R. G. Towell, P. R. Wade, J. M. Waite,
 and A. N. Zerbini. 2017. Alaska marine mammal stock assessments, 2016. U.S. Dep.
 Commer., NOAA Tech. Memo. NMFS-AFSC-355.
- Nachtigall, P. E., J. L. Pawloski, and W. W. L. Au. 2003. Temporary threshold shifts and recovery following noise exposure in the Atlantic bottlenosed dolphin (*Tursiops truncatus*). Journal of the Acoustical Society of America **113**:3425-3429.
- Nachtigall, P. E., A. Y. Supin, J. L. Pawloski, and W. W. L. Au. 2004. Temporary threshold shifts after noise exposure in the bottlenose dolphin (*Tursiops truncatus*) measured using evoked auditory potentials. Marine Mammal Science **20**:672-687.
- NCEI. 2016. State of the climate: global analysis for annual 2015. National Centers for Environmental Information, National Oceanic and Atmospheric Administration, U.S. Department of Commerce, Published online at: http://www.ncdc.noaa.gov/sotc/global/201513.
- Nisbet, I. C. 2000. Disturbance, habituation, and management of waterbird colonies. Waterbirds:312-332.
- NMFS. 1995. Status review of the United States Steller sea lion (Eumetopias jubatus) population. NOAA, NMFS, AFSC, National Marine Mammal Laboratory, Seattle, Washington.
- NMFS. 2008a. Recovery Plan for the Steller Sea Lion (*Eumetopias jubatus*). Revision. National Marine Fisheries Servies, Silver Spring, MD.
- NMFS. 2008b. Recovery plan for the Steller sea lion (*Eumetopias jubatus*). Revision. Office of Protected Resources, National Marine Fisheries Service, National Oceanic and Atmospheric Administration, U.S. Department of Commerce, Silver Spring, Maryland.
- NMFS. 2010. Final recovery plan for the fin whale (*Balaenoptera physalus*). Office of Protected Resources, National Marine Fisheries Service, National Oceanic and Atmospheric Administration, U.S. Department of Commerce, Silver Spring, Maryland.

- NMFS. 2013a. Effects of oil and gas activities in the Arctic Ocean: supplemental draft Environmental Impact Statement. Office of Protected Resources, National Marine Fisheries Service, National Oceanic and Atmospheric Administration, U.S. Department of Commerce, Silver Spring, Maryland.
- NMFS. 2013b. Supplemental Draft Environmental Impact Statement for the Effects of Oil and Gas Activities in the Arctic Ocean. USDOC, National Oceanic and Atmospheric Administration, National Marine Fisheries Service, Office of Protected Resources.
- NMFS. 2016a. Humpback whale (*Megaptera novaeangliae*). Office of Protected Resources, National Marine Fisheries Service, National Oceanic and Atmospheric Administration, U.S. Department of Commerce, http://www.nmfs.noaa.gov/pr/species/mammals/whales/humpback-whale.html.
- NMFS. 2016b. NOAA APPS: scientific research permits issued for bowhead, fin, and humpback whales in Alaska. National Marine Fisheries Service, National Oceanic and Atmospheric Administration, U.S. Department of Commerce.
- NMFS. 2016c. Occurrence of Distinct Population Segments (DPSs) of Humpback Whales off Alaska. National Marine Fisheries Service, Alaska Region. Revised December 12, 2016.
- NMFS. 2016d. Recovery Plan for the Cook Inlet Beluga Whale (*Delphinapterus leucas*).

 National Marine Fisheries Service, Alaska Region, Protected Resources Division, Juneau, AK.
- NMFS. 2016e. Technical Guidance for Assessing the Effects of Anthropogenic Sound on Marine Mammal Hearing: Underwater Acoustic Thresholds for Onset of Permanent and Temporary Threshold Shifts. U.S. Dept. of Commer., NOAA. NOAA Technical Memorandum NMFS-OPR-55, 178 p.
- Nowacek, D. P., L. H. Thorne, D. W. Johnston, and P. L. Tyack. 2007. Responses of cetaceans to anthropogenic noise. Mammal Review **37**:81-115.
- NRC (Nation Research Council). 2003. Ocean Noise and Marine Mammals. Ocean Study Board, National Academy Press, Washington, DC.
- Okey, T. A., H. M. Alidina, V. Lo, and S. Jessen. 2014. Effects of climate change on Canada's Pacific marine ecosystems: a summary of scientific knowledge. Review of Fish Biology and Fisheries **24**:519-559.
- Oleson, E. M., S. M. Wiggins, and J. A. Hildebrand. 2007. The impact of non-continuous recording on cetacean acoustic detection probability. Page 19 3rd International Workshop on the Detection and Classification of Marine Mammals Using Passive Acoustics, Boston, MA
- Owings, D. H., M. P. Rowe, and A. S. Rundus. 2002. The rattling sound of rattlesnakes (*Crotalus viridis*) as a communicative resource for ground squirrels (*Spermophilus beecheyi*) and burrowing owls (*Athene cunicularia*). Journal of Comparative Psychology **116**:197-205.
- Parker, G. A. 1974. Courtship Persistence and Female-Guarding as Male Time Investment Strategies. Behaviour **48**:157-184.
- Parks, S. E. 2003. Response of North Atlantic right whales (*Eubalaena glacialis*) to playback of calls recorded from surface active groups in both the North and South Atlantic. Marine Mammal Science **19**:563-580.
- Parks, S. E. 2009. Assessment of acoustic adaptations for noise compensation in marine mammals. Office of Naval Research.
- Patricelli, G. L., and J. L. Blickley. 2006. Avian communication in urban noise: Causes and

- consequences of vocal adjustment. Auk 123:639-649.
- Payne, K., and R. Payne. 1985. Large scale changes over 19 years in songs of humpback whales in Bermuda. Zeitschrift fur Tierpsychologie **68**:89-114.
- Payne, R. 1978. A note on harassment. Pages 89-90 *in* K. S. Norris and R. R. Reeves, editors. Report on a workshop on problems related to humpback whals (*Megaptera novaeangliae*) in Hawaii. Sea Life Inc., Makapuu Pt., HI.
- Payne, R. S. 1970. Songs of the humpback whale. Capitol Records, Hollywood, CA.
- Pirotta, E., K. L. Brookes, I. M. Graham, and P. M. Thompson. 2014. Variation in harbour porpoise activity in response to seismic survey noise. Biology Letters **10**.
- Pitcher, K. W., P. F. Olesiuk, R. F. Brown, M. S. Lowry, S. J. Jeffries, J. L. Sease, W. L. Perryman, C. E. Stinchcomb, and L. F. Lowry. 2007. Abundance and distribution of the eastern North Pacific Steller sea lion (*Eumetopias jubatus*) population. Fishery Bulletin **107**:102-115.
- Popov, V. V., V. O. Klishin, D. I. Nechaev, M. G. Pletenko, V. V. Rozhnov, A. Y. Supin, E. V. Sysueva, and M. B. Tarakanov. 2011a. Influence of acoustic noises on the white whale hearing thresholds. Doklady Biological Sciences **440**:332-334.
- Popov, V. V., A. Y. Supin, D. Wang, K. Wang, L. Dong, and S. Wang. 2011b. Noise-induced temporary threshold shift and recovery in Yangtze finless porpoises *Neophocaena phocaenoides asiaeorientalis*. Journal of the Acoustical Society of America **130**:574-584.
- Quijano, J., and M. Austin. 2017. Biorka Island Dock Replacement: Modeling Pile Installation Sound Footprints. Document 01309, Version 2.1. Technical report by JASCO Applied Sciences for R&M Consultants, Inc.
- Raum-Suryan, K. L., K. W. Pitcher, D. G. Calkins, J. L. Sease, and T. R. Loughlin. 2002. Dispersal, rookery fidelity, and metapopulation structure of Steller sea lions (Eumetopias jubatus) in an increasing and a decreasing population in Alaska. Marine Mammal Science **18**:746-764.
- Richardson, W. J. 1998. Marine mammal and acoustical monitoring of BP Exploration (Alaska)'s open-water seismic program in the Alaskan Beaufort Sea, 1997. LGL Rep. TA2150-3. Rep. from LGL Ltd. (King City, Ont.), Greeneridge Sciences Inc.
- Richardson, W. J. 1999. Marine mammal and acoustical monitoring of Western Geophysical's openwater seismic program in the Alaskan Beaufort Sea, 1998. . TA2230-3, Report from LGL Ltd., King City, Ontario, and Greeneridge Sciences Inc., Santa Barbara, CA, for western Geophysical, Houston, TX and National Marine fisheries Service, Anchorage, AK.
- Richardson, W. J., C. R. Greene, Jr., C. I. Malme, and D. H. Thomson. 1995. Marine mammals and noise. Academic Press, Inc., San Diego, CA.
- Richardson, W. J., and C. I. Malme. 1993. Man-made noise and behavioral responses. Pages 631-700 *in* J. J. Burns, J. J. Montague, and C. J. Cowles, editors. The bowhead whale. Society for Marine Mammology, .
- Richardson, W. J., B. Wursig, and C. R. Greene. 1986. Reactions of bowhead whales, *Balaena mysticetus*, to seismic exploration in the Canadian Beaufort Sea. Journal of the Acoustical Society of America **79**:1117-1128.
- Ridgway, S. H., D. A. Carder, R. R. Smith, T. Kamolnick, C. E. Schlunt, and W. R. Elsberry. 1997. Behavioural responses and temporary shift in masked hearing threshold of bottlenose dolphins, *Tursiops truncatus*, to 1-second tones of 141 to 201 dB re 1 μPa. Naval Command, Control and Surveillance Center, RDT&E Division, San Diego,

- California.
- Rolland, R. M., S. E. Parks, K. E. Hunt, M. Castellote, P. J. Corkeron, D. P. Nowacek, S. K. Wasser, and S. D. Kraus. 2012. Evidence that ship noise increases stress in right whales. Proceedings of the Royal Society of London Series B Biological Sciences **279**:2363-2368.
- Romano, T., M. Keogh, C. Kelly, P. Feng, L. Berk, C. Schlundt, D. Carder, and J. Finneran. 2004. Anthropogenic sound and marine mammal health: measures of the nervous and immune systems before and after intense sound exposure. Canadian Journal of Fisheries and Aquatic Sciences **61**:1124-1134.
- Romano, T. A., J. Olschowka, S. Felten, V. Quaranta, S. Ridgway, and D. Felten. 2002. Immune response, stress, and environment: Implications for cetaceans. Cell and Molecular Biology of Marine Mammals; CJ Pfeiffer, ed. Krieger Publishing Co., Inc.
- Ryan, M. J. 1985. The túngara frog: a study in sexual selection and communication. The University of Chicago Press, Chicago, IL.
- Salden, D. R. 1993. Effects of research boat approaches on humpback whale behavior off Maui, Hawaii, 1989-1993. Page 94 Tenth Biennial Conference on the Biology of Marine Mammals, Galveston, Texas.
- Schlundt, C. E., J. J. Finneran, D. A. Carder, and S. H. Ridgway. 2000. Temporary shift in masked hearing thresholds of bottlenose dolphins, *Tursiops truncatus*, and white whales, *Delphinapterus leucas*, after exposure to intense tones. Journal of the Acoustical Society of America **107**:3496-3508.
- SEASWAP. 2017. Sperm Whale Data Tracker. Accessed September 7, 2017 at: http://seaswap.info/whaletracker/.
- Secretariat of the Convention on Biological Diversity. 2014. An Updated Synthesis of the Impacts of Ocean Acidification on Marine Biodiversity. Montreal, Technical Series No. 75, 99 pp.
- Serreze, M. C., and R. G. Barry. 2011. Processes and impacts of Arctic amplification: a research synthesis. Global and Planetary Change **77**:85-96.
- Sharpe, F. A., and L. M. Dill. 1997. The behavior of Pacific herring schools in response to artificial humpback whale bubbles. Canadian Journal of Zoology-Revue Canadienne De Zoologie **75**:725-730.
- Sigler, M. F., J. N. Womble, and J. J. Vollenweider. 2004. Availability to Steller sea lions (Eumetopias jubatus) of a seasonal prey resource: a prespawning aggregation of eulachon (Thaleichthys pacificus). Canadian Journal of Fisheries and Aquatic Sciences **61**:1475-1484.
- Silber, G. K. 1986. The relationship of social vocalizations to surface behavior and aggression in the Hawaiian humpback whale (*Megaptera novaeangliae*). Canadian Journal of Zoology-Revue Canadienne De Zoologie **64**:2075-2080.
- Simmonds, M. P., and W. J. Eliott. 2009. Climate change and cetaceans: Concerns and recent developments. Journal of the Marine Biological Association of the United Kingdom **89**:203-210.
- Simmonds, M. P., and S. J. Isaac. 2007. The impacts of climate change on marine mammals: Early signs of significant problems. Oryx **41**:19-26.
- Sinclair, E. H., and T. K. Zeppelin. 2002. Seasonal and spatial differences in diet in the western stock of Steller sea lions (Eumetopias jubatus). Journal of Mammalogy **83**:973-990.
- Slabbekorn, H., and E. A. Ripmeester. 2008. Birdsong and anthropogenic noise: Implications

- and applications for conservation. Molecular Ecology Resources 17:72-83.
- Southall, B. L., A. E. Bowles, W. T. Ellison, J. J. Finneran, R. L. Gentry, C. R. Greene, Jr., D. Kastak, D. R. Ketten, J. H. Miller, P. E. Nachtigall, W. J. Richardson, J. A. Thomas, and P. L. Tyack. 2007. Marine mammal noise exposure criteria: initial scientific recommendations. Aquatic Mammals 33:411-521.
- Stafford, K. M., D. K. Mellinger, S. E. Moore, and C. G. Fox. 2007. Seasonal variability and detection range modeling of baleen whale calls in the Gulf of Alaska, 1999-2002. Journal of the Acoustical Society of America **122**:3378-3390.
- Straley, J., T. O'Connell, S. Mesnick, L. Behnken, and J. Liddle. 2005. Sperm whale and longline fisheries interactions in the Gulf of Alaska. North Pacific Research Board R0309 Final Report:15.
- Straley, J., G. Schorr, A. Thode, J. Calambokidis, C. Lunsford, E. Chenoweth, V. O. Connell, and R. Andrews. 2014. Depredating sperm whales in the Gulf of Alaska: local habitat use and long distance movements across putative population boundaries. Endangered Species Research 24:125-135.
- Straley, J. M. 1990. Fall and winter occurrence of humpback whales (Megaptera novaeangliae) in southeastern Alaska. Report of the International Whaling Commission Special:319-323
- Thompson, P. O., W. C. Cummings, and S. J. Ha. 1986. Sounds, source levels, and associated behavior of humpback whales, Southeast Alaska. Journal of the Acoustical Society of America **80**:735-740.
- Thompson, T. J., H. E. Winn, and P. J. Perkins. 1979. Mysticete sounds. Pages 403-431 *in* H. E. Winn and B. L. Olla, editors. Behavior of Marine Animals: Current Perspectives in Research Vol. 3: Cetaceans. Plenum Press, New York, NY.
- Todd, S., P. Stevick, J. Lien, F. Marques, and D. Ketten. 1996. Behavioural effects of exposure to underwater explosions in humpback whales (*Megaptera novaeangliae*). Canadian Journal of Zoology-Revue Canadienne De Zoologie **74**:1661-1672.
- Tyack, P., and H. Whitehead. 1983. Male competition in large groups of wintering humpback whales. Behaviour **83**:132-154.
- Tyack, P. L. 1981. Interactions between singing Hawaiian humpback whales and conspecifics nearby. Behavioral Ecology and Sociobiology **8**:105-116.
- Vu, E. T., D. Risch, C. W. Clark, S. Gaylord, L. T. Hatch, M. A. Thompson, D. N. Wiley, and S. M. Van Parijs. 2012. Humpback whale song occurs extensively on feeding grounds in the western North Atlantic Ocean. Aquatic Biology 14:175-183.
- Wade, P. R., T. J. Quinn II, J. Barlow, C. S. Baker, A. M. Burdin, J. Calambokidis, P. J. Clapham, E. Falcone, J. K. B. Ford, C. M. Gabriele, R. Leduc, D. K. Mattila, L. Rojas-Bracho, J. Straley, B. L. Taylor, J. Urbán R., D. Weller, B. H. Witteveen, and M. Yamaguchi. 2016. Estimates of abundance and migratory destination for North Pacific humpback whales in both summer feeding areas and winter mating and calving areas. Paper SC/66b/IA21 submitted to the Scientific Committee of the International Whaling Commission, June 2016, Bled, Slovenia.
- Ward, W. D. 1997. Effects of high-intensity sound. Pages 1497-1507 *in* M. J. Crocker, editor. Encyclopedia of Acoustics. Wiley, New York, New York.
- Warner, G., and M. Austin. 2016a. Alaska DOT Hydroacoustic Pile Driving Noise Study: Ketchikan Monitoring Results. JASCO Document 01167, Version 1.0. Technical report by JASCO Applied Sciences for Alaska Department of Transportation and Public

- Facilities.
- Warner, G., and M. Austin. 2016b. Alaska DOT Hydroacoustic Pile Driving Noise Study: Kodiak Monitoring Results. JASCO Document 01167, Version 2.0. Technical report by JASCO Applied Sciences for Alaska Department of Transportation and Public Facilities.
- Wartzok, D., W. A. Watkins, B. Wursig, and C. I. Malme. 1989. Movements and behaviors of bowhead whales in response to repeated exposures to noises associated with industrial activities in the Beaufort Sea. Report from Purdue Univ., Fort Wayne, IN, for Amoco Production Co., Anchorage, AK.
- Watkins, W. A. 1986. Whale Reactions to Human Activities in Cape-Cod Waters. Marine Mammal Science **2**:251-262.
- Weir, C. R. 2008. Overt responses of humpback whales (*Megaptera novaeangliae*), sperm whales (*Physeter macrocephalus*), and Atlantic spotted dolphins (*Stenella frontalis*) to seismic exploration off Angola. Aquatic Mammals **34**:71-83.
- Wiese, K. 1996. Sensory capacities of euphausiids in the context of schooling. Marine & Freshwater Behaviour & Phy **28**:183-194.
- Wieting, D. 2016. Interim Guidance on the Endangered Species Act Term "Harass". National Marine Fisheries Service, Office of Protected Resources. Silver Spring, MD. October 21, 2016.
- Winn, H. E., P. J. Perkins, and T. C. Poulter. 1970a. Sounds of the humpback whale. Pages 39-52 7th Annual Conference on Biological Sonar and Diving Mammals, Stanford Research Institute, Menlo Park.
- Winn, H. E., P. J. Perkins, and T. C. Poulter. 1970b. Sounds of the humpback whale. Pages 39-52 Seventh Annual Conference on Biological Sonar and Diving Mammals, Stanford Research Institute, Menlo Park, California.
- Winn, H. E., and N. E. Reichley. 1985. Humpback whale, *Megaptera novaeangliae* (Borowski, 1781). Pages 241-274 *in* S. H. Ridgway and S. R. Harrison, editors. Handbook of marine mammals. Academic Press, London, England.
- Winship, A. J., and A. W. Trites. 2003. Prey consumption of Steller sea lions (Eumetopias jubatus) off Alaska: how much prey do they require? Fishery Bulletin **101**:147-167.
- Womble, J. N., M. F. Sigler, and M. F. Willson. 2009. Linking seasonal distribution patterns with prey availability in a central-place forager, the Steller sea lion. Journal of Biogeography **36**:439-451.
- Womble, J. N., M. F. Willson, M. F. Sigler, B. P. Kelly, and G. R. VanBlaricom. 2005. Distribution of Steller sea lions Eumetopias jubatus in relation to spring-spawning fish in SE Alaska. Marine Ecology Progress Series **294**:271-282.
- Wright, A. J., N. A. Soto, A. Baldwin, M. Bateson, C. Beale, C. Clark, T. Deak, E. Edwards, A. Fernandez, A. Godinho, L. Hatch, A. Kakuschke, D. Lusseau, D. Martineau, L. Romero, L. Weilgart, B. Wintle, G. Notarbartolo Di Sciara, and V. Martin. 2007. Anthropogenic noise as a stressor in animals: A multidisciplinary perspective. International Journal of Comparative Psychology 201:250-273.
- Ydenberg, R. C., and L. M. Dills. 1986. The economics of fleeing from predators. Advances in the Study of Behavior **16**:229-249.
- Zuberbuhler, K., R. Noe, and R. M. Seyfarth. 1997. Diana monkey long-distance calls: Messages for conspecifics and predators. Animal Behaviour **53**:589-604.